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## Interactions between Titanium Dioxide Nanoparticles (NPs) and mucin

. Boimin  
*University of Massachusetts Amherst*

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**INTERACTIONS BETWEEN TITANIUM DIOXIDE NANOPARTICLES (NPS)  
AND MUCIN**

A Dissertation Presented

by

BOIMIN

Submitted to the Graduate School of the  
University of Massachusetts Amherst in partial fulfillment  
of the requirements for the degree of

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February 2021

Food Science





## INTERACTIONS BETWEEN TITANIUM DIOXIDE NANOPARTICLES (NPS) AND MUCIN

A Dissertation Presented

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## **DEDICATION**

*I dedicate this thesis to*

*My country, INDONESIA*

*My beloved families and friends*

*for their inspiration, unconditional love, and support*

*My respected professors*

*for their guidance.*

## ACKNOWLEDGEMENTS

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ABSTRACT  
**INTERACTION BETWEEN TITANIUM DIOXIDE NANOPARTICLES (NPS)  
AND MUCIN**

FEBRUARY 2021

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Titanium dioxide nanoparticles (TiO<sub>2</sub> NPs) are widely used in many food, consumer, and industrial products. However, little is known about the overall effects of TiO<sub>2</sub> NPs on the environment or human health. In order to elucidate the fate, transformation, transport, and toxicological impact of TiO<sub>2</sub> NPs, a better understanding is needed of how the physicochemical properties of TiO<sub>2</sub> NPs (e.g. size, charge, curvature, hydrophobicity, and surface functionality) interact with their microenvironments (e.g. pH, temperature, bile acids, microbiome, enzymes, surface-active components, and biopolymers).

Living organisms including humans have a natural mechanism to protect themselves from physical, biological, and chemical perils by generating mucin—the main gel-forming polymers of mucus and consists of core protein domains and densely O-

linked oligosaccharide chains. However, there is a very limited study even no study examining the interaction between TiO<sub>2</sub> NPs and mucin comprehensively.

This thesis was divided into three parts: first, a literature review focusing on the major routes of TiO<sub>2</sub> NPs entered the environment and human body, and the mechanistic interactions of biomolecules on the surface of TiO<sub>2</sub> NPs; the second part, the effect of TiO<sub>2</sub> NPs-mucin interaction on the alteration of physicochemical properties of TiO<sub>2</sub> NPs and mucin during the formation of biomolecular corona (BMC), aggregation and accumulation of TiO<sub>2</sub> NPs in water; the last part, the effect of phosphate-buffered saline (PBS) and pH to the change of surface charge, the formation of BMC, and hetero aggregation.

There were several major pathways of TiO<sub>2</sub> NPs entered the environment and the human system. TiO<sub>2</sub> NPs interacted with the human system via the respiratory system, skin, and gastrointestinal (GI) tract. Interaction between TiO<sub>2</sub> NPs and mucin might be induced by polyvalent binding as evidenced by three or more ligands (TiO<sub>2</sub> NPs) that were likely to interact with one molecule receptor (mucin). TiO<sub>2</sub> NPs-mucin interaction (at a mass ratio of TiO<sub>2</sub> NPs and mucin= 0.25) was likely to cause the massive hetero aggregation since evidenced by the increase of size, the hypochromic effect and redshift of UV-Vis spectra, and the appearance of spectral peaks of TiO<sub>2</sub> NPs and mucin by the surface-enhanced Raman Spectroscopy (SERS). The formation of unstable hetero aggregation in PBS required a higher concentration of TiO<sub>2</sub> NPs than in DIW. In deionized water (DIW), cationic BMC was potentially developed in acidic conditions; while in PBS there was, no cationic BMC formed. It might be caused by the effect of buffers and pH on physicochemical properties of TiO<sub>2</sub> NPs, particularly the surface

charge. Surface charges of  $\text{TiO}_2$  NPs in DIW were changed from positive to negative with the increase of pH, while surface charges of  $\text{TiO}_2$  NPs in PBS were all negative in various pH. Albeit both  $\text{TiO}_2$  NPs and mucin in PBS had a negative surface charge, SERS exhibited that  $\text{TiO}_2$  NPs-mucin interaction still occurred. In conclusion, the interaction between  $\text{TiO}_2$  NPs and mucin is potentially polyvalent binding and thereby induces irreversible hetero aggregation through various interactions. Further studies to know the types of interaction, reaction products, and possible consequences are required.

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## **CHAPTER I**

### **1. INTRODUCTION**

The estimated global production of TiO<sub>2</sub> NPs was approximately 2000 metric tons worth \$70 million in 2005 [1] and had increased to 5000 tons by 2010. This trend is expected to increase until 2025 continuously [2]. TiO<sub>2</sub> NPs are widely used in aquaculture, agriculture, and are listed as the top five nanoparticles (NPs) found in consumer products including food [3]. These facts have raised concerns related to ecological and human health [4-15]. But, determining the fate, transformation, transport, and toxicity of TiO<sub>2</sub> NPs remains challenging. It profoundly depends on the chemical reactivity and physicochemical property of TiO<sub>2</sub> NPs in a complex environment. Taking all these facts together, studies to understand the bioactivity of TiO<sub>2</sub> NPs through investigating the alteration of their chemical reactivity and physicochemical property are a necessity.

The reactivity and physicochemical property of NPs in the aquatic environment is strongly influenced by water properties (such as pH, water hardness, ionic composition, and temperature) [16-22] suspended particulate matter (SPM) in water [23-28] and the presence of living organisms [17, 29-32]. Water is also called the “universal solvent” [33] because of the capability of water to dissolve more substance than any other solvent, and thereby it is pivotal for living thing on earth including for human, particularly as a great solvent for nutrients in the gastrointestinal (GI) tract. Interactions of water with surrounding substances and smooth alterations in water chemistry may significantly modify the surface properties of nanoparticles (NPs). In summary, the chemical reactivity

and physicochemical properties of TiO<sub>2</sub> NPs in the aquatic environment and human GI tract may be influenced by water properties.

Natural organic matter (NOM) in water is expected to provide a profound impact on the surface properties of NPs through coating and surface charge alterations [25, 27, 34-35]. Mucin is well known as NOM that protects the host from biological, physical, and chemical hazards [36-39]. Mucin is the main gel-forming polymers of mucus and consists of core protein domains and densely O-linked oligosaccharide chains; hence, it confers negative charge to the mucin through carboxyl and sulfate groups [36].

Interactions between TiO<sub>2</sub> NPs and mucin and other NOM cause aggregations and accumulations of NPs in fish [40] and invertebrates [41]. Yin, C. et al. (2019), for instance, suggested that the accumulation of TiO<sub>2</sub> NPs may pass through aquatic organisms to the higher trophic levels (particularly to human) via seafood [42].

Subsequently, this food may be digested in the human gastrointestinal (GI) tract.

Therefore, TiO<sub>2</sub> NPs trapped in the food matrix are released and absorbed by the human gut [43]. In general, TiO<sub>2</sub> NPs-mucin interaction may lead to hetero aggregation and accumulation of TiO<sub>2</sub> NPs in aquatic organisms and humans.

Although the absorption of TiO<sub>2</sub> NPs in the human GI tract is considered at low levels [44], it may be toxic enough in cells. Bare TiO<sub>2</sub> NPs generated more free radicals (hydrogen peroxide (H<sub>2</sub>O<sub>2</sub>)) in phosphate-buffered saline (PBS) than in pure water [45]. As PBS is well known as a cell buffer, it indicates that bare TiO<sub>2</sub> NPs may be toxic in cells. However, TiO<sub>2</sub> NPs contained in digested food models (1.5 % w/w) did not significantly perturbed cellular proteome due to their interaction with food matrix [46] and mucin (mucus) [47]. Mucin excludes foreign or hazardous molecules from the host,

and permit the useful one such as nutrients [36, 48]. It can be noticed that buffer and mucin may play an important role in determining the toxicity of TiO<sub>2</sub> NPs in cells.

Disputes associated with the accumulation of TiO<sub>2</sub> NPs in tissue have been caused by the increase in organ levels was not always detected [44] due to very low absorptions and concentration-dependent. Although the absorption of TiO<sub>2</sub> NPs is low, they could be approved by visual detection in organs [49]. So, the detection of TiO<sub>2</sub> NPs accumulation may depend on the absorption and concentration of TiO<sub>2</sub> NPs.

The interaction between TiO<sub>2</sub> NPs and mucin occurs in the GI tract at various pH. TiO<sub>2</sub> NPs are poorly dissolved in the model gastric and intestinal environment [44]. TiO<sub>2</sub> NPs exposed with juices mimicking the gastric and intestinal compartment (pH= 2 & 7) tend to alter their size and surface charge due to agglomeration and protein adsorptions on their surface [50, 51]. However, cationic nanoparticles/biomolecules are reported to have the capability to cross the epithelial barrier [52]. *In vitro* studies on Caco-2 intestinal cells [53-56] and TR146 buccal cells [57] have shown the translocation of NPs through the epithelial barrier. Studies associated with the penetration of cell barrier by NPs have been conducted; but, the condition was not very representative of the real gut environment, particularly they did not consider the presence of mucin (Mercier-Bonin et al., 2018). It is interesting to be noted that pH may affect the formation of cationic TiO<sub>2</sub> NPs/BMC, the change of physicochemical properties, and the penetration of the epithelial barrier.

Taking all these together, the long-term goal of this study is to simplify the development of theory in dealing with the fate, transformation, transport, and toxicity of TiO<sub>2</sub> NPs by understanding the mechanistic interaction between TiO<sub>2</sub> NPs and a

complex environment. To achieve this goal, the overall objective of this study is to develop a basic theory about the mechanistic interaction between TiO<sub>2</sub> NPs and mucin by examining the effect of pH and buffer on the change of physicochemical properties of TiO<sub>2</sub> NPs/BMC. The central hypothesis of this study is that the interaction between TiO<sub>2</sub> NPs and mucin may alter the physicochemical properties of TiO<sub>2</sub> NPs (particularly surface charge and size). Additionally, pH and buffer may play a crucial role in terms of the physicochemical alteration of TiO<sub>2</sub> NPs/BMC properties, the formation of hetero aggregation, and cationic TiO<sub>2</sub> NPs/BMC.

In order to test the central hypothesis and achieve the objective, some specific aims are pursued:

1. ***Determination of the surface charge of TiO<sub>2</sub> NPs and mucin in deionized water (DIW).*** The surface charge of TiO<sub>2</sub> NPs and mucin will be determined by using dynamic light scattering (DLS) in various pH to evaluate pH values near the zero points of charge (pH<sub>pzc</sub>) and the formation of cationic TiO<sub>2</sub> NPs. It is expected to observe a homo aggregation of TiO<sub>2</sub> NPs at pH<sub>pzc</sub>, while dispersion at cationic TiO<sub>2</sub> NPs.
2. ***Determine the surface charge, size, and turbidity of BMC and characterize the TiO<sub>2</sub> NPs-mucin interaction in DIW.*** DLS will be utilized to determine the surface charge of BMC at a different pH, while the size will be measured by using Mastersizer. UV-Vis spectrophotometry will be performed to determine the turbidity (O.D. 600) and the maximum absorption of BMC at different concentrations of TiO<sub>2</sub> NPs. Surface-enhanced Raman spectroscopy (SERS) will be employed to characterize TiO<sub>2</sub> NPs-mucin interaction. It is an



expectation to observe the formation of a cationic BMC and a hetero aggregation at a certain mass ratio of TiO<sub>2</sub> NPs and mucin.

3. ***Determination of the type of binding and the change of energy during TiO<sub>2</sub> NPs-mucin interaction in DIW.*** Isothermal titration calorimetry (ITC) will be used to examine the change of energy during the formation of BMC and to determine the type of binding between TiO<sub>2</sub> NPs and mucin. This experiment anticipates observing exothermic energy and polyvalent binding.
4. ***Determination of the surface charge of TiO<sub>2</sub> NP, mucin, and BMC, and characterize the TiO<sub>2</sub> NPs- mucin interaction in PBS.*** The surface charge of TiO<sub>2</sub> NPs and BMC will be examined by using DLS. It is expected to observe an interaction between TiO<sub>2</sub> NPs and mucin, although the surface charge of both TiO<sub>2</sub> NPs and mucin is the same.

## CHAPTER II

### 2. LITERATURE REVIEW

#### Introduction

Nanotechnology has been widely applied in consumer products, environmental protection, agriculture, aquaculture, and medicine. A nanoparticle is described as a particle of matter that is between 1 and 100 nanometers (nm) in at least one dimension [58]. TiO<sub>2</sub> NPs, also known as nanocrystalline titanium dioxide, ultrafine titanium dioxide, or microcrystalline titanium dioxide, are one of the most widely used nanoparticles in the world [59]. Consequently, TiO<sub>2</sub> NPs are listed as the top five NPs found in consumer products [3].

TiO<sub>2</sub> NPs are commonly used as a colorant because of their brightness and opacifying strength. TiO<sub>2</sub> NPs consist of three different types of crystal polymorphs: anatase, brookite, and rutile. They are resistant to chemical disruptions and exhibit excellent thermal stability. When exposed to temperatures above 800 °C, brookite and anatase can transform into rutile, a more stable TiO<sub>2</sub> NP. Most importantly, TiO<sub>2</sub> NPs can absorb and scatter ultraviolet light due to their high refractive index. TiO<sub>2</sub> NPs have photoactive properties which are characterized by the different band gaps in their electron structure. Anatase is the most photoactive due to its highest bandgap among all crystal forms of TiO<sub>2</sub> NPs [60-62].

Due to widespread use, the potentially detrimental effects of TiO<sub>2</sub> NPs have raised environmental and health concerns, and the cause of those effects is mostly still debatable. In particular, the formation of biological corona (BMC), the adsorption of

biomolecules on the surface of TiO<sub>2</sub> NPs, has been a critical determinant of the bio-fate of the nanoparticles (NPs) [63]. Bare TiO<sub>2</sub> NPs tend to aggregate in aqueous media; consequently, these aggregates reduce the cell viability and cause expression of stress-related genes, for instance, those encoding interleukin-6 (IL-6) and heat shock protein 70B'[64]. Although TiO<sub>2</sub> NPs are likely to induce inflammatory and heat shock response, the conjugation of TiO<sub>2</sub> NPs with PEG, for instance, can eliminate aggregation due to steric hindrance and reduce their toxicity [64]. There is a growing interest in studying BMC due to its possible association with certain essential phenomena in biological systems [65]. BMC research stem faces specific challenges as the mechanism of TiO<sub>2</sub> NPs-biomolecules interactions is poorly understood [66]. As illustrated in the conceptual framework on potential TiO<sub>2</sub> NPs-biomolecule interactions, investigations on the formation of BMC in conditions relevant to environmental and/or physiological situations represent a research challenge and a gap in the literature (**Fig. 1**).

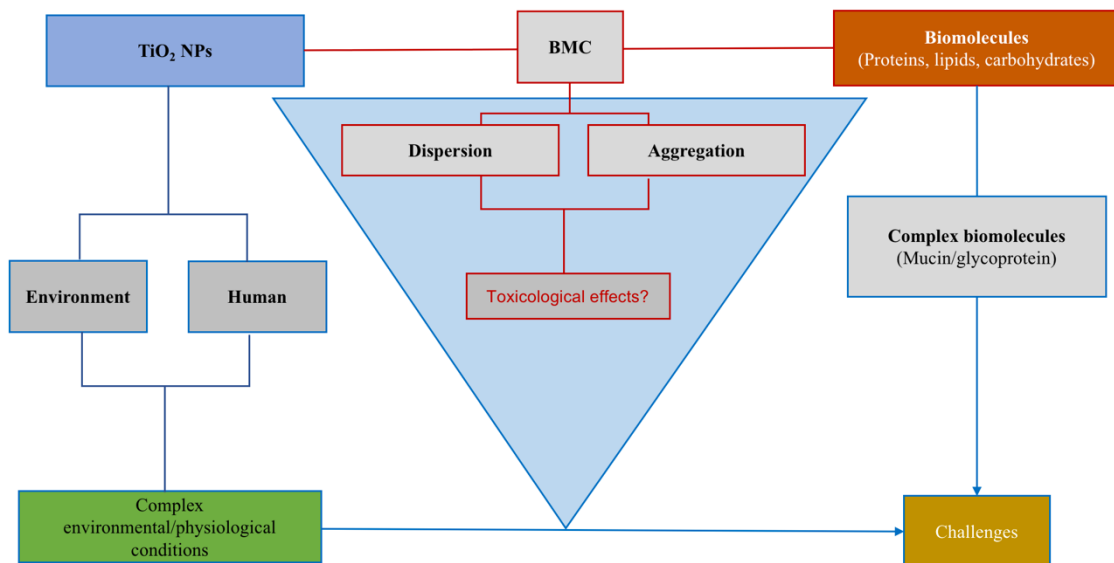


Figure 1: Conceptual framework of potential TiO<sub>2</sub> NPs-biomolecule interactions in environmental and human physiological conditions.

The application of TiO<sub>2</sub> NPs across multiple industries increases their release into the environment with emerging concerns regarding the adverse impacts on aquatic and terrestrial organisms [12]. TiO<sub>2</sub> NPs in water can interact with biomolecules or natural organic matter (NOM) such as humic acid or fulvic acid [67]. These organic substances are adsorbed by TiO<sub>2</sub> NPs, covering the surface of TiO<sub>2</sub> NPs, and forming biomolecular corona (BMC). The formation of BMC can induce either dispersion or aggregation which influence the cellular/tissue response of TiO<sub>2</sub> NPs.

Furthermore, TiO<sub>2</sub> NPs can enter terrestrial organisms via three primary routes: the respiratory system, gastrointestinal tract (GI), and skin. The GI tract route has gained the attention of researchers due to its significance to the immune system and digestive system. Studies report that when TiO<sub>2</sub> NPs interact with the outer cell membrane of organisms, they have various forms, such as micelles, droplets, vesicles and the pristine one. Once TiO<sub>2</sub> NPs enter the cell, they can interact with cytoplasmic molecules and organelle cells, and therefore generate radical oxygen species (ROS) and induce toxicity [11, 68].

The purpose of this systematic literature review is to provide a mechanistic understanding of the interaction between TiO<sub>2</sub> NPs and biomolecules (particularly mucin). This review examines the fate, conformation, transport, and toxicity of TiO<sub>2</sub> NPs and focuses on the change of physicochemical properties of TiO<sub>2</sub> NPs.

## **Materials and Methods**

The Preferred Reporting Items for Systematic Reviews and Meta-Analyses Protocols (PRISMA-P) guidelines were applied to this systematic review.

### **Eligibility Criteria**

Inclusion criteria were studies focusing on TiO<sub>2</sub> NPs pathways into the environment; ingestion and absorption pathways of TiO<sub>2</sub> NPs by humans; and interactions between TiO<sub>2</sub> NPs and various biomolecules. This review considered research articles, books and documents, clinical trials, meta-analyses, randomized controlled trials, review papers, and systematic reviews. Research papers not written in English were excluded from the study.

### **Search Strategy**

An extensive bibliographic search using PubMed (core collection), Web of Science, and Agricola databases was conducted to screen articles written in English and published with no limitations on the year of publication. Keywords were employed as search terms in all selected databases. Search terms included but were not limited to TiO<sub>2</sub> NPs and aquatic environment, TiO<sub>2</sub> NPs and gastrointestinal tract, interactions TiO<sub>2</sub> NPs and biomolecules, BMC formation, TiO<sub>2</sub> NPs and toxicity, TiO<sub>2</sub> NPs and dispersion, TiO<sub>2</sub> NPs and aggregation, TiO<sub>2</sub> NPs and mucin, etc. (Table 1).

### **Data Management, Screening, and Selection**

Titles and abstracts of articles were initially screened for further review. After duplicate and irrelevant articles were excluded, a careful review of full-length articles

was conducted. Articles were categorized under key search terms and systematically recorded in an excel spreadsheet (Office Excel software, Microsoft Corporation, One Microsoft Way, Redmond, WA 98052-6399, USA) (Table 1).

Table 1: Literature review based on PRISMA guidelines: Web of Science (highly cited), Agricola, and PubMed (including books, documents, clinical trials, meta-analyses, randomized controlled trials, reviews, systematic reviews)

No	Subject	WoS	Agricola	Pub Med	Articles with duplicates	Close duplicates	Articles found	Selected articles
1	TiO <sub>2</sub> NPs in environments	301	1,782	175	2,258	158	2,178	
2	TiO <sub>2</sub> NPs in human	22	93	136	251	12	245	311
3	Interactions TiO <sub>2</sub> NPs and biomolecules	35	467	96	598	22	587	
4	TiO <sub>2</sub> NPS (dispersion / aggregation)	107	887	83	1,077	60	1,047	
5	NPs and mucin	28	217	122	367	48	342	
6	TiO <sub>2</sub> NPs and mucin	2	3	1	6	2	5	777
	Total	495	3,449	613	4,557	302	4,404	1,088

## **Results and Discussions**

### **Search Results**

The initial search identified 4,557 articles comprising 603 titles from PubMed, 495 from Web of Science, and 3,449 from Agricola. After duplicates had been removed, 4,404 articles remained. The further selection was conducted and selected 1,088 articles, consisting of 311 articles related to the major route of TiO<sub>2</sub> NPs into environment and human, and 777 articles associated with the interaction between TiO<sub>2</sub> NPs and biomolecules. 301 articles had their titles and abstracts screened, and 120 were excluded

for not meeting inclusion criteria. The full-text screening did not exclude any studies, resulting in the inclusion of 181 articles. More specifically, this literature review included five (5) articles addressing the research on TiO<sub>2</sub> NPs and mucin:

1. Mercier-Bonin, M., Despax, B., Raynaud, P., Houdeau, E., & Thomas, M. (2018). Mucus and microbiota as emerging players in gut nanotoxicology: The example of dietary silver and titanium dioxide nanoparticles. *Critical reviews in food science and nutrition*, 58(6), 1023–1032.  
<https://doi.org/10.1080/10408398.2016.1243088> (PubMed & Agricola) [47].
2. Brun, E., Barreau, F., Veronesi, G., Fayard, B., Sorieul, S., Chaneac, C., . . . Carriere, M. (2014}). Titanium dioxide nanoparticle impact and translocation through ex vivo, in vivo and in vitro gut epithelia. *Particle and Fibre Toxicology*, 11} doi: {10.1186/1743-8977-11-13 (Web of Science (Highly cited)) [50].
3. Hajirezaee, S., Mohammadi, G., & Naserabad, S. S. (2020}). The protective effects of vitamin C on common carp (cyprinus carpio) exposed to titanium oxide nanoparticles (TiO<sub>2</sub>-NPs). *Aquaculture*, 518} doi: {10.1016/j.aquaculture.2019.734734 (Web of Science (Highly cited)) [40].
4. Bourgeault, A., Veronique Legros, Gonnet, F., Daniel, R., Aurelie Paquirissamy, Clemence Benatar, . . . Pin, S. (2017). Interaction of TiO<sub>2</sub> nanoparticles with proteins from aquatic organisms: The case of gill mucus from blue mussel. *Environmental Science and Poll (Agricola)* [69].

5. Tsai, S., Duran-Robles, E., Goshia, T., Mesina, M., Garcia, C., Young, J., . . . Chin, W. (2018). CeO<sub>2</sub> nanoparticles attenuate airway mucus secretion induced by TiO<sub>2</sub> nanoparticles. *Science of the Total Environment*, 631, 262-269. doi:10.1016/j.scitotenv.2018.03.001 (Agricola) [70].



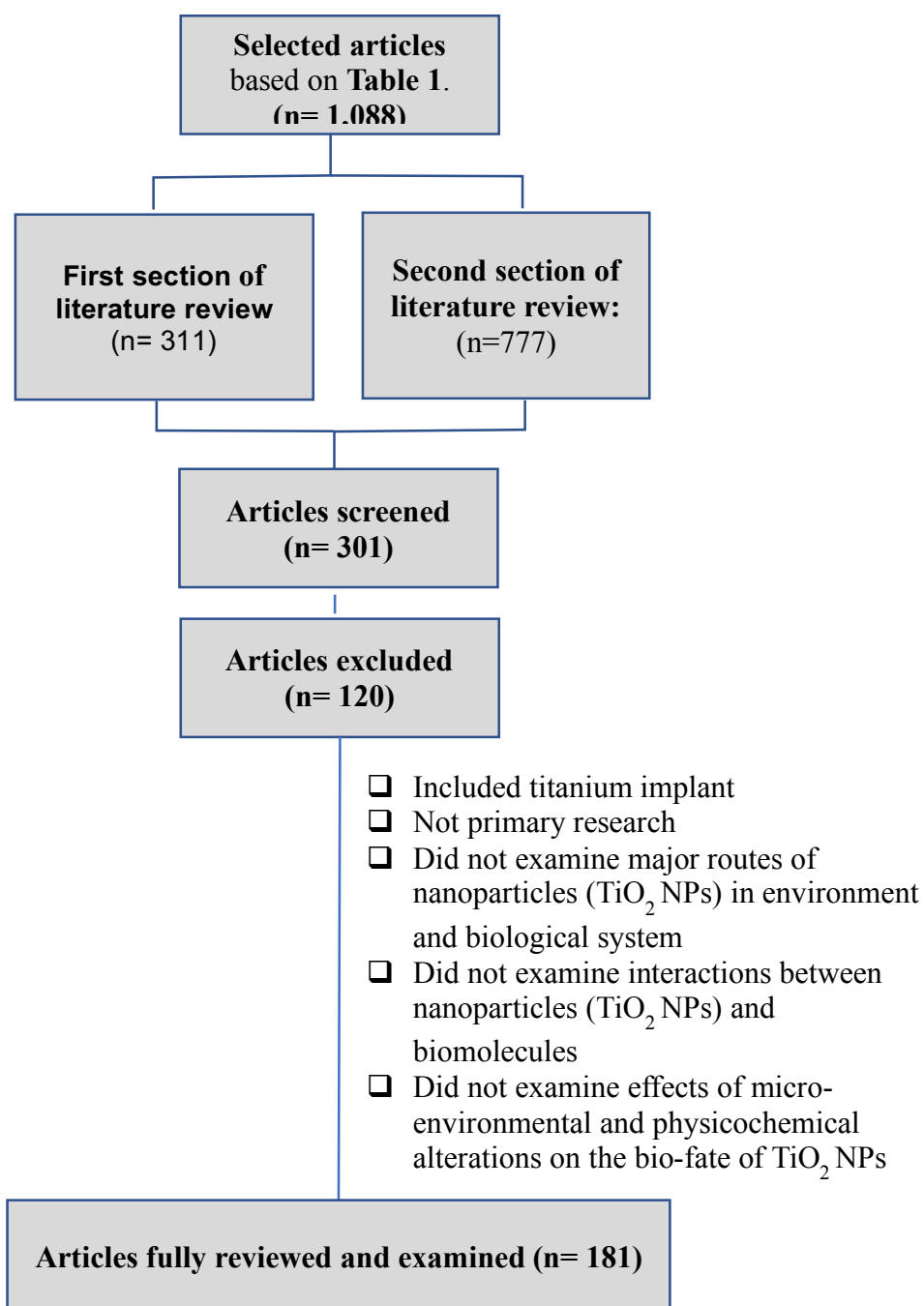


Figure 2: Flow chart for selection of articles adapted from PRISMA.

## Discussions

### **Impact of NP Properties on Biomolecular Corona (BMC) Formation**

There are two important terms in the formation of BMC, biological identity and synthetic identity. Adsorbed layers of biomolecule are assigned as the biological identity of NP, while physicochemical properties of NPs are defined as synthetic identity. The depiction of the biological and synthetic identity may reveal the existing gap in knowledge in the formation of BMC [71]. Biological identity is a derived property of an NP and depends on the myriad of physiological biomolecules. The first protein interacting with NPs has the highest abundance in the BMC due to the “memory” effect, which affects the subsequent protein-protein interaction during BMC formation [72]. Almost all the BMC research focuses on protein-NPs interaction. However, the biological identity of TiO<sub>2</sub> NPs which interact with lipids, carbohydrates, and complex molecules (such as glycoproteins) is mostly unknown.

The nano-bio interface study, initiated by Dawson and co-workers [63, 73-74] delivered the new field, which might be categorized as the safe version of NPs. Nevertheless, understanding the Nano-bio interface is very challenging due to its complexity. The Nano-bio interaction is highly associated with a dynamic physiological environment and presents difficulty in finding appropriate methods to analyze the interaction [66]. Here, we systematically present existing evidence in the literature and explain the current gap of knowledge on decoding the interaction between TiO<sub>2</sub> NPs and various types of biomolecules.

### **Size, Shape, Curvature, and Charge**

The surface properties of TiO<sub>2</sub> NPs are considered a fundamental determinant of BMC bioactivity. A slight change in the physicochemical properties of TiO<sub>2</sub> NPs may alter the biological effect of BMC, and thereby give rise to the unpredictability of the toxicity and bio-fate. Studies on BMC reveal the unpredictable behavior of TiO<sub>2</sub> NPs.

Studies report that the particle size is a determinant of NPs delivery and toxicity [75-76]. Particle size affects the qualitative and quantitative composition of BMC [77-78]. Additionally, current studies show that other surface properties of NPs, such as shape, curvature, hydrophobicity, and charge, also influence the binding of NPs with biomolecules. The higher curved surface and the smaller size of NPs results in less protein-protein interactions and oligomerizations of adsorbed proteins [79-80]. However, smaller particle size can lead to higher surface area and hydrophobicity, and thus faster BMC formation [81].

The formation of BMC may be driven by the size as well as other surface properties of TiO<sub>2</sub> NPs. The charge of both NPs and biomolecules play important role in the formation of BMC. The negative charge of NPs has an affinity towards protein with an isoelectric point (PI)>5.5, whereas the positive charge of NPs exhibits affinity towards protein with PI<5.5 [82]. However, the isoelectric point of TiO<sub>2</sub> NPs with protein and other biomolecules, particularly complex biomolecules such as glycoproteins, are not fully known.

More than 130 prior experiments showed that the different methods, proteins, and NPs systems, can cause a notable difference in the estimation of the dissociation

coefficient [83]. Studies have observed that positively charged NPs tend to have smaller dissociation coefficients than negatively charged NPs [83]. In addition, Calatayud et al. (2014) reported that the charge density of surface modified magnetic NPs in cell culture mediums exhibits a fivefold escalation in the particle size within a few minutes of incubation [84]. The final charge and size of BMC is frequently different from the initial charge and size of NPs. These findings suggest that the characterization of the final properties of TiO<sub>2</sub> NPs in environmentally and physiologically relevant bio-fluids is needed.

The shape of NPs matter [85]. Few studies observed the effect of shape on the formation of BMC [86]. NPs with more irregular shapes, for example, are reported to have less uptake in the human macrophage cell line [85]. Moreover, NPs with the same shape and mass concentration but different species can provide different toxicity [85]. For example, rod-shaped silver NPs are toxic, while rod-shaped NPs with the same concentration were found safe on the human lung epithelial cell [85]. More investigations are needed on the shape of TiO<sub>2</sub> NPs and the effect of shape on uptake and toxicity in cells.

### **Surface Chemistry**

Protein adsorption may be adjusted by adding different functional groups on the NP surface. The application of polyethylene glycol (PEG) to gain “stealth” behavior, for example, is the gold standard for controlling surface properties and maintaining corona-free conditions [66]. Some studies have reported the application of PEG on TiO<sub>2</sub> NPs to assess its potential to reduce the cytotoxicity of this NP [64]. Manipulation of surface

functional groups of NPs and their density exhibit a direct association with the model of protein adsorption [87. ]The adsorption of proteins on the surface of colloidal alumina NPs, for instance, has a linear association with the NPs functionality and the composition of amino acids on adsorbed proteins [88].

In addition, previous studies have reported that surface chemistry influences the surface charge of NPs and plays an important role in the composition and evolution of BMC [66]. The different composition of BMC can translate to different biological and physicochemical properties of NPs [66]. For example, the BMC composition of a native SiO<sub>2</sub> NPs is different from the amine (--NH<sub>2</sub>) and carboxyl (--COO(-)) modified surface of SiO<sub>2</sub> NPs owing to the different surface functionality of SiO<sub>2</sub> NPs [66]. Carboxyl-modified NPs show more stability (i.e. the size of aggregates and agglomeration rate) and less toxic effects on cells than native and amine-modified ones [89].

Another key regulator in the formation of BMC is exhibited by the arrangement of lipid functional groups in liposome-protein interactions [66]. Small alterations in lipid composition may substantially influence the formation of BMC [90], and thereby may contribute to changes in the biological impact of BMC. Studies suggest that each surface chemistry is specific and has its own potential bioactivity [66]. Limited data have reported the surface chemistry of TiO<sub>2</sub> NPs; therefore, further studies about the surface chemistry of TiO<sub>2</sub> NPs in association with the formation of BMC are required.

### **Hydrophobicity**

Modifying surface hydrophobicity may prevent the exposed surface of NPs from adsorbed protein; so, their biological properties are more likely to be governed by

physicochemical properties of NPs than adsorbed protein [66]. For example, zwitterion NPs may not form hard coronas at physiological serum concentrations [91] and the hydrophobic NPs adsorb albumin more readily than hydrophilic NPs [92]. Besides hydrophobicity, topography and surface curvature may affect the denaturation of adsorbed proteins; and the effect may be multiplied by the surface chemistry [93]

The smaller size of NPs may lead to more hydrophobic proteins being adsorbed [92]. Hydrophobicity tends to result in curvature increases and shields NPs from hydrophilic/aqueous environments. Hydrophobic interactions are like electrostatic interactions, playing an essential role in the qualitative and quantitative composition of BMC, and having wide toxicological effects [66]. Understanding the interplay of different physicochemical properties of TiO<sub>2</sub> NPs may provide a more comprehensive overview of BMC formation. Furthermore, exposing TiO<sub>2</sub> NPs with different surface chemistry to various biomolecules may provide a better understanding of the formation of BMC. However, studies to examine the effect of different physicochemical properties and surface chemistry of TiO<sub>2</sub> NPs on the formation of BMC are very limited, even unknown. Figure 6 illustrates findings from several studies on different types of NPs, including TiO<sub>2</sub> NPs. It is important to note that various NPs may lead to differential results).

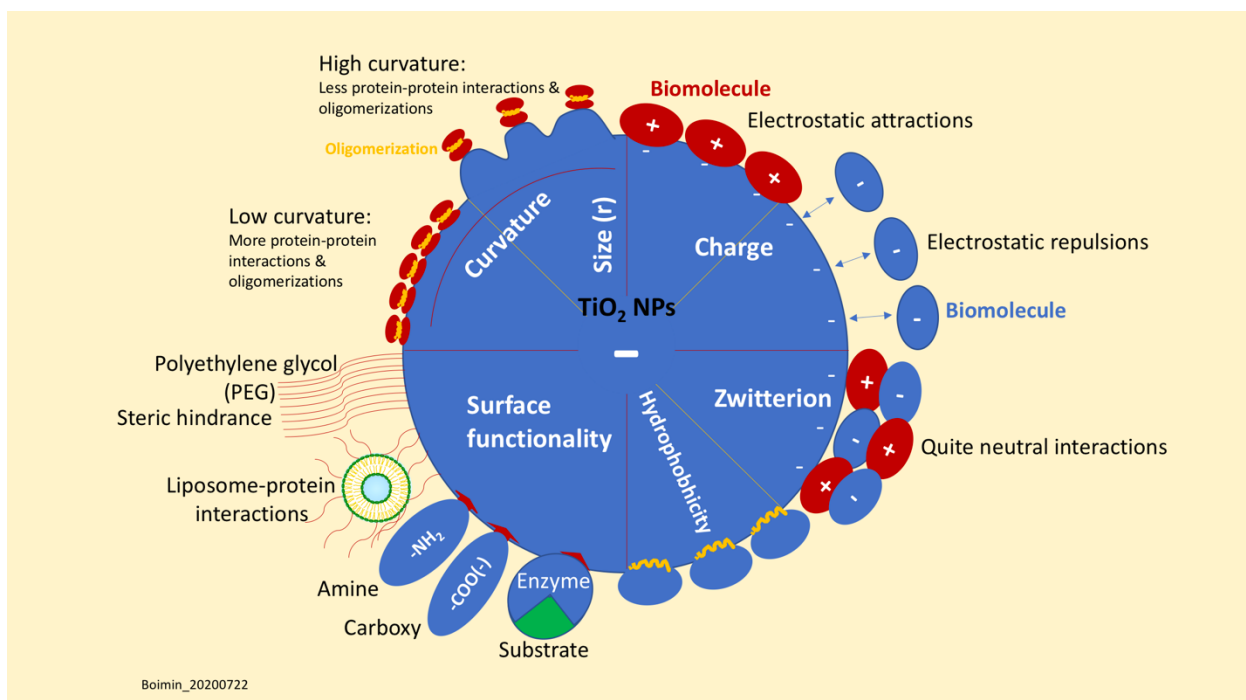


Figure 3: Illustrations based on the literatures [66; 75-92]: possible surface chemistry and/or physicochemical properties of NPs, including TiO<sub>2</sub> NPs. Different results may arise depending on the type and properties of NPs.

### **Interaction between TiO<sub>2</sub> NPs with Different Biomolecules and Potential Host Impact**

Once TiO<sub>2</sub> NPs enter the environment, they may interact with organic substances (biomolecules), form BMC, and transform their physicochemical properties and biological fate. The main components of biomolecules are proteins, lipids, and carbohydrates. However, some constituents, such as surfactants, colorants, minerals, preservatives, nutraceuticals, bases, acids, and buffers, are frequently included in the interactions between NPs and biomolecules [94-95], which cause complexities.

McClements, Xiao, and Demokritou (2017) argued that since TiO<sub>2</sub> NPs have different physicochemical properties compared to TiO<sub>2</sub> fine particles, bio-safety assessment for negative health and environmental outcomes is essential. Properties of

TiO<sub>2</sub> NPs may be altered by various components of industrial processing, the food matrix, or other elements, thus affecting dissolution, precipitation, adsorption, de-adsorption, release, entrapment, association (aggregation), and dissociation (dispersion) [43]. Zhang, D. et al. (2020) explained the main types of nanoparticle dispersions in liquids such as electrostatic repulsion, reversed micelle wrapping, and steric hindrance. The higher dispersion may indicate greater toxicity caused by TiO<sub>2</sub> NPs [10].

Klaine et al. (2008) described three possible models of aggregation between inorganic particles and biomolecules in liquids, based on Buffle et al. (1998) [96]. Sugars, amino acids, and other very fine biological substances may interact with inorganic particles and formed gels; while the larger biological substances such as polysaccharide fibrils may create small aggregation (stable suspension) and large aggregation (unstable suspension) [97]. **Fig. 4** illustrates the potential effects on TiO<sub>2</sub> NPs when they are dispersed into liquids containing various biomolecules [10, 43, 96].



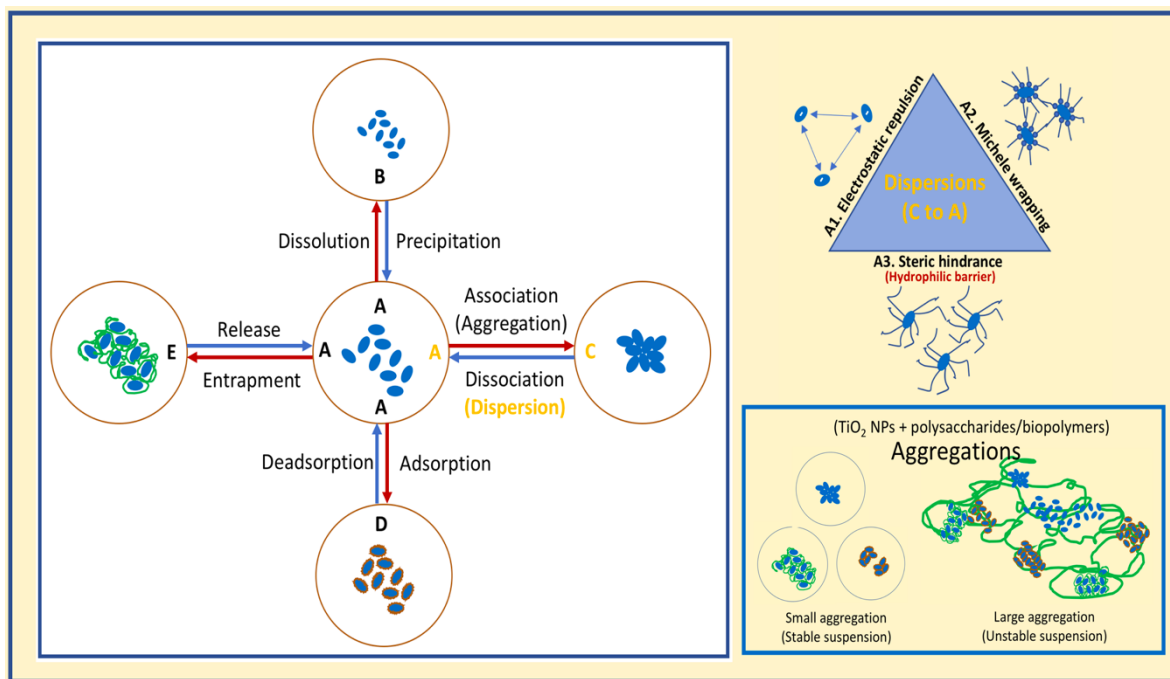


Figure 4: Potential effects of TiO<sub>2</sub> NPs when exposed to liquids which contain various biomolecules (this illustration is inspired and reflects findings by Buffle et al. (1998), Klaine et al. (2008), McClements, Xiao, & Demokritou (2017), and Zhang, D. et al. (2020).

The mechanisms by which exposure to TiO<sub>2</sub> NPs potentially leads to adverse environmental or human outcomes are not fully understood. The potential pathways of the interaction between TiO<sub>2</sub> NPs and cells are illustrated in **Fig. 5**. Initially, TiO<sub>2</sub> NPs - having some configurations such as pristine form, micelles, droplets, and vesicles - are dispersed in a liquid and adsorbed by the outer membrane of a cell before entering the cell [11]. Hypothetically, TiO<sub>2</sub> NPs interact with cytoplasmic molecules, organelle cells (particularly mitochondria), and generate radical oxygen species (ROS), which in turn likely induces toxicity [11].

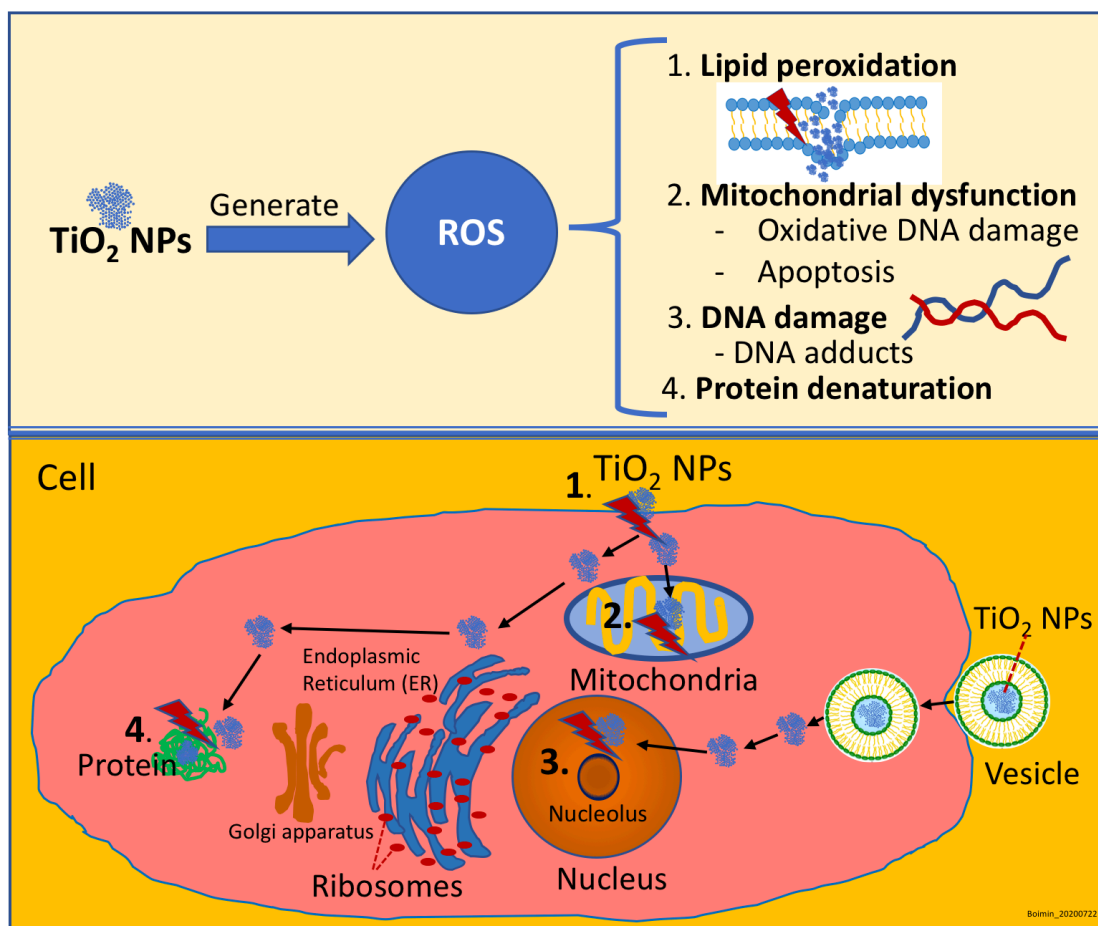


Figure 5: Possible effects of ROS generated by  $\text{TiO}_2$  NPs on cell.

### Interaction of $\text{TiO}_2$ NPs and Proteins

Interaction between protein and inorganic nanoparticles has been extensively researched [98]. McClements, Xiao and Demokritou (2017) elucidate inorganic nanomaterials within the food matrix and their implications for gastrointestinal fate [43]. They also report the main properties of various food proteins (Table 2).

Table 2: Summary of molecular characteristics of some common food-grade proteins [43].

Name	Main source	Molecular weight (kDa)	Main structural type	pI	~T <sub>m</sub> (°C)	Solubility
β-lactoglobulin	Milk	18.4	Globular	~5.0	~75	Water
Casein	Milk	19.0-25.2	Flexible	~4.6	~125-140	Water
Bovine serum albumin (BSA)	Milk/blood	66.5	Globular	~4.7	~180	Water
Lactoferrin	Milk	80	Globular	~8.0	~60-90	Water
Ovalbumin	Egg white	45	Globular	~4.6	~74; 82 <sup>s</sup>	Water
Lysozyme	Egg white	14.3	Globular	~11.0	~74	Water
Phosvitin	Egg yolk	36-40	Globular	~4.0	~80	Water
Gelatin	Animal collagen	Varies	Flexible	~8 <sup>A</sup>	~5 (fish)	Water
				~8 <sup>B</sup>	~40 (animal)	Water
Soy Glycinin	Soybean	360	Globular	~5.0	~67 <sup>7s</sup> ; 87 <sup>11s</sup>	Water
Zein	Corn	18-26	Globular	~6	~90	Organic solvent
Gliladin	Wheat	28-55	Globular	~6	~90	Organic solvent

Proteins can potentially have five different physicochemical mechanisms by which they can change the GI tract fate of nanoparticles [43]. Entrapment is where proteins may build a food matrix that traps NPs, and thus breaking the matrix is required prior to releasing NPs [99]. Surface active adsorption is when proteins can be adsorbed by NPs; hence, the interfacial characteristics of NPs are altered [100-101]. Proteins are good thickening and gelling agents and can modify the rheology and transport of NPs in the GI tract [102-103]. Proteins may lead to NPs aggregation due to their potential in bridging molecules and depleting flocculation [104]. And finally, proteins are effective buffering agents and in a sufficient amount, they can modulate the pH of GI tract fluids

[102].

Cao, X. and co-workers (2019) examined the interaction between TiO<sub>2</sub> NPs and casein (milk proteins), using a simulated GI tract model. Findings suggest that the aggregation takes place due to the bridging effect and reduction of electrostatic repulsion. This study reports that the gastric digestion of caseins which interact with TiO<sub>2</sub> NPs takes longer and their rate of digestion is lower than just caseins. Hence, the formation of BMC may inhibit pepsin to interact with peptide bonds [105].

Adsorption effects of TiO<sub>2</sub> NPs and temperature on the structure of plasmatic proteins such as bovine serum albumin (BSA) and fibrinogen (Fib) have been studied. The study found that the structure of adsorbed BSA on the surface of TiO<sub>2</sub> NPs differs from the solution one; while adsorbed Fib is no different from the Fib solution. Increasing the temperature causes the structure of adsorbed BSA does not change, but BSA solution is changed or denatured; meanwhile, both adsorbed Fib and Fib solution are denatured [106].

Nikpasand et al. (2019) reported that TiO<sub>2</sub> NPs could potentially accelerate wound healing in the early stages of injury due to the fibroblast proliferation and the rearrangement of tissue and collagen fibers granulation [107]. Although the TiO<sub>2</sub> NPs-protein interactions have been well-studied, there is limited data on the potential risks of TiO<sub>2</sub> NPs to human health.

### **Interaction between TiO<sub>2</sub> NPs and Lipids**

Lipids, generally consist of either neutral lipids (e.g. triacylglycerols (TAGs), diacylglycerol (DAGs), monoacylglycerols (MAGs), terpenes and hydrocarbons) or polar

lipids (e.g. free fatty acids (FFA), surfactants, and phospholipids) [108]. TAGs, DAGs, and phospholipids usually own a hydrolyzable ester bond which can induce the release of free FFAs and MAGs. These simple lipids can be both integrated into mixed micelles and adsorbed by epithelium cells [109].

TiO<sub>2</sub> NPs may damage lipid membranes via lipid peroxidation and lipid removal. NPs interact with lipid membranes in various structures, compositions, and dimensions, with each structure having its own potential bioactivity [110]. However, lipid peroxidation can be minimized through adsorption or layering the surface of TiO<sub>2</sub> NPs with serum protein [111]. TiO<sub>2</sub> NPs are scavengers of Ca<sup>2+</sup> ions, which may weaken the interaction between the membrane and support membrane; therefore, the addition of TiO<sub>2</sub> NPs can potentially remove the lipid membrane [112]. The removal of the lipid membrane can be prevented or minimized by neutralizing the membrane charge, particularly in the hydrophilic area. Previous research found that bilayers of membrane containing more zwitterion (relatively neutral regions) are more resistant to TiO<sub>2</sub> NPs perturbation than bilayers of membrane populated by anionic lipids [113]. These interesting findings raise questions about the overall impact of TiO<sub>2</sub> NPs on lipid membranes co-exposed to other chemicals contained in various skincare products.

### **Interaction of TiO<sub>2</sub> NPs and Carbohydrates**

Carbohydrates are typically categorized into two main groups, digestible and indigestible polysaccharides such as starch, cellulose pectin, alginate, carrageenan, and xanthan [110]. While some starches are easily hydrolyzed by amylases in the mouth and small intestine and transform into oligosaccharides and glucose, dietary fibers,

carbohydrate polymers with >10 monomeric units, are not hydrolyzed by digestive enzymes [114-115].

Carbohydrates have physicochemical mechanisms that may change the GI fate of TiO<sub>2</sub> NPs [110]. First, TiO<sub>2</sub> NPs are trapped into a solid matrix, a necessary step to release the nanoparticles [4, 116]. Second, some types of polysaccharides are good thickening and gelling agents that can change the rheological properties of the GI fluids and thereby alter the transport of TiO<sub>2</sub> NPs in the GI tract [117-119]. Third, some carbohydrates have electrostatic and hydrophobic interactions with TiO<sub>2</sub> NPs due to their nonpolar properties and charge, and thus lead to the adsorption [120-121]. Fourth, polysaccharides have mechanisms, such as bridging processes and depleted flocculation, that can promote aggregation. Starches are easily hydrolyzed in the mouth, small intestine, and stomach [122], and hence the interaction between starches and TiO<sub>2</sub> NPs in the GI tract is insignificant. Given that dietary fibers are persistently intact in the upper GI tract [123], they may interact and provide a significant impact on the physicochemical properties of TiO<sub>2</sub> NPs.

Although dietary fibers cannot be digested in the upper GI tract, the human gut microbiome can facilitate fermentation and produce metabolites like short-chain fatty acids. These in turn can impact metabolic regulation or become a substrate for microbial uptake [124]. Some studies have reported a closed relationship between dietary fibers and the microbiome, but no studies have documented the interactions between dietary fibers and the microbiome in the presence of TiO<sub>2</sub> NPs (Mercier-Bonin et al., 2018).

Regarding other forms of carbohydrates, Chen, Z., et al. (2015) observed the interactions between TiO<sub>2</sub> NPs and glucose. In this study, the toxicological effect of

varying doses of orally administered TiO<sub>2</sub> NPs and glucose to young animal models was monitored. Findings suggested that oral exposure to both TiO<sub>2</sub> NPs and TiO<sub>2</sub> NPs plus glucose caused liver, kidney, and heart injuries, as well as altered the number of white and red blood cells in a time, dose, and gender-dependent manner [125]. However, this study was not able to delineate the causal effects of TiO<sub>2</sub> NPs or glucose on observed injuries.

Another study by Chen and co-workers (2015) investigated the effect of TiO<sub>2</sub> NPs, glucose, and TiO<sub>2</sub> NPs plus glucose in animal models. Results indicated that the toxic effect of high-dose glucose was greater than both TiO<sub>2</sub> NPs and the combination of TiO<sub>2</sub> NPs with glucose. They suggested that it might be more essential to control the uptake of sugar than TiO<sub>2</sub> NPs [125]. Based on these findings, TiO<sub>2</sub> NPs seemed less harmful in animal models than glucose, but the unstable characteristics of TiO<sub>2</sub> NPs should be carefully considered prior to clinical trials.

### **Implications: Environmental and Human Pathways and Potential Risks**

#### **TiO<sub>2</sub> NPs & Their Major Routes into Environment**

Disposal and domestic wastewater, containing NPs, are inevitably released into the aquatic environment [1, 126-129] posing concerns about their environmental behavior and associated ecological risks. Previous studies have reported that NPs cause significant toxicity by inhibiting protein expression [130] and generating oxidative stress [131]. NPs exposed to water will generally either aggregate or disperse, resulting in the expression of different physicochemical properties and varying levels of toxicity. Dispersed NPs are likely to exhibit higher toxicity than the aggregated ones [10]; however, the aggregation

of NPs causes their accumulation in aquatic organisms [41]. Accumulation of TiO<sub>2</sub> NPs to higher trophic levels, particularly to humans due to the consumption of fish and seafood [42], is another concern. The toxicity of TiO<sub>2</sub> NPs in aquatic organisms is debated by scholars fueled by limited data on interactions between TiO<sub>2</sub> NPs and biomolecules in water.

Aquatic organisms can be susceptible to the toxic effects of TiO<sub>2</sub> NPs. Articles have described the environmental transformation of TiO<sub>2</sub> NPs in aquatic system and their accumulation in fish, bivalve mollusk, algae, some surimi products, as well as their possible toxic effects [9, 40, 42, 69, 132-133]. Shi, W., et al., (2019) observed the effect of ocean acidification (OA) on the bioaccumulation of TiO<sub>2</sub> NPs in three different bivalves and investigated the realistic amount of TiO<sub>2</sub> NPs that impacted the bivalves' health. This study reported that the accumulation of TiO<sub>2</sub> NPs in bivalves alleviates significantly due to OA, and thereby may increase the health risk of seafood consumers [133].

Studies showed that the oxidative stress and toxicity of fish and bivalve mollusks are altered due to TiO<sub>2</sub> NPs [11, 69]. Bivalve mollusks and TiO<sub>2</sub> NPs which are co-exposed to different organic and inorganic particles exhibit various effects including (1) enhanced accumulation and toxicity (TCDD, Cu, Pb, PCPs, Phenanthrene (Phe), As, Cd, Cu, Pb) [134-140], (2) decreased accumulation and increased toxicity (benzo(a)pyrene) [141], (3) decreased accumulation and toxicity (Cd<sup>2+</sup>) [142-143], (4) no change in accumulation and toxicity (Cd<sup>2+</sup>, TCDD, PBDEs) [140, 144-146].

Evaluation of TiO<sub>2</sub> NPs concentration in white-colored seafood and surimi-based food products found that TiO<sub>2</sub> NPs concentrations are relatively high [42]. These kinds of



food are considered an important route of TiO<sub>2</sub> NPs uptake, particularly among younger generations, aged 20-20 years, of the human population [42]. Lee and An (2013) investigated the eco-toxicity effects of TiO<sub>2</sub> NPs on the green algae (*Pseudokirchneriella subcapitata*) under the irradiation of visible UVA and UVB light. Their results showed that the increase of TiO<sub>2</sub> NPs concentration inhibits alga growth due to the destabilization of cell membranes and inhibition of alga growth was not alleviated by UV pre-irradiation conditions [132]. Indeed, the aquatic environment is complex due to the diversity of organisms, pH variations in fresh and sea water, and potential interactions with biomolecules in the aquatic environment. Further research is needed to understand the interactions of TiO<sub>2</sub> NPs and potential toxicity in aquatic environments.

Soil and terrestrial plants are also affected by the widespread distribution of TiO<sub>2</sub> NPs in the environment [147]. Tan, W. et al. (2018) reviewed studies on the interactions of TiO<sub>2</sub> NPs with soil and plants [8]. This review article described the change of physicochemical properties of TiO<sub>2</sub> NPs during the interactions. TiO<sub>2</sub> NPs are released into the environment via three primary routes: groundwater (18.5 %), soil (13.8%), and air (2.2%) [8]. TiO<sub>2</sub> NPs are then absorbed and translocated from the soil into leaves and fruits. As a result, plants such as cucumber fruits have increased phosphorus and potassium content [148]; barley and rice experience a decrease in kernels and grain yield [149]; an alteration of nutritional elements occurs in lettuce roots and shoots [150], tomato stems and leaves [151], and soybean roots [152]; and a change occurs in the antioxidant system of pinto bean plants sprayed with TiO<sub>2</sub> NPs at different stages of growth [153]. In a review article, Liu, Y., et al. (2019) explained that NPs enhance plant photosynthesis [154]. Many challenges and questions remain in association with TiO<sub>2</sub>

NPs-plants interactions, such as the specific effect of  $\text{TiO}_2$  NPs on plant proteins, the transgenerational effect of  $\text{TiO}_2$  NPs on plants, and the impact of other substances in soil on these interactions. **Fig. 6** illustrates the pathways of environmental exposure to  $\text{TiO}_2$  NPs as described by studies in this systematic review.

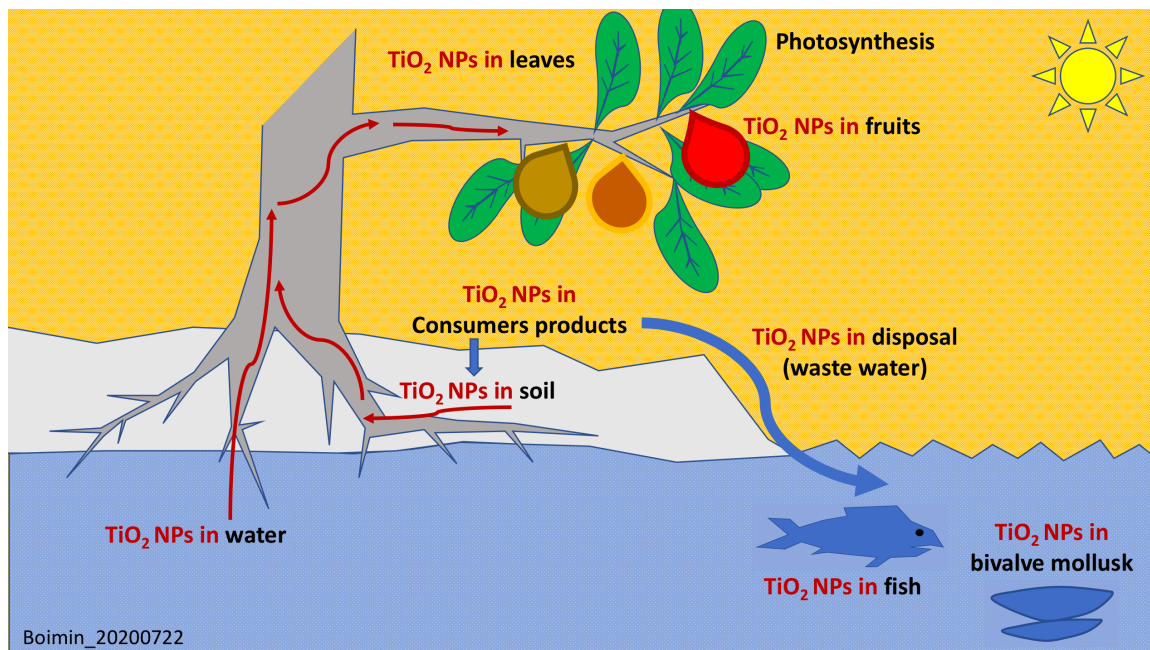


Figure 6: The major routes of  $\text{TiO}_2$  NPs into the environment.

### **$\text{TiO}_2$ NPs, Pathways into the Human Body, and Health Risks**

The International Agency for Research on Cancer has categorized  $\text{TiO}_2$  NPs as a possible carcinogen in humans [60] and the U.S. Institute of Occupational Safety and Health suggests  $\text{TiO}_2$  NPs exposure be limited to  $2.4 \text{ mg/m}^3$  [55].  $\text{TiO}_2$  NPs are widely used in toothpaste, sunscreen, cosmetics, paints, plastics, self-cleaning devices, pharmaceuticals, and food additives, as well as in industrial and medical applications, because of their strong photocatalytic activity [4, 156-158]. Investigations on the cytotoxicity and genotoxicity potential of  $\text{TiO}_2$  NPs have been conducted in both in vitro

and in vivo studies, however, toxicological data are disputable [159]. Scientific evidence suggests that TiO<sub>2</sub> NPs cause cytotoxic, genotoxic, and oxidative effects through oxidant generation, inflammation, and apoptosis [15, 157-158, 160-162]. Other studies have found that TiO<sub>2</sub> NPs induced low or no toxic effects [163-164].

There are three major pathways for human exposure to TiO<sub>2</sub> NPs – through the skin, respiratory system, and via ingestion to the gastrointestinal tract (Figure 4).

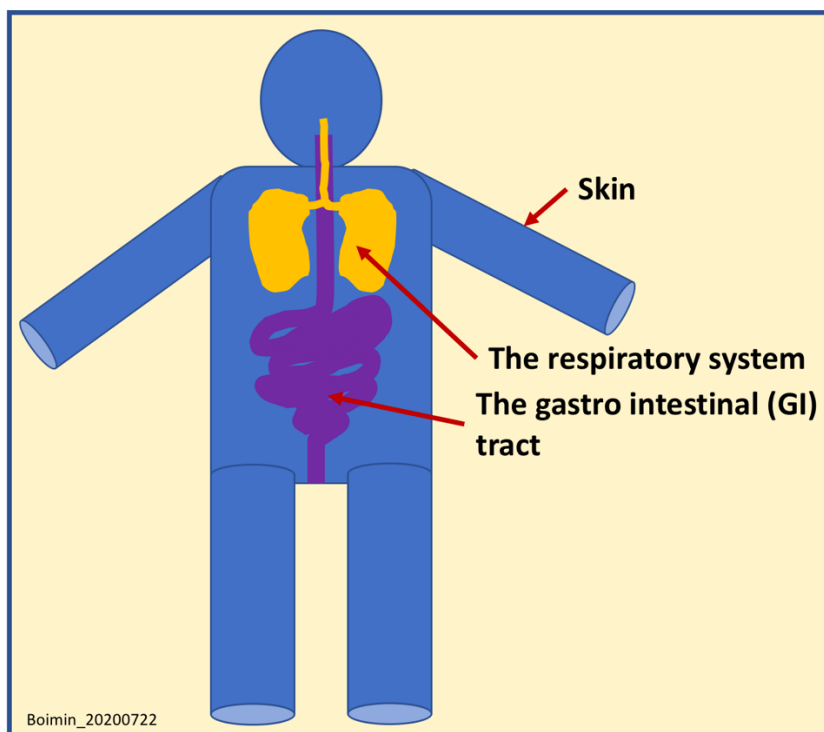


Figure 7: The major pathways of TiO<sub>2</sub> NPs into the human body.

## The Skin Pathway

Geppert et al. (2020) reviewed the interaction of  $\text{TiO}_2$  NPs with ingredients from modern lifestyle products and their effect on human skin cells [165]. They found that the outer layer of human skin was firm, inducing a limited penetration of inorganic  $\text{TiO}_2$  NPs [166]; however, the penetration of  $\text{TiO}_2$  NPs on the skin is possible [167-168]. Although most studies reported that no penetration of  $\text{TiO}_2$  NPs on the skin [169-170], Wu, J. et al. (2009) observed that  $\text{TiO}_2$  NPs penetrated the hairless skin of mice without causing breakage. Furthermore, higher doses given to the mice resulted in a higher number of  $\text{TiO}_2$  NPs found in organs such as the heart, brain, spleen, and liver. Liver exposure to  $\text{TiO}_2$  NPs in mice showed the alteration of malondialdehyde and superoxide dismutase levels [171]. A recent systematic review article reported that using a 3D skin model has exhibited to be promising to evaluate the toxicity of  $\text{TiO}_2$  NPs [172].

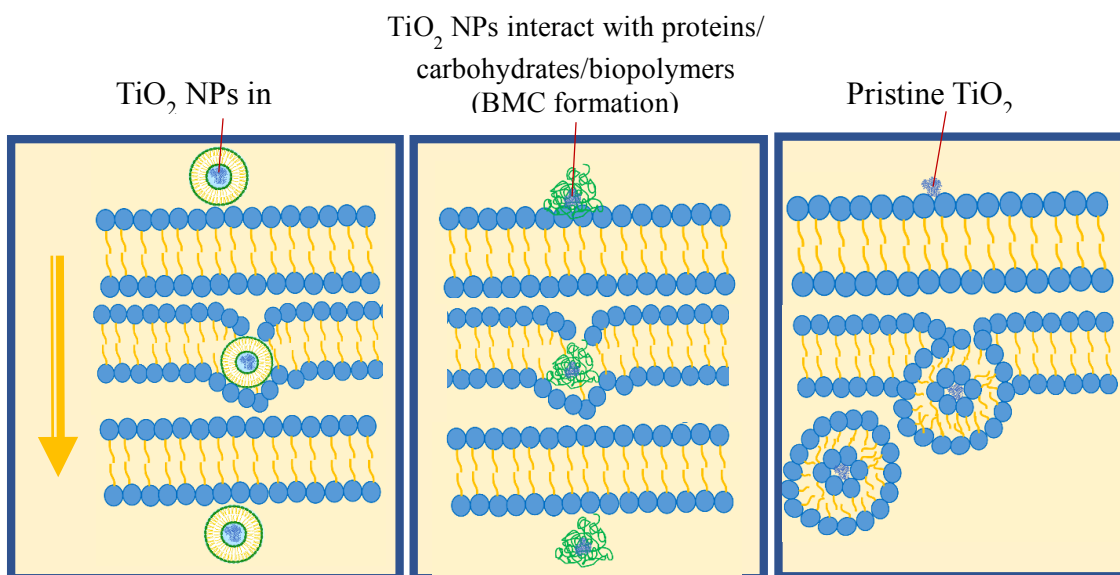


Figure 8: Interactions of  $\text{TiO}_2$  NPs (in various forms) with bilayers membrane.

Evidence indicates that dermal exposure to TiO<sub>2</sub> NPs does not pose a significant risk to human health, however, there is a considerable number of studies reporting a toxic effect of TiO<sub>2</sub> NPs in animal models [15]. Effects of TiO<sub>2</sub> NPs on the skin, which are co-exposed with an external stressor such as UV light and some chemicals, have also been studied [61-173]. Photo-toxicity to human skin keratinocytes with UV irradiation of TiO<sub>2</sub> NPs with four different sizes (<25 nm, 31 nm, <100 nm, and 325 nm) and two different crystal forms (i.e. rutile and anatase) was assessed [159-162]. Results suggested that all types of TiO<sub>2</sub> NPs generate ROS, and the smaller size TiO<sub>2</sub> NPs produce higher levels of photo-toxicity than larger size TiO<sub>2</sub> NPs [159-162].

Pelclova et al. (2019) reported that TiO<sub>2</sub> NPs are present in samples of human plasma and urine after 6-48 hours of applying sunscreen, indicating penetration of NPs to the protective layer of human skin cells [174]. Wright et al. (2017) proposed that all sizes of TiO<sub>2</sub> NPs tend to have a dose-dependent intensification of caspase 8 and 9 activity, superoxide production, and apoptosis on human keratinocyte cells (HaCaT) [175]. Crosera et al. (2015) corroborated the study findings presented by Wright et al. (2017) by demonstrating that TiO<sub>2</sub> NPs cause cytotoxic effects on HaCaT cells (EC<sub>50</sub> 10<sup>-4</sup>-10<sup>-5</sup> mol/L) [169]. In addition, human dermal fibroblast exposure to TiO<sub>2</sub> NPs results in a decrease in the cell area, mobility, proliferation, and a contractive ability of collagen [176]. Given these findings, the mechanistic interactions between TiO<sub>2</sub> NPs and skin cells, particularly the lipid membrane, needs to be investigated further.

## **The Respiratory System Pathway**

The major route by which ambient particles enter the human body is the respiratory tract, specifically through the nasal epithelium, the trachea-bronchiolar region, and the alveolar interstitium, i.e. respiratory bronchiole and pulmonary alveoli [177]. These particles can be retained in the lung, be cleared through the airways, and translocated and distributed within the body. Understanding the deposition of inhaled nanoparticles in the respiratory tract is crucial to evaluating air contamination, drug delivery, and health risks [178].

Since the first observation of particle deposition in the respiratory tract in 1881[179], there have been extensive investigations on particles entering the human body. Early studies on the deposition of aerosol particles in the respiratory system, involving in allergens, occupational and atmosphere dust, clinical aerosol, cigarette smoke, radioactive particles, and consumer aerosol products have been conducted [180]. The particle deposition in the respiratory tract depended on the properties of specific particles, including size, charge, density, as well as the breathing pattern of an individual [181].

It appears that smaller particles tend to be stored in the more distal region of human respiratory tract [181]. Particles with sizes  $>10\ \mu\text{m}$ ,  $>2.5\ \mu\text{m}$ ,  $<2.5\ \mu\text{m}$ , and  $<0.1\ \mu\text{m}$  tend to rest in the nose and upper airway (by filtration), the tracheobronchial region (by impaction and deposition), the pulmonary region (by sedimentation and diffusion), and the alveolar region (by penetration of alveolar epithelium to the bloodstream) [181-184].

Scientists have indicated that airborne NPs are deposited in three regions of the human respiratory tract. Airborne NPs with a size of 1 nm, for instance, can be found in the nasopharyngeal (90 %), tracheobronchial (10 %), and (3) alveolar regions of the body [185]. These nanoparticles can be found in the same regions in line with proportion. For example, 5 nm airborne NPs have a 30 %, 30 %, 30% distribution, and 20 nm particles have a 15%, 15%, 50% distribution to the nasopharyngeal, tracheobronchial, and alveolar regions, respectively. In addition, small nanoparticles exhibit penetration to cells (transcytosis) and can circulate in the circulatory and lymphatic systems [186].

Besides particle size, individual breathing patterns may affect particle deposition; and several studies reported that the deposition of particles in nasal breathing was higher than in oral breathing for both hydrophobic and hygroscopic aerosols [187]. For ethical reasons, there are currently no human studies on the deposition of TiO<sub>2</sub> NPs in the human respiratory tract due to their high toxic possibility.

Current meta-analysis studies in correlation with NPs and the respiratory system have been reported, such as transcriptional profiling to identify physicochemical properties of NPs that are determinants of the in vivo pulmonary response [188], occupational TiO<sub>2</sub> exposure, and lung cancer mortality [189], and transcriptomic responses to identify pulmonary disease outcomes for engineered NPs [190].

### **Gastrointestinal (GI) Tract Pathway**

NPs enter the human body and gastrointestinal (GI) tract through the ingestion of foods, water, drugs, and via topical application of some cosmetic products [185]. Payne et al. (1960) and Sanders & Ashworth (1961) presented key studies on the uptake and

absorption of particles in the GI tract. Payne et al. (1960) reported uptake of small resin by the alimentary canal of the calf [191]. Sanders & Ashworth (1961) observed the intestinal absorption of particulates and hepatocellular uptakes of PS latex particles [192]. These findings provide the fundamental knowledge base for researchers worldwide on uptake and absorption of particles in the GI tract [193-198].

Travel of NPs through the environment of the human GI tract is complex [199]. NPs firstly are exposed to digestive elements in the mouth (neutral pH, mucin, amylase, and electrolytes), then pass through to the esophagus and into the stomach where they are exposed to a pH of 2-3, lipase, pepsin, and electrolytes. NPs then move through the small intestine with exposure to a pH of 5-7, mucin, bile salts, phospholipids, pancreatic lipase, proteases, amylase, and electrolytes. Finally, NPs reach the colon with exposure to a pH of 6-7, gut microbiota, and undigested food substances [110]. Through the digestion process, NPs trapped within the food can be released into GI tract fluids due to the disruption of the food matrix [109, 200-202]. Moreover, the region where NPs are released in the GI tract depends on the structure and composition of the food [110].

McClements and Xiao (2017) explained that the riches of fluids, which have certain properties in the GI tract, may alter NPs properties such as 1) pH and ionic strength: these parameters could specify the surface potential and electrostatic interactions that affect the aggregation state and interactions of NPs with other components; and 2) surface-active components contained in gastrointestinal fluid: these components, which consist of surfactants, proteins, bile salts, phospholipids, and free fatty acids (FFAs), can be adsorbed by NPs and change their biological fate. The interfacial properties of inorganic NPs, for instance, is altered when they enter the GI



tract [110]. As a consequence, the cellular and tissue response to the NPs is also changed [108, 203]. Furthermore, their biological fate can be modified as well [204-207]. Notwithstanding pH and ionic strength effect to the alteration of the physicochemical property of TiO<sub>2</sub> NPs, the TiO<sub>2</sub> NPs-food component interactions and associated biological fate in the human GI tract are poorly understood.

Enzyme activity and enzymatic digestion can alter NPs properties [110]. NPs carrying carbohydrates, protein, or lipids, for example, can be digested by amylases, proteases, and lipases. As a result, the properties of NPs exhibited in certain regions of the GI tract may be different from the initial properties of NPs [110]. Biopolymers adsorbed on the surface of NPs can change the NPs interfacial properties. Adsorbing polymers may promote bridging flocculation while non-absorbing biopolymers may oppose flocculation [110].

The mucus layer in the surface of the GI tract may provide new insights into the inhibition of entry of NPs beyond the GI tract to other organs. The human GI tract can be analogized as a series of tubes and chambers with specific surface morphologies. The surface of the GI tract consists of villi and microvilli and a thin layer of mucus that serves as a gut barrier between NPs and epithelial cells [208]. The gut microbiome is primarily in the large intestine. The microbiome can ferment foods and generate metabolites that may change the NPs properties; reciprocally, ingested NPs may alter the properties of the gut microbiome [209]. Studies report that many types of NPs can disturb the balance of gut microbiome (dysbiosis) and cause adverse effects [209-210].

There are some subsequent mechanical forces that occur during NPs digestion in the GI tract, such as mastication in the mouth, peristaltic movements in the esophagus

and small intestine, and churning in the stomach. These forces may alter NPs properties, particularly their aggregation state [110]. Studies associated with the interaction between TiO<sub>2</sub> NPs and mucus or biopolymer in the presence of gut microbiome in the human GI tract are very limited due to complex biomolecules and microenvironments [47].

Besides the digestive process and the riches of fluids in the GI tract, some studies have reported and effects of NPs ex vivo, in vitro, and in vivo gut cell [5, 50, 211-223]. Studies have described the production of reactive oxygen species (ROS) which can induce toxicity [224-226]. The impact and translocation of TiO<sub>2</sub> NPs through ex vivo, in vivo, and in vitro gut epithelia have been observed by Brun et al., (2014). TiO<sub>2</sub> NPs were likely to translocate via both Peyer's patches and the regular epithelium lining, would cause epithelium disruption, then would be stuck out in intestinal cells and induce chronic damage [50].

### **Conclusions and Perspectives**

The formation of BMC, which constitutes the adsorption of biomolecules on the TiO<sub>2</sub> NPs surface, is influenced by many types of interactions, particularly electrostatic and hydrophobic interactions. These interactions may alter the physicochemical properties and toxicological effect of TiO<sub>2</sub> NPs. The applications of TiO<sub>2</sub> NPs have raised concerns because of their possible toxicity. Some detrimental effects induced by TiO<sub>2</sub> NPs on cells, such as lipid peroxidation, leads to disruption of the bilayers membrane, protein denaturation that induces protein oxidation and enzyme dysfunction, and mitochondria dysfunctions that cause oxidative DNA damage and apoptosis. The production of ROS is considered a primary cause of cell toxicity.

The BMC formations are complicated due to the richness of fluids, such as varying pH and isoelectric point (PI), mechanical forces, surface active components, biopolymers, microbiome, enzymes, and mucus layers. Although some studies have revealed the interaction between TiO<sub>2</sub> NPs and proteins, lipids, and carbohydrates, more comprehensive studies are required, particularly on the mechanistic interactions of TiO<sub>2</sub> NPs with complex molecules such as mucus. Mucus as epithelial barrier provides pivotal protection to living organisms, especially in protecting the respiratory system, GI tract, eyes, and reproductive organs from physical, chemical, and biological hazards.

The physicochemical mechanisms of biomolecules that may affect the GI fate of TiO<sub>2</sub> NPs are entrapment, thickening, and gelling process, adsorption, aggregation, and pH buffer which is specifically owned by proteins. Interactions of NPs with proteins have been studied more rigorously than other biomolecules such as lipids and carbohydrates. Recent studies have examined the interactions between TiO<sub>2</sub> NPs and gut microbiomes, as well as dietary fibers and microbiomes. However, no studies have investigated the interactions between dietary fibers and the microbiome in the presence of TiO<sub>2</sub> NPs. Applied research examining the potential toxicity and associated environment and human health risks of TiO<sub>2</sub> NPs are essential. Research is also needed on the mechanistic interactions between titanium dioxide nanoparticles (TiO<sub>2</sub> NPs) and biomolecules, with a focus on mechanisms that prevent or minimize entry of TiO<sub>2</sub> NPs into the cells or organisms.

## CHAPTER III

### 3. INTERACTIONS OF TiO<sub>2</sub> NPs AND MUCIN INDUCE THE ALTERATION OF PHYSICOCHEMICAL PROPERTIES, AGGREGATION, AND ACCUMULATION OF TiO<sub>2</sub> NPs IN WATER

#### Introduction

Domestic and industrial water effluent contained TiO<sub>2</sub> NPs are released inevitably in aquatic environments such as inland surface water, seas, and ground water. Concerning for nanoparticles (NPs) release, fate and effects on ecosystems and human health are emerging [31, 227-231]. Some studies have reported the assessment of TiO<sub>2</sub> NPs toxicity in aquatic organisms, including the rainbow trout [232-233] and bivalve mollusks [234-237]. TiO<sub>2</sub> NPs can generate radical oxygen species (ROS), and hence to cause toxicity [11, 238]. However, determining the fate, transport, transformation, and toxicity of TiO<sub>2</sub> NPs in the aquatic environment remains challenging and depends on the chemical reactivity and physicochemical property of TiO<sub>2</sub> NPs.

The reactivity and physicochemical property of NPs are strongly influenced by water properties (such as pH, water hardness, ionic composition, and temperature) [16-22], the present organism living in water [17, 29-32], and suspended particulate matter (SPM) [23-28]. Smooth alterations in water chemistry and interactions with surrounding substances can significantly modify NPs surface properties and cause either dispersion or aggregation [17, 239-240]. The presence of NOM is expected to provide a profound impact on the surface properties of NPs (such as coating and surface charge alterations) and the behavior of hetero-aggregation of NPs [25, 27, 34-35].

Naturally, aquatic organisms produce NOM, named mucus (mucin), to protect the hosts from biological, physical and chemical perils [39]. Mucin is the main gel-forming polymers of mucus and consists of core protein domains and densely O-linked oligosaccharide chains; hence, it confers negative charge to the mucin through carboxyl and sulfate groups [36]. Meanwhile, mucus is a complex mixture of water, lipids, salts, nucleic acids, and various proteins, including protein [241-245]. A recent review article has reported the significant effect of electrostatic, steric, and hydrophobic interaction on selective permeability of mucin [37]. Studies, in relation with biological and ecological roles of mucin in aquatic metazoans, have been conducted such as: Atlantic salmon [246-255].

NOM-TiO<sub>2</sub> NPs interaction in water consecutively leads to the formation of biomolecular corona (BMC), hetero-aggregation of BMC that can disrupt to BMC stability in water [34], and dispersion of BMC due to steric hindrances [67, 256]. BMC aggregation is potentially to facilitate the ingestion and accumulation of TiO<sub>2</sub> NPs in the aquatic organisms such as fish and invertebrates. Yin, C., et al. (2019), for instance, suggested that the accumulation of TiO<sub>2</sub> NPs may pass through to the higher trophic levels (particularly to human) via seafood. However, no studies have reported the interaction between TiO<sub>2</sub> NPs and mucin—one of NOMs produced by aquatic animals. Thus, the purpose of this study is to know TiO<sub>2</sub> NPs-mucin interaction, emphasizing on their physicochemical alterations during the formation of BMC in water.

## **Materials and Methods**

### **Materials**

TiO<sub>2</sub> NPs (anatase; 5-15 nm) was purchased from US Research Nanomaterials (TX, US) and mucin (Type II, porcine gastric mucin) was purchased from Sigma-Aldrich Chemicals (St. Louis, MO, USA). The following products were purchased from Life Technologies (Thermo Fisher Scientific, Agawam, MA, USA): hydrochloric acid, sodium hydroxide, sodium chloride, calcium chloride, dimethyl sulfoxide. Silver (Ag) dendrites were prepared based on He, Lin, Li, & Kim (2010) [257]. Deionized water (DIW) was applied for preparation of all solutions.

### **Methods**

The change of physicochemical properties is considered as the main factors affecting the formation of aquatic biomolecular corona (BMC). The formation of BMC was identified by observing the change of physicochemical properties of TiO<sub>2</sub> NPs and mucus (e.g. charge and size) in the function of pH. These alterations may lead to a specific interaction (e.g. electrostatic interaction or hydrophobic interaction) and induce a new formation (e.g. dispersion or aggregation), and thereby change the bioactivity of BMC. Some experiments that have been done to measure these alterations: *first*, observing the appearance of TiO<sub>2</sub> NPs and mucin in the various pH; *second*, examining zeta potential ( $\zeta$ ) of TiO<sub>2</sub> NPs and mucin in the different pH using a dynamic light scattering (DLS); *third*, measuring the change of turbidity (O.D. 600) and aggregation of BMC by utilizing UV-Vis. (wave length 200 nm-750 nm); *fourth*, observing the charge

alteration of BMC by employing dynamic light scattering (DLS); *fifth*, measuring the change of size through by utilizing a static light scattering (Master sizer) and optical microscopy; *sixth*, investigating the change of energy during the formation of BMC by using isothermal titration calorimetry (ITC); *seven*, characterizing the interaction occurs between TiO<sub>2</sub> NPs and mucus through employing surface-enhanced Raman spectroscopy (SERS). In the aquatic environment, water is considered as an important buffer, especially in the formation of biomolecular corona (BMC). Thus, this study used the deionized water (DIW) as the buffer.

#### **Alterations of BMC Appearance & the Basic Experiment (BE)**

0.1% v/v of TiO<sub>2</sub> NPs were dispersed in DIW. 5 mg/ml mucin was stirred into DIW overnight and then centrifuged. Before experiments were conducted, pH of TiO<sub>2</sub> NPs and mucin should have been adjusted. The change of BMC appearance was observed every time a certain amount of TiO<sub>2</sub> NPs was added. The experiment was started with observing the appearance of control (mucin) by adding 1 mL mucin (5mg/mL) into a glass reaction tube and add 9 mL DIW. While treated groups, the amount of TiO<sub>2</sub> NPs (0.1% v/v) was increased 0.5 mL and the amount of DIW was reduced simultaneously; therefore, the total volume of sample was 10 mL. Vortex was applied for each tube (10 s), and eventually the alteration of appearance was observed, particularly to identify the possibility of aggregation during TiO<sub>2</sub> NPs-mucin reaction. This experiment was basic experiment (BE) and should be understood before doing other experiments.

### **Particle Size and Charge Characterization**

The particle size of BMC was measured by a static light-scattering instrument (Master sizer 2000, Malvern Instruments, Worcestershire, U.K.), while zeta potential ( $\zeta$ ) was measured by Zetasizer nano ZS series, Malvern Instruments Ltd, Worcesterhire, UK).

### **UV-Vis Spectrophotometry**

TiO<sub>2</sub> NPs and mucin were mixed and prepared as the same to BE. The absorption spectrum was determined using a SpectraMax M2 microplate reader (Molecular Devices, Sunnyvale, CA, U.S.A.).

### **Isothermal Titration Calorimetry (ITC)**

The change of energy ( $\Delta H$ ), when TiO<sub>2</sub> NPs were titrated into either mucin solution or buffer solution (DIW), was measured by an ITC instrument (Microcalorimeter VP-ITC, MicroCal Inc., Northampton, MA, USA). Twenty-nine 10  $\mu$ L aliquot of TiO<sub>2</sub> NPs suspension were injected sequentially into a 1,450  $\mu$ L cell (mucin) initially. The mass ratio of total samples (TiO<sub>2</sub> NPs: mucin) is 0.25.

### **Surfaced-Enhanced Raman Spectroscopy (SERS)**

Mucin and TiO<sub>2</sub> NPs were dispersed and stirred in buffer overnight. Three different categories of solution (mucin, TiO<sub>2</sub> NPs, and the mixture of TiO<sub>2</sub> NPs and mucin with the mass ratio= 0.25) were prepared. These solutions were homogenized for 10 seconds and added Ag dendrites. These mixtures were homogenized for 10 seconds and



sedimented. These sediments were centrifuged and rinsed with buffer three times. Each sediment (2  $\mu$ L) was deposited on a glass slide covered with aluminum foil as a sample. The sample was air-dried before observing it under the Raman laser. A DXR Raman Microscope (Thermo Scientific, Madison, WI) equipped with a 785 nm-excitation laser and a 50x objective was employed. Spectra were collected with a 5.0 mW laser power and a 50  $\mu$ m slit aperture for 2 seconds scanning time. All SERS experiments were done minimally twice. There were 7-9 spots per sample were characterized by SERS. The elicited spectra were analyzed using TQ analyst software, version 8.0 (Thermo Fisher Scientific).

### **Statistical Analysis**

All experiments were conducted at least twice and reported as the results of means and deviations.

## **Results and Discussions**

### **The change of Physicochemical Properties of TiO<sub>2</sub> NPs & Mucin**

The alteration of physicochemical properties is essential in determine the fate, transformation, transport and toxicity of TiO<sub>2</sub> NPs [258-262]. The physicochemical property of TiO<sub>2</sub> NPs was strongly affected by the pH, as evidenced by the change of their appearance. They were opaquer, when exposed to the more basic microenvironment (**Fig. 9 (A)**). The appearance of TiO<sub>2</sub> NPs at acidic condition (pH 2 to pH 4) were relatively clear and changed gradually. It was becoming cloudier with the increase of pH; while in the neutral to alkaline conditions (pH 6 to pH 10), the appearance of TiO<sub>2</sub> NPs

was all opaque. Knowing the charge is essential to determine the possible interactions caused by the ionic strength such as electrostatic repulsion or attraction. The surface charge data showed that in acidic conditions,  $\text{TiO}_2$  NPs were potentially positive charge; while in the quite neutral to the alkaline condition they had negative charge (**Fig. 10 (A)**). Since the electrostatic attraction at  $\text{pH} > 7$  is not the case, the change of  $\text{TiO}_2$  NPs appearance in DIW may be induced by aggregation. The addition of hydrochloric acid (HCl) was required to reduce pH, while sodium hydroxide (NaOH) was also required to increase pH. The pH adjustment from acidic to alkaline condition could cause the interaction between HCl and NaCl, and thereby develop sodium chloride (NaCl). It could be concluded that the cloudiness can be caused the formation of NaCl, although it is not significant.

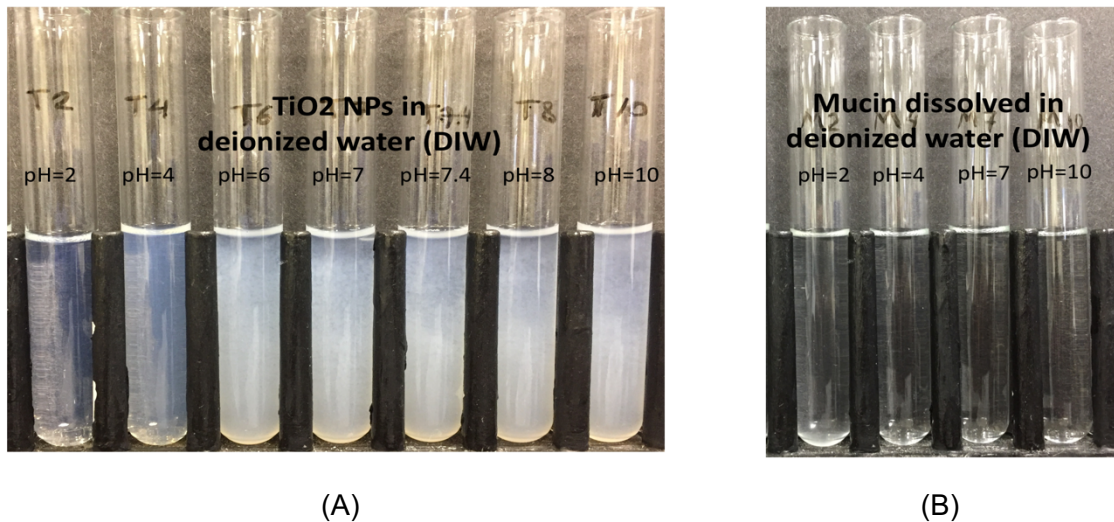


Figure 9: The appearance of  $\text{TiO}_2$  NPs (A) and mucin (B) in deionized water (DIW) at various pH

In aqueous media,  $\text{TiO}_2$  NPs tend to aggregate [64]. The average size of  $\text{TiO}_2$  NPs was much bigger (from 5-15 nm to 30 nm) when they were exposed in deionized water [263]. In theory,  $\text{TiO}_2$  NPs are supposed to aggregate for pH value near the zero points of charge ( $\text{pH}_{\text{pzc}}$ ) [67]. The data showed that the charge of  $\text{TiO}_2$  NPs at  $\text{pH}=6.0$  to  $\text{pH}=7.4$

was low or close to zero ( $<10$ ). This means that  $pH_{pz}$  of  $TiO_2$  NPs in water is more likely 6.0-7.4 and potentially causes aggregation. Whereas, Domingos and co-workers reported that  $pH_{pz}$  for  $TiO_2$  NPs in water was pH range of 4.5-5.2 [67]. This difference is possible and may be induced by the different size range, producer, and preparation of  $TiO_2$  NPs.

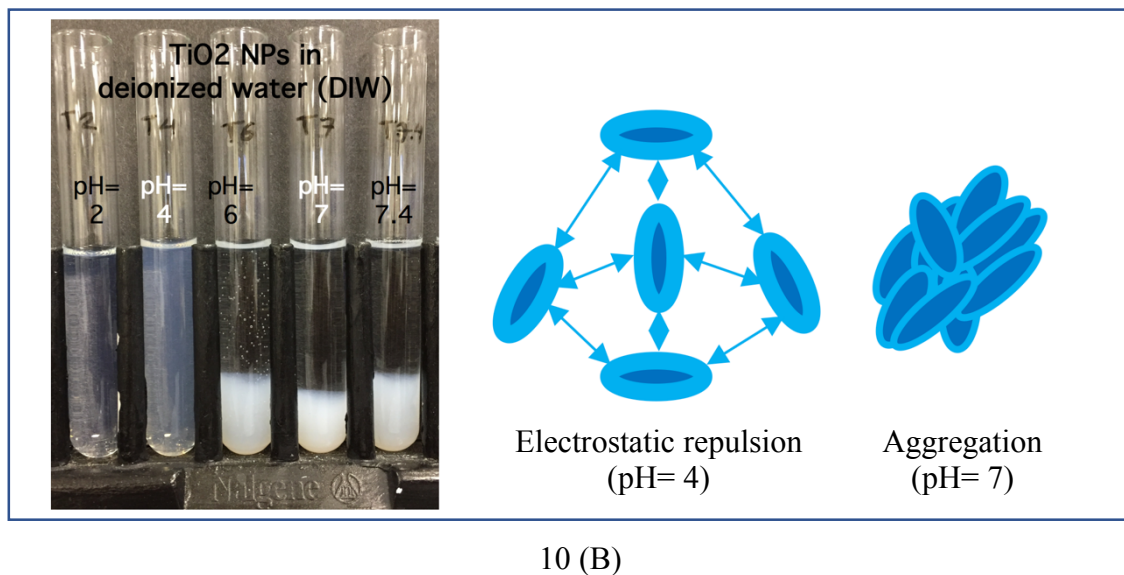
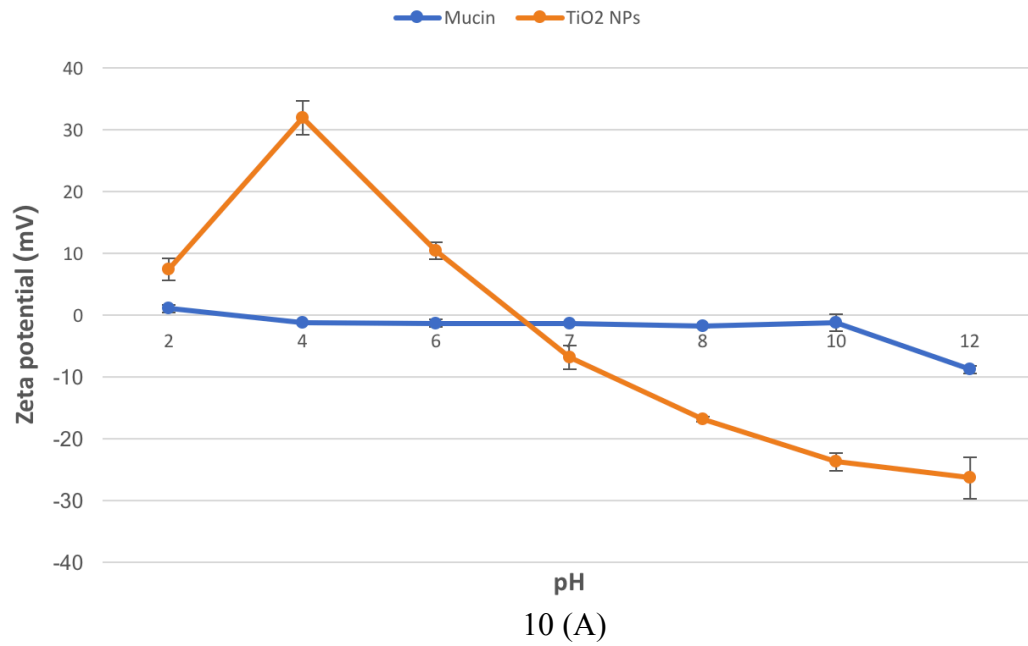
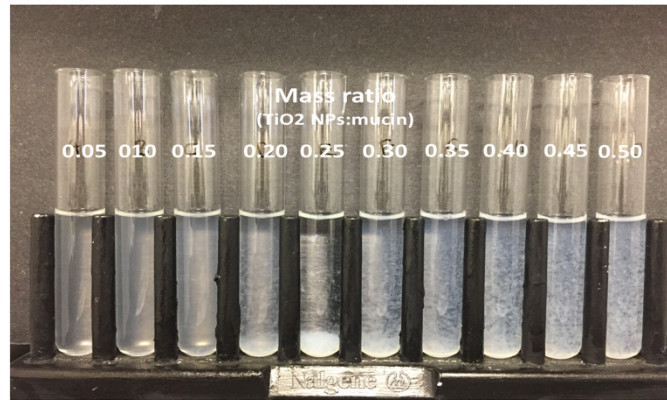


Figure 10: (A) Zeta potential of  $TiO_2$  NPs and mucin in the different pH; (B) The illustration of possible interactions between  $TiO_2$  NPs at pH=4 (electrostatic repulsion) & pH=7 (aggregation).

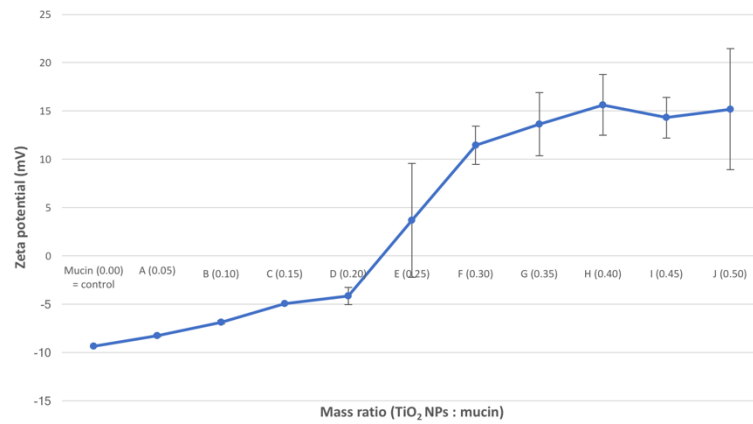
### **Formations of BMC & Hetero Aggregation**

The addition of TiO<sub>2</sub> NPs into mucin induced the formation of BMC. The different concentration of TiO<sub>2</sub> NPs resulted the different level of turbidity. The clarity of control (mucin) was the same to DIW; while the TiO<sub>2</sub> NPs treated groups, the turbidity was gradually increased depending on the amount of TiO<sub>2</sub> NPs added (**Fig. 11 (A) & (C)**). Therefore, it concluded that the more TiO<sub>2</sub> NPs added is the higher turbidity resulted. Interestingly, a massive aggregation was occurred in the mass ratio 0.25; and since that point, the aggregate was relatively stable with the addition of TiO<sub>2</sub> NPs (**Fig. 11 (B)**).

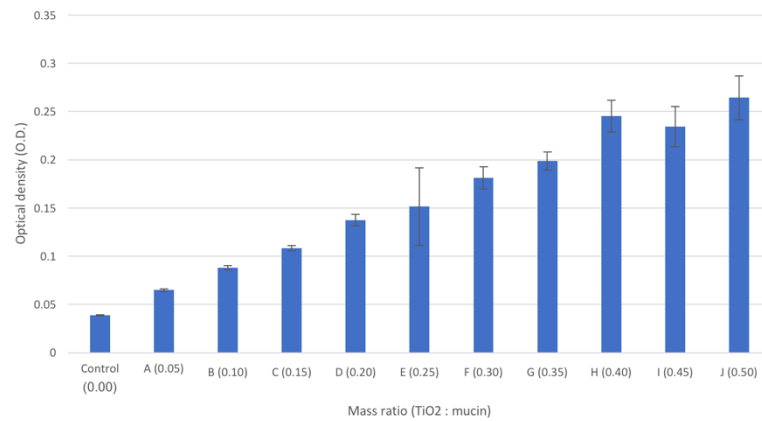
Also, the hydrophobic interaction might affect significantly on TiO<sub>2</sub> NPs-mucin interactions. Zeta potentials of BMC treated with low concentrations of TiO<sub>2</sub> NPs (mass ratio 0.05 to 0.20) were positively to neutrally charged. Meanwhile, zeta potentials were changed from neutral to positive charge, since at mass ratio (TiO<sub>2</sub> NPs: mucin) = 0.25 (**Fig. 11 (B)**). Generally, the formation of BMC at pH=4 is likely to be induced by hydrophobic interactions, and affected by the amount of TiO<sub>2</sub> NPs added.



**11 (A)**  
Zeta potential ( $\zeta$ ) of BMC



**11 (B)**  
Turbidity (O.D. 600)



**11 (C)**

Figure 11: Formations of BMC at pH=4: (A) The massive aggregation may occur in the mass ratio (TiO<sub>2</sub> NPs: mucin) = 0.25; (B) zeta potential alterations of BMC (from negative to positive charge) are affected by the amount of TiO<sub>2</sub> NPs added into the interaction; (C)

The significant increase of BMC size might indicate a massive hetero aggregation. Hetero aggregation can be described as the aggregation of dissimilar particles, and can render significant effect on NPs fate, uptake, transport, bioavailability, and eco-toxicity in environment [231, 264-265]. The size augmentation of aggregate was observed through observing under optical microscopy (**Fig. 12 (A)**) and applying Master sizer depicted in **Fig. 12 (B)**. **Fig. 12 (B)** showed that aggregations were started at C ( $\text{TiO}_2$  NPs: mucin=0.15), and massively developed since E ( $\text{TiO}_2$  NPs: mucin=0.25). Smaller aggregates were identified at J ( $\text{TiO}_2$  NPs: mucin=0.5), and might be caused by separation phase. Further observation to identify separation phase using TEM might be required.

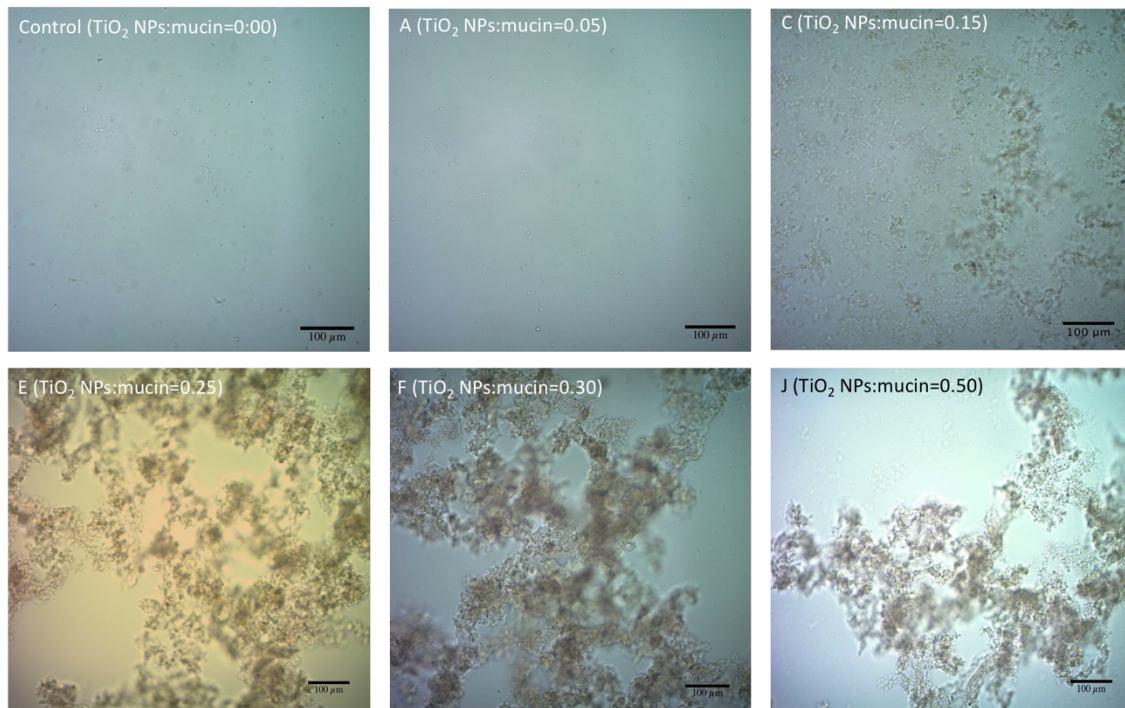


Figure 12: Observations of BMC under optical microscopy. The massive aggregation occurs at mass ratio ( $\text{TiO}_2$  NPs: mucin) 0.25.



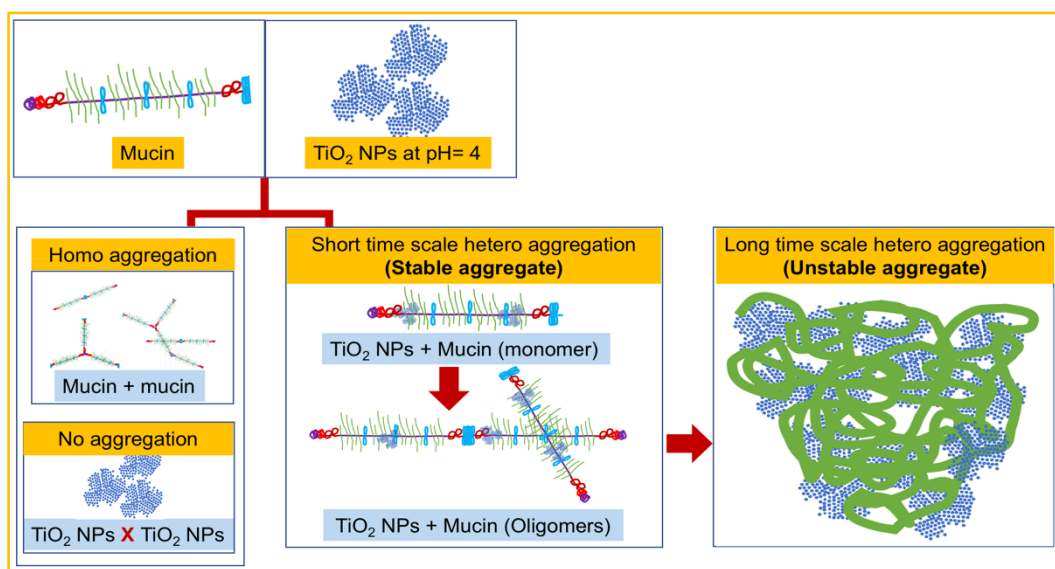
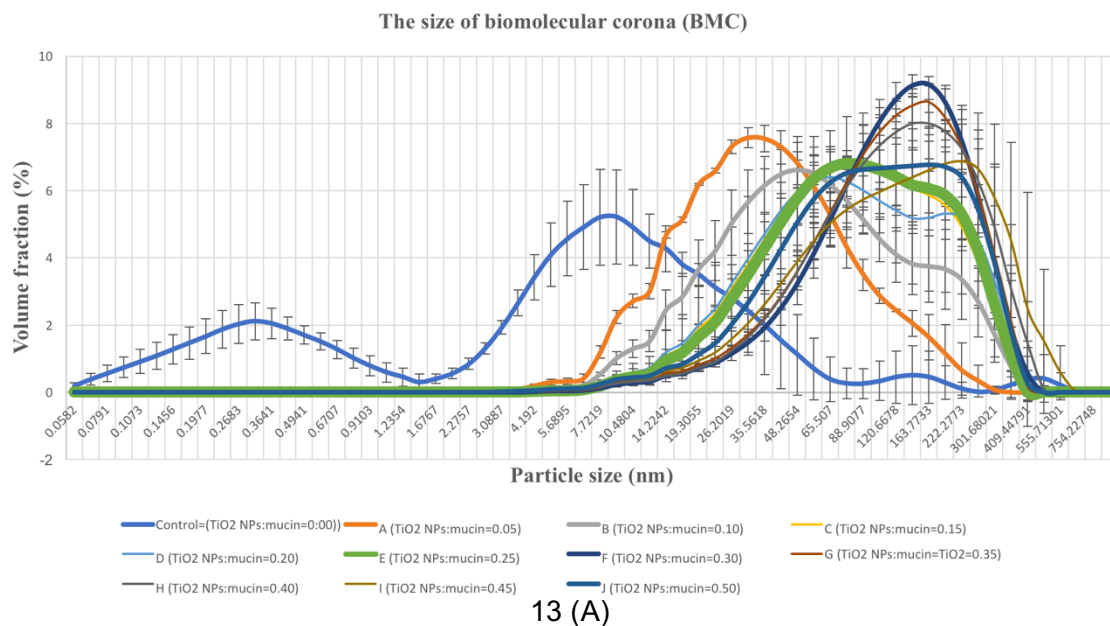


Figure 13: (A) Size measurements using Master sizer show that the size of BMC tends to be bigger with the more addition of TiO<sub>2</sub> NPs; (B) Illustrations of hetero aggregation between TiO<sub>2</sub> NPs and mucin at pH= 4, which is inspired by Clavier, Praetorius, & Stoll, (2019).

Respective of the size, the control or mucin was poly-dispersed molecule, as evidenced by the presence of some peaks with the wide range of size (Fig. 13 (A)). It confirmed previous theory that the size of mucin is varies (200 KDa-200 MDa) [266]. While the treated groups (mucin + TiO<sub>2</sub> NPs) showed that the addition of different amount TiO<sub>2</sub> NPs induced the various size of BMC. Experiment A, for example, showed although the amount of TiO<sub>2</sub> NPs added was very little, but it exhibited the peak shift. It means that the size is much bigger than the control. Fig. 13 (A) also exhibited the shifting peak was persistent unless the mass ratio of TiO<sub>2</sub> NPs and mucin reached 0.25. Once their mass ratio reached 0.25 (E), not only the shifting peak but also the formation of two peaks occurred. It is interesting to note that these phenomena may represent aggregations. The maximum size gained by experiment F; afterwards, the size and volume fraction of BMC were decreased. These reductions might be caused by a phase separation, and resulted an insoluble aggregate. In general, the size of BMC became bigger, when the higher amount of TiO<sub>2</sub> NPs added into mucin, and the mass ratio 0.25 to 0.30 were a critical point, where the maximum size may be achieved (**Fig. 12 & Fig. 13 (A)**).

**Fig. 13 (B)** illustrated the development of hetero aggregation between TiO<sub>2</sub> NPs and mucin. This illustration is inspired by **Fig. 12 & Fig. 13 (A)** and **Fig. 11 (A)** showed enormous aggregates and a massive precipitation. Interactions of some similar molecules/particles are an example of mono aggregation such as dimer and trimer; while hetero aggregation is categorized into two types of aggregation: stable hetero aggregation (which is formed by dimer/trimer aggregate) and unstable hetero aggregation (which is



developed by large and complex aggregate) [34]. Therefore, it can be noticed that the interaction between TiO<sub>2</sub> NPs and mucin is likely to generate unstable hetero aggregation.

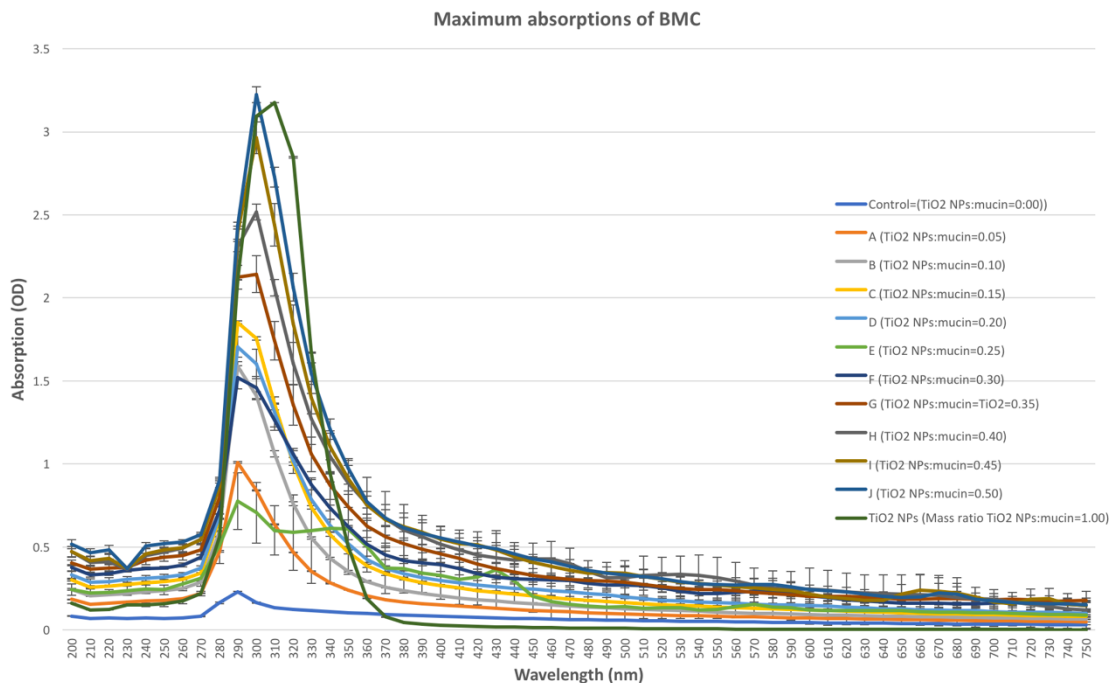


Figure 14: UV-Vis absorbance spectra of BMC with different concentrations of TiO<sub>2</sub> NPs (wavelength 200-750 nm). At E (TiO<sub>2</sub> NPs: mucin= 0.25), there are multiple peaks, hypochromic effect, and redshift which may indicate the massive aggregation.

Furthermore, UV-Vis spectroscopy was employed to examine the formation of hetero aggregates by scanning the maximum absorption of BMC with the range of wavelength 200-750 nm. **Fig. 14** exhibited that the more TiO<sub>2</sub> NPs added was likely to create more bounds between TiO<sub>2</sub> NPs and mucin, then led to aggregation. This was evidenced by the increase of BMC absorbance. The spectrum absorbance of control groups did not show any absorbance at wave length longer than 400 nm (mucin) and 390 nm (TiO<sub>2</sub> NPs); meanwhile the spectrum of mucin exposed with TiO<sub>2</sub> NPs showed a vivid absorbance at 400 nm even at >600 nm, depending on the mass ratio of TiO<sub>2</sub> NPs to

mucin. In general, the treated groups resulted in the red shift of their absorbance spectra. It suggests that the molecular interaction and direct binding may occur. It is an interesting note that the red shift was persistent to all treated groups.

However, there was an exception shown by the experiment E (the mass ratio of  $\text{TiO}_2$  NPs to mucin = 0.25): a hypochromic effect was significantly appeared, and there were multiple peaks found when the mass ratio reached 0.25 (**Fig. 14**). The red shift and hypochromic effect accompanied by multiple peaks may be an indication that the unstable hetero aggregation massively takes place at mass ratio ( $\text{TiO}_2$  NPs: mucin) = 0.25. In theory, UV-Vis spectrophotometry is commonly applied to detect aggregation due to electron transfer from an absorbed molecule to empty conduction band on  $\text{TiO}_2$  which induce the decrease of ionization energy (IE) of electron donor [267]. Aggregation can be identified through a bathochromic shift of UV-vis spectra (the red shift) [268]. Therefore, it was concluded that the interaction between  $\text{TiO}_2$  NPs and mucin leads to hetero aggregation.

#### **$\text{TiO}_2$ NPs-Mucin Interaction: Polyvalent Binding and Irreversible Aggregate**

Although mucin is mostly constituted by carbohydrate, mucin's backbone consists of apomucin—the apoprotein of mucin [36]. Although extensive studies have observed protein-NPs interactions [98], the mechanistic interaction between  $\text{TiO}_2$  NPs and mucin limited information have been unclear.  $\text{TiO}_2$  NPs may interact with apomucin, and induce the alteration of mucin structure. Studies revealed that the mucin aggregate could be constituted by five-, six-, and seven-sided ring structure; where these ring structures were composed by multiple mucin trimers [269]. Mucin trimers (**Fig. 15**) were formed in low

and high pH conditions [269]. Mucin can build net-like structure through hydrophobic cross linking, and interact with other biomolecules or particles via some possible interactions such as hydrophobic interactions, hydrogen binding and electrostatic interactions (**Fig. 15**) [36, 270-271].

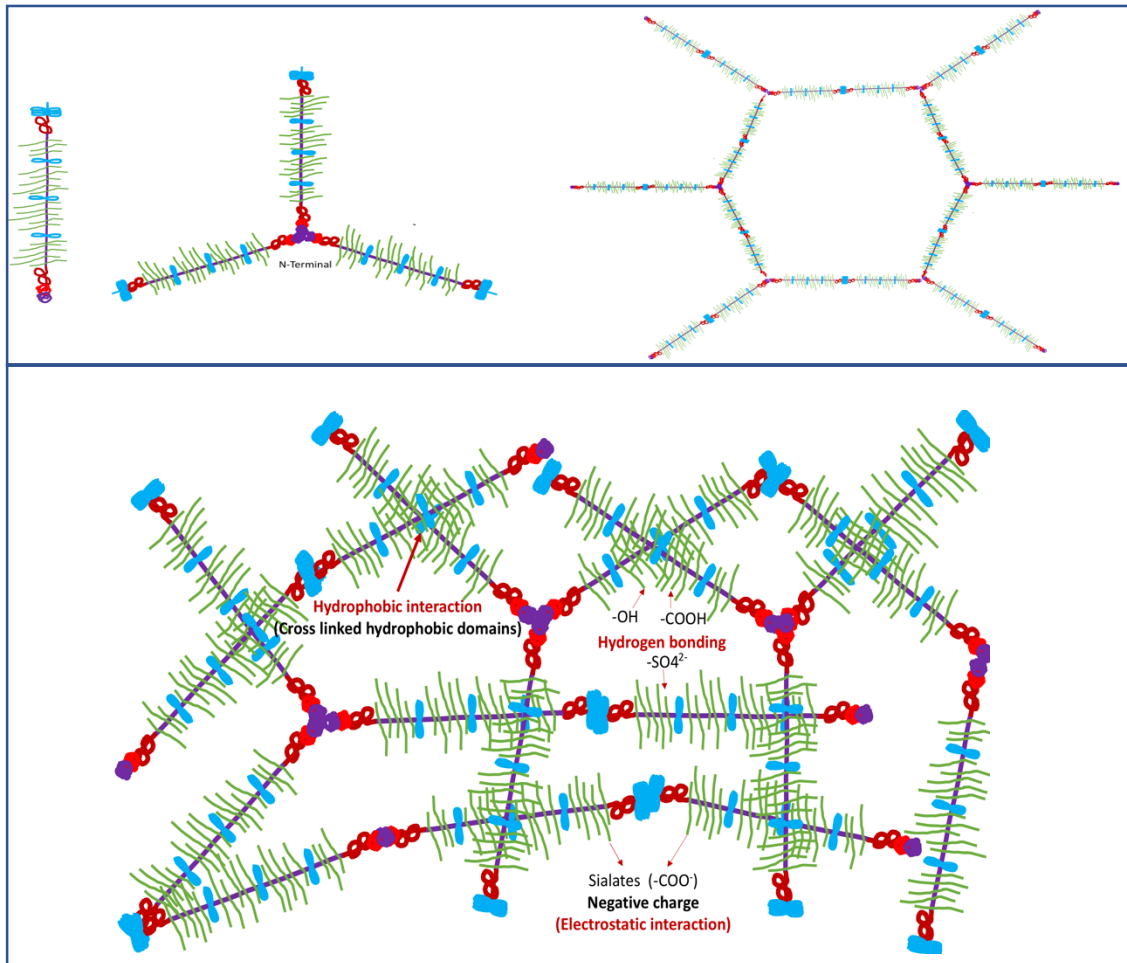


Figure 15: Illustrations: The structure of mucin which is like a tube brush (inspired by Bansil, & Turner (2006)), and mucin trimer and six sided-ring structure which are inspired by Ambort et al., (2012). Possible interactions of mucin with biomolecules or inorganic particles which are inspired by Wagner et al., (2017); Yang et al., (2012); & Bansil & Turner, (2006).

Some studies have reported their findings related to interactions between  $\text{TiO}_2$  NPs and natural organic matter (NOM) in aquatic environment such as humic acid/fulvic acid [67, 255]. They explained that once  $\text{TiO}_2$  NPs are exposed into water, their stability is reduced, and hence to lead to agglomeration. Besides,  $\text{TiO}_2$  NPs will adsorb NOM existing in aquatic environment, form BMC, increase their stability owing to steric hindrance, and cause dispersion. The addition of electrolytes lead to disrupt the stability of BMC due to charge neutralization, depletion, and bridging effect [67, 255]. Mucin is another essential NOM generated by aquatic organisms to elicit protection from physical, biological, chemical hazards; however, mechanistic interactions between  $\text{TiO}_2$  NPs and mucin in aquatic environment are still questioned.

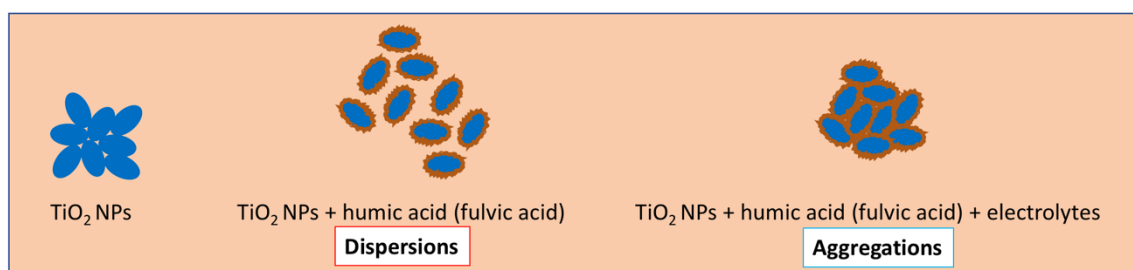


Figure 16: Illustrations of the interactions between  $\text{TiO}_2$  NPs and NOM (humic acid/fulvic acid) in the aquatic environment, which is based on findings by Domingos, Tufenkji, & Wilkinson, (2009) and Wang et al., (2019).

Witten and co-workers (2018) reviewed the biochemical mechanisms emphasizing on mucin penetration and binding, particularly the significance of electrostatic, steric, and hydrophobic interactions [37]. Impact induced by mucin-particle interactions depends on not only the specific biochemistry but also the number of binding sites on the particle which can interact with mucin [272]. The illustration of polyvalent binding between  $\text{TiO}_2$  NPs and mucin was showed by **Fig. 17 (A) & (B)**. The size of mucin was not actual size. It was just the estimated size based on the Mastersizer data

( $\pm 8500$  nm), while the size of  $\text{TiO}_2$  NPs varies depending on pH (aggregation). **Fig.** 17 tried to provide a simple illustration; therefore, the polyvalent binding between  $\text{TiO}_2$  NPs and mucin and the cross linking of mucin were more understandable.

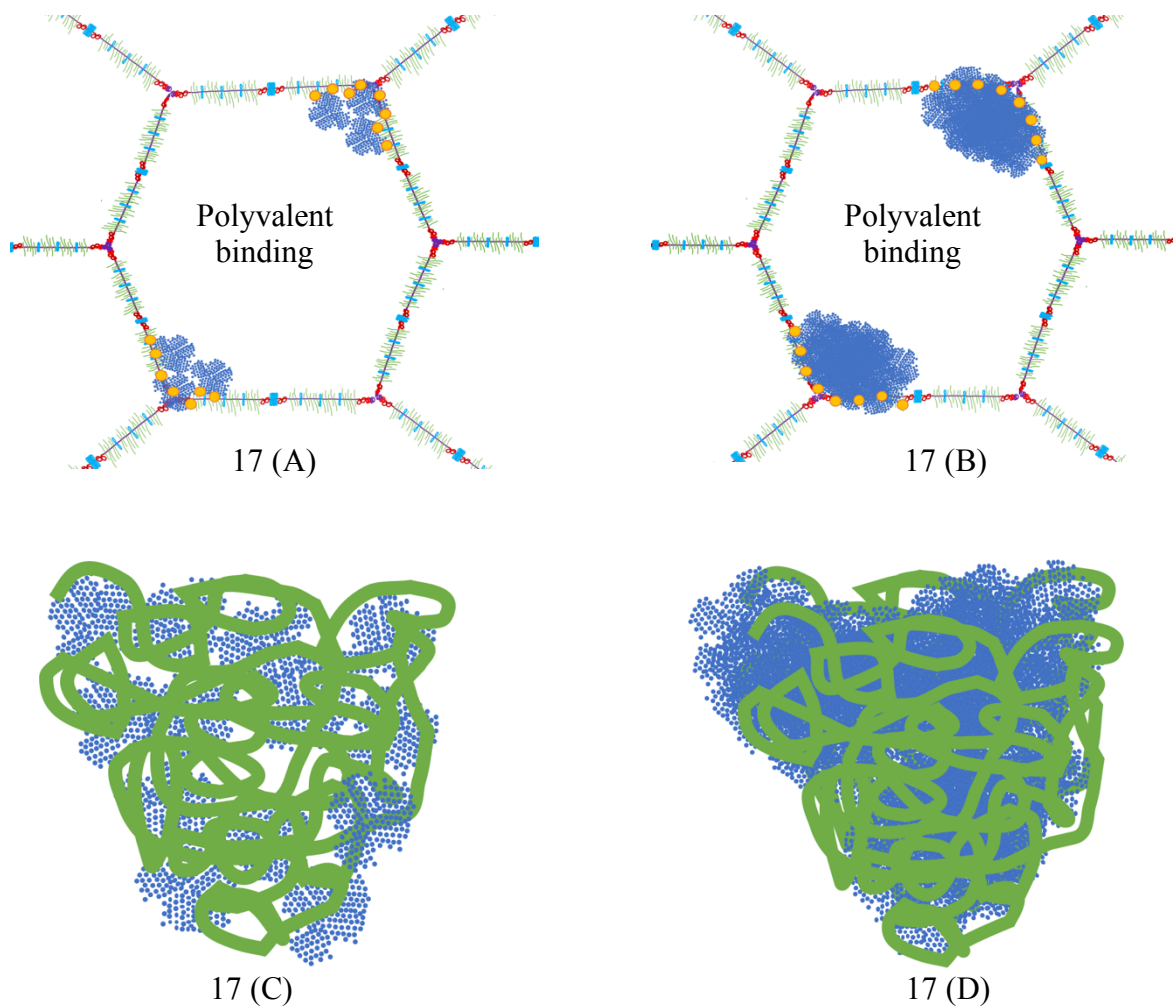


Figure 17: Illustrations of polyvalent binding between  $\text{TiO}_2$  NPs and mucin at pH=4 (A) & pH=7 (B) (inspired by Cone, R.A., (2009)). Effect of polyvalent binding and hydrophobic cross-linking on the formation of irreversible aggregate at pH=4 (C) & pH=7 (D).

Based on the measurement of turbidity (**Fig.** 11 (A) & (B)), size (**Fig.** 12 & **Fig.** 13 (A)), UV-Vis maximum absorbance (**Fig.** 14 (A) & (B)), and exothermic energy (**Fig.** 18), interactions between  $\text{TiO}_2$  NPs and mucin caused irreversible aggregates. It can be

noticed that polyvalent binding may play important role in the formation of irreversible aggregate (**Fig. 17 (A) & (B)**). In theory, almost all NPs own multi binding sites; therefore, mucin can reduce the diffusivity of NPs more than small particles. Although each binding is weak; the total effect of polyvalent binding is likely to induce near-irreversible binding [37, 272]. Multiple bindings presented by NPs are termed polyvalent binding [272]. Besides, the hydrophobic cross linking that form net-like mucin structure [36, 271] is potentially to facilitate the entrapment of TiO<sub>2</sub> NPs.

One of the common approaches to determine the type of energy change and binding during the interaction between biomolecule and ligand is ITC. The combination of thermodynamic data and structural biomolecule has enhanced the understanding of macromolecular interactions in solutions (particularly protein), and thereby allowing predictions in terms of size, thermodynamic properties of binding interface, protein solvation, and conformational changes of protein [273-274]. It was difficult to determine the number of binding sites from **Fig. (18)**. Based on the theory **Fig. (18)** showed a multiple binding [275]. It means that three or even more ligand molecules (TiO<sub>2</sub> NPs) bind a receptor molecule (mucin). The highest affinity binding site has the largest exothermic enthalpy change, while the next two or more binding processes have overlapping the equilibrium constant ( $K$ ) values, and thereby they indistinguishable [275]. It may be induced by a complex sequence of modifications that are not being fully understandable when TiO<sub>2</sub> NPs interact with mucin.

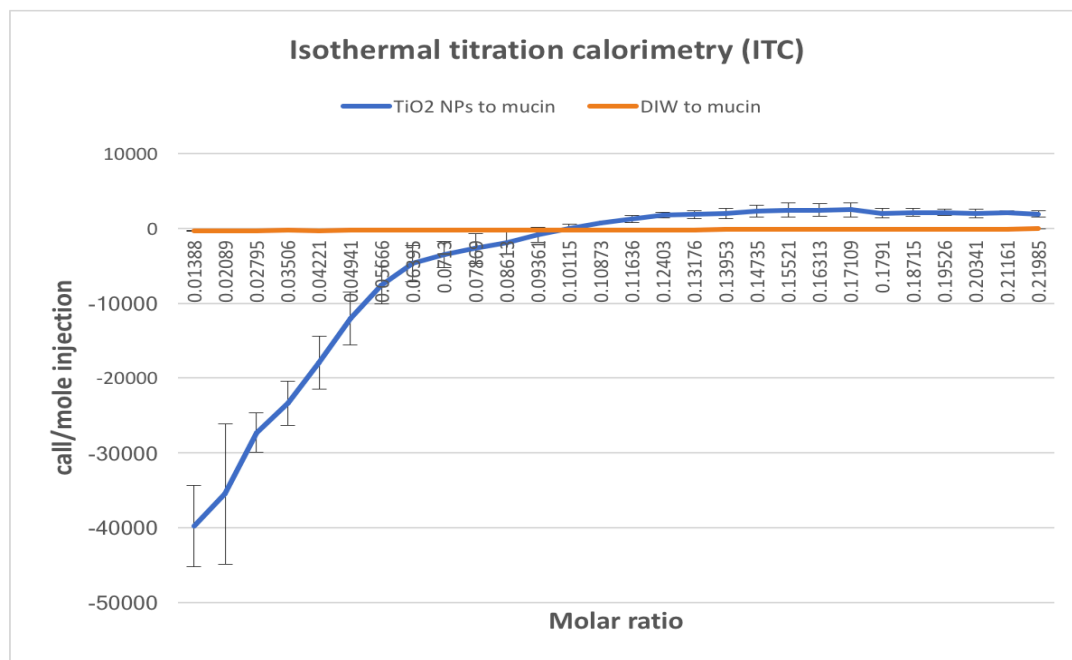


Figure 18: Isothermal titration calorimetry (ITC) shows that exothermic energy. It indicates that bonds are formed and energy is released.

**Fig. 18** showed that the reaction of  $\text{TiO}_2$  NPs and mucin resulted exothermic energy and insignificant endothermic energy. The exothermic energy was predominantly generated from the first to thirteenth injection; afterwards, a small endothermic energy was consistently generated until the last injection. The surface properties of NPs and protein stability provide a notable effect to the alteration of protein structure and its function [274]. The disturbance of the protein structure is usually escorted by an inclusion of additional energy to the system, and thereby to shift the protein structure from the “potential hole” [274]. In thermodynamic terms, a minimal entropy is the most favorable state of the polypeptides chain. Entropy influences the intrinsic stability of protein. If the entropy of the “alternative state” is lower than the “native state”, it will induce a reversible transition. However, if the entropy of the “alternative state” reaches

the critical maximum value, protein will be denatured and form insoluble aggregates or fibrils due to the irreversible transition. The reversible transition indicates the formation of “soft corona”; while the irreversible transition results “hard corona” [276].

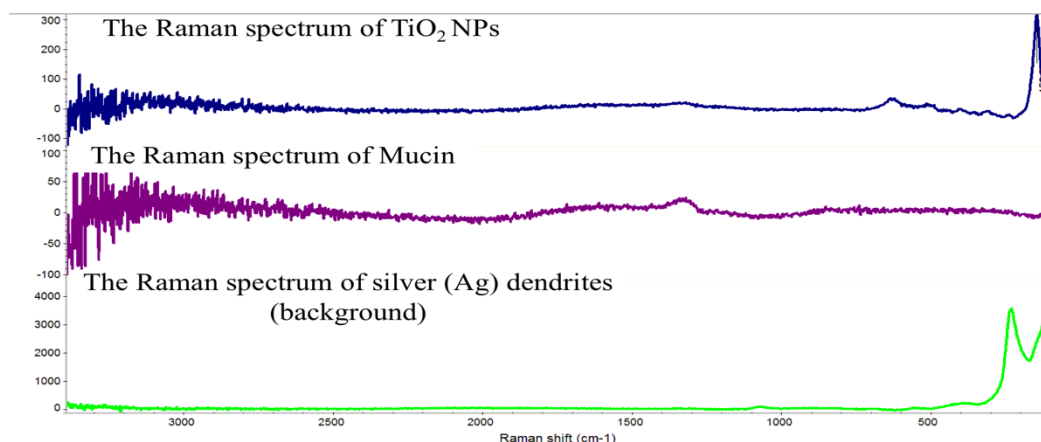
Considering **Fig. 11, 12, 13, 14, & 18**, it can be noticed that the aggregation induced by TiO<sub>2</sub> NPs-mucin interaction forms “hard corona” and it is irreversible.

### **SERS and TiO<sub>2</sub> NPs-Mucin Interaction**

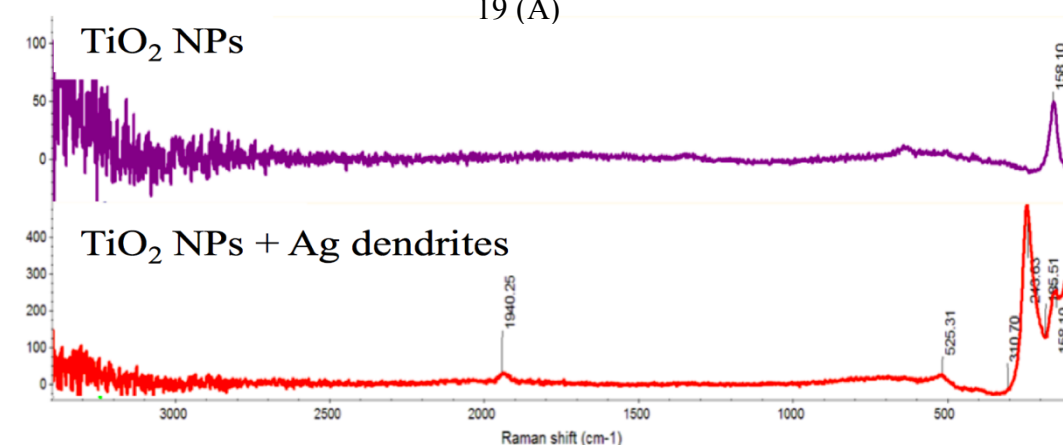
#### **Silver (Ag) Dendrites and the Enhancement of TiO<sub>2</sub> NPs & Mucin Spectra**

Regarding SERS enhancement mechanisms, NPs can be classified into three groups such as NPs which support electromagnetic enhancement, chemical enhancement, and neither electromagnetic nor chemical enhancement [277]. TiO<sub>2</sub> NPs are included in the second group, where SERS enhancement is attributed with the charge transfer (CT) between NPs and adsorbed biomolecule ligands [278]. Silver (Ag) dendrites show a high consistency and satisfactory performance with an analytical enhancement factor  $\sim 10^4$  [257].

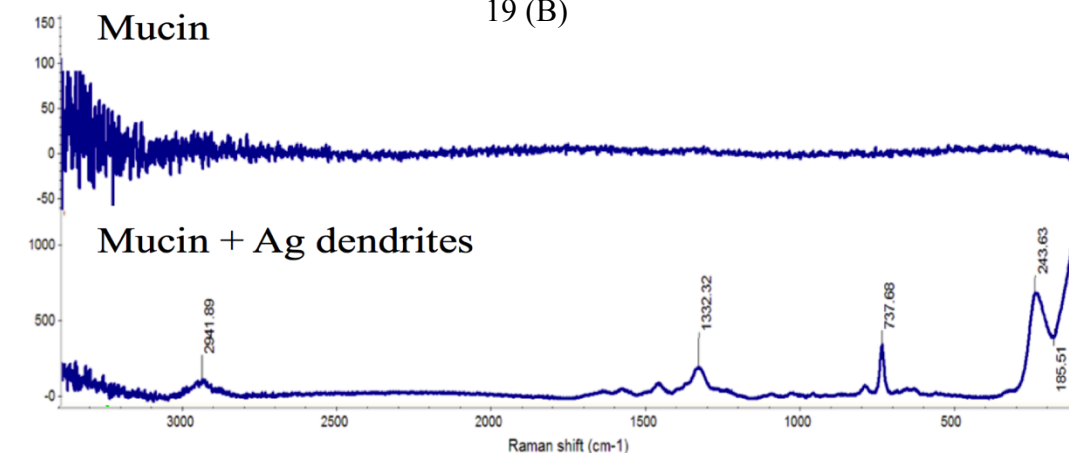




19 (A)



19 (B)



19 (C)

Figure 19: The spectral peaks of  $\text{TiO}_2$  NPs, mucin and Ag dendrites (background) (A); Ag dendrites enhance the spectral peaks of  $\text{TiO}_2$  NPs (B) and mucin (C).

The specific spectrum of every chemical substance used in the experiment is essential to be known, particularly the background (Ag dendrite). The background-result comparison will provide an idea about what happens during the experiments. TiO<sub>2</sub> NPs showed the spectral peak at wave number 158.51 cm<sup>-1</sup>; while mucin did not exhibit a significant spectral peak. Ag dendrites represented the specific peak at wave number 243.63 cm<sup>-1</sup> (**Fig. 19 (A)**).

The spectral peaks of mucin which interacted with Ag dendrite, as substrate, was significantly enhanced (**Fig. 19 (C)**). The spectral peak of TiO<sub>2</sub> NPs (at wave number 158.51 cm<sup>-1</sup>) was also enhanced by Ag dendrites, approximately 4 times higher than TiO<sub>2</sub> NPs without Ag dendrites (**Fig. 19 (B)**). Several main peaks of mucin were depicted clearly when mucin interacted with Ag dendrites such as at wave number  $\pm 737.68$  cm<sup>-1</sup>,  $\pm 1332.32$  cm<sup>-1</sup>, and  $\pm 2941.89$  cm<sup>-1</sup> (**Fig. 19 (C)**). Assignments of Raman bands (at wave number  $\pm 737.68$  cm<sup>-1</sup>,  $\pm 1332.32$  cm<sup>-1</sup>, and  $\pm 2941.89$  cm<sup>-1</sup>) might be close associated with N-acetyl-D-glucosamine (GlcNAc) and D-(+)-galactosamine (GalNAc) [279-281]. GalNAc and Glc NAc are a type of amino-sugar which is secreted by mucin producing-epithelial cell in the GI tract [282]. Studies have characterized and analyzed spectral feature of biomolecules including mucin, generally, at wave number 0 to 2000 cm<sup>-1</sup> [279-281, 283-285].

The main aim of SERS experiments was to determine the development of hetero aggregation by characterizing the interaction between TiO<sub>2</sub> NPs and mucin. This study preferred using the spectral peak of mucin at wave number  $\pm 2941.89$  cm<sup>-1</sup> than using the spectral peaks at wavenumber  $\pm 0$ -2000 cm<sup>-1</sup> for two reasons. First, GlcNAc and GalNAc generally showed a higher spectrum at a wave number  $\pm 2941.89$  cm<sup>-1</sup> than at wave

number  $\pm 1332.32 \text{ cm}^{-1}$  or  $\pm 737.68 \text{ cm}^{-1}$  [281]. In other words, using a higher spectral peak to determine hetero aggregation between  $\text{TiO}_2$  NPs and mucin was more practical or efficient than the lower ones. Second, based on this study, the spectral peak of mucin at wave number  $\pm 2941.89 \text{ cm}^{-1}$  was not only more efficient but also more consistent when  $\text{TiO}_2$  NPs interacted with mucin in various pH and buffer (Fig. 21 (A) and 33 (A)). Therefore, the preference of this study went to the spectral peak at wave number  $\pm 2941.89 \text{ cm}^{-1}$ .

A recent study characterized the interaction between  $\text{TiO}_2$  NPs and mucin using SERS [285]. Their preference was to analyze the  $\text{TiO}_2$  NPs-mucin interaction at wave number  $\pm 0\text{-}2000 \text{ cm}^{-1}$ . The spectral peaks were characterized in their study were similar to this study with a little bit different. It is normal due to  $\text{TiO}_2$  NPs used in the experiments might be different such as the type and size of  $\text{TiO}_2$  NPs. They preferred using the solid  $\text{TiO}_2$  NPs (powder) for the original sample. Meanwhile, this experiment preferred to use the liquid ones with the original size range between 5-15 nm. They only used SSF, while this experiment used DIW and PBS as a buffer.

**Fig. 20** depicted two sets of experiments to examine the interaction between Ag dendrites and  $\text{TiO}_2$  NPs or mucin: 1. ( $\text{TiO}_2$  NPs + Ag dendrites) + mucin, 2. (mucin + Ag dendrites) +  $\text{TiO}_2$  NPs. The results showed that these two experiments had similar results (**Fig. 20 (A; III)** & **Fig. 20 (B; III)**). The specific spectral peaks  $\text{TiO}_2$  NPs and mucin were persistently appeared and enhanced (**Fig. 20 (A; III)** & **Fig. 20 (B; III)**); it means that Ag dendrites are a good substrate for both  $\text{TiO}_2$  NPs and mucin. Also, the principal component analysis (PCA) of mucin and  $\text{TiO}_2$  NPs represented a wider area when they interacted with Ag dendrites. The PC score of mucin was magnified many times

compared to mucin alone (**Fig. 20 (C; O) & (D; O)**). These findings confirmed the previous study that Ag dendrites are a good substrate [257].

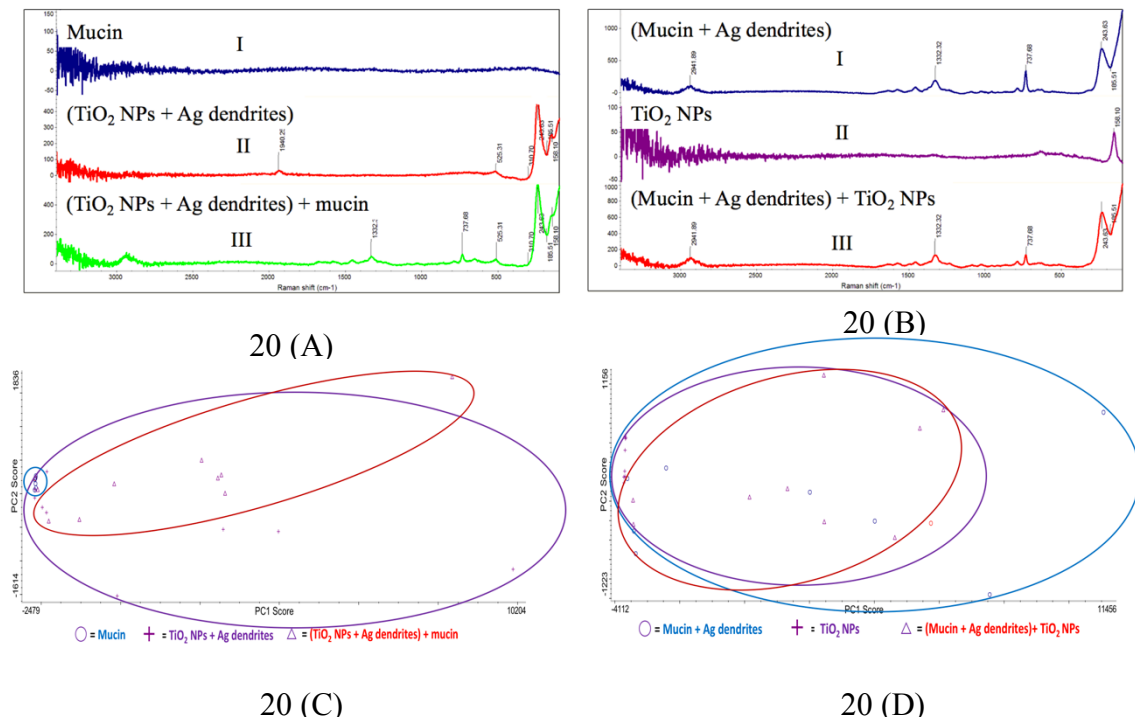
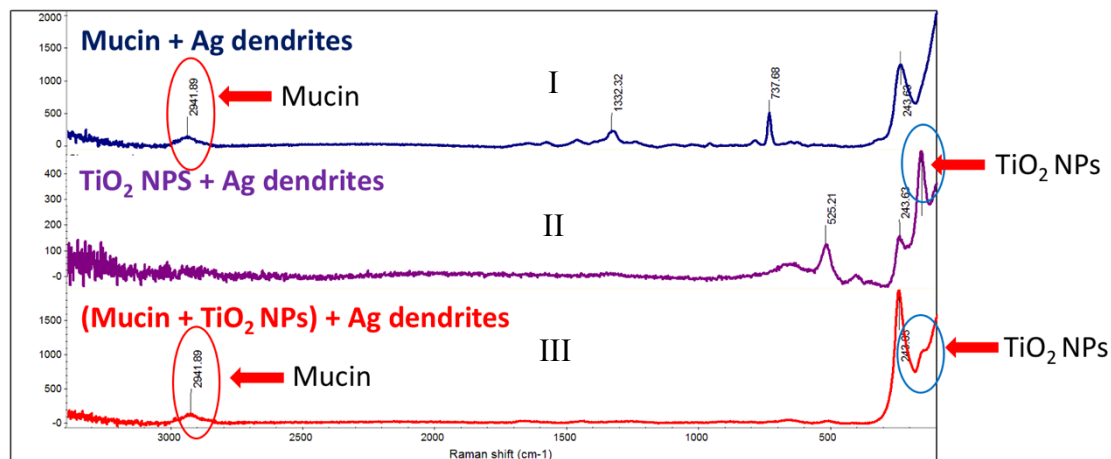


Figure 20: (A) The spectral peaks of mucin (I), TiO<sub>2</sub> NPs + Ag dendrites (II), and (TiO<sub>2</sub> NPs + Ag dendrites) + mucin (III); (B) The spectral peaks of mucin + Ag dendrites (I), TiO<sub>2</sub> NPs (II), (mucin + Ag dendrites) + TiO<sub>2</sub> NPs (III); (C) PCA of mucin (O), TiO<sub>2</sub> NPs + Ag dendrite (+), and (TiO<sub>2</sub> NPs + Ag dendrites) + mucin (Δ); (D) PCA of mucin + Ag dendrites (O), TiO<sub>2</sub> NPs (+), (mucin + Ag dendrites) + TiO<sub>2</sub> NPs (Δ).

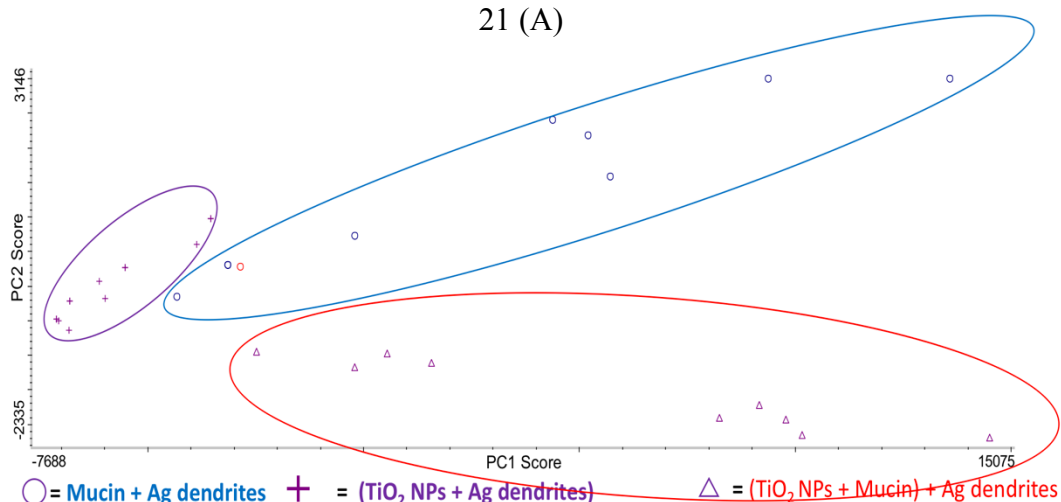
### Characterizations of TiO<sub>2</sub> NPs-Mucin Interaction

The characterization of TiO<sub>2</sub> NPs-mucin interaction had some steps: First, mixing TiO<sub>2</sub> NPs with mucin, then adding Ag dendrites ((TiO<sub>2</sub> NPs + mucin) + Ag dendrites). The results exhibited that the spectral peaks of both TiO<sub>2</sub> NPs and mucin appeared (**Fig. 21 (A; III)**). It indicates that the interaction between TiO<sub>2</sub> NPs and mucin occurs. Second, Analazing of PC score. The results depicted three distinctive clusters: TiO<sub>2</sub> NPs cluster, mucin cluster, and BMC cluster. It is interesting to note that the TiO<sub>2</sub> NPs-mucin

interaction may alter the synthetic identity of TiO<sub>2</sub> NPs and the biological identity of mucin.



21 (A)



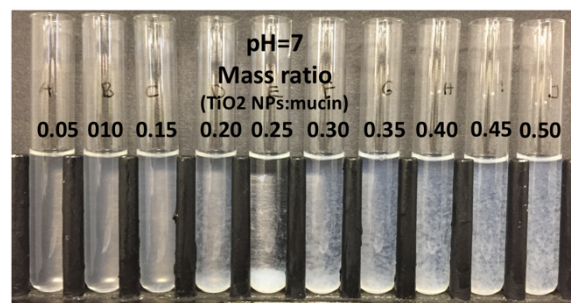
21 (B)

Figure 21: (A) the spectral peaks of mucin, TiO<sub>2</sub> NPs, and BMC; (B) PCA of mucin, TiO<sub>2</sub> NPs, and BMC.

### TiO<sub>2</sub> NPs-Mucin Interaction at Neutral Condition (pH=7)

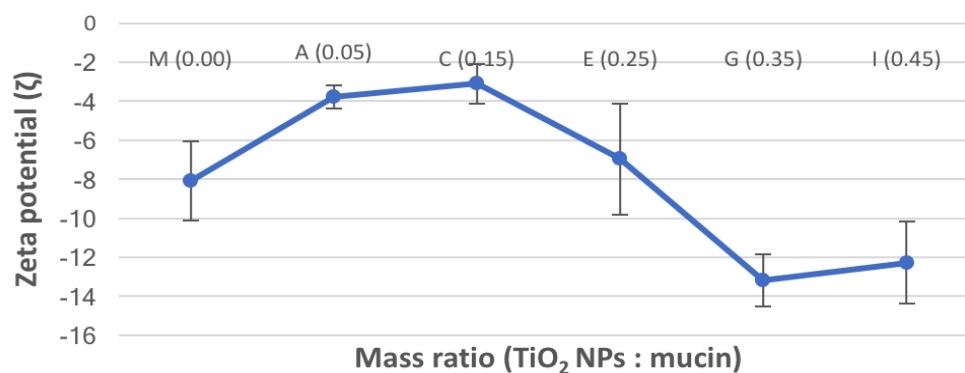
This study also observed TiO<sub>2</sub> NPs-mucin interaction in neutral condition (pH=7). The result showed that although the charge of both TiO<sub>2</sub> NP and mucin were negative (Fig. 22 (B)); aggregations were still observed (Fig. 22 (A)). Zeta potentials of BMC

with different concentrations of TiO<sub>2</sub> NPs presented negative charge, although the charge is relatively low (-2 to -15). In theory, aggregations are supposed to be occurred where molecules/particles at pH<sub>pzc</sub>, which is pH where the charge is near zero [67]. It suggests that in aquatic environment, hydrophobic interactions may play significant role in the interaction between TiO<sub>2</sub> NPs and mucin. Therefore, the use of Nano-isothermal titration calorimetry (nITC) coupled with other methods (such as fluorescence spectroscopy and circular dichroism) to get detailed information about the interaction types, reaction products, alterations on TiO<sub>2</sub> NPs surface, and the possible consequences of TiO<sub>2</sub> NPs-mucin interaction on aquatic environment may be required.



22 (A)

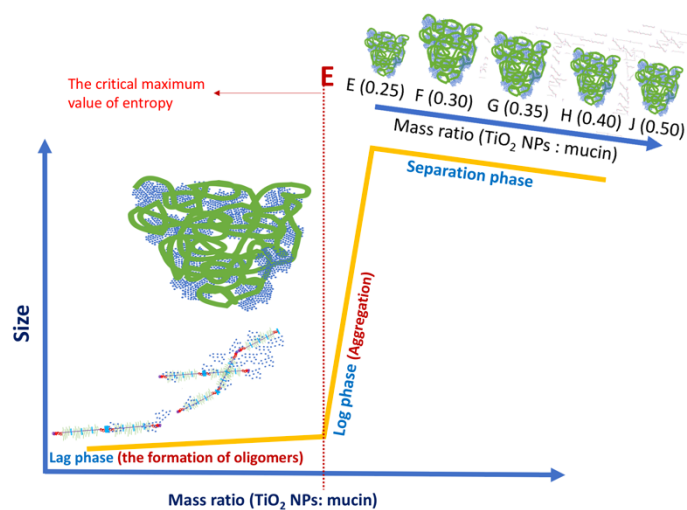
### Zeta potential ( $\zeta$ ) of BMC at pH=7



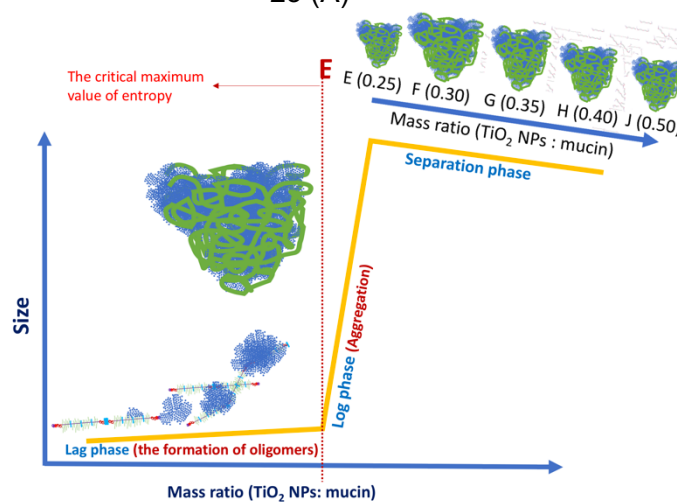
22 (B)

Figure 22: (A) Aggregations still occur; although TiO<sub>2</sub> NPs & mucin have negatively charge; (B) Zeta potential of BMC at pH=7.

In summary, the mechanistic interactions between  $\text{TiO}_2$  NPs and mucin in water can be illustrated by **Fig. 23**. Initially, mucin oligomers might be developed; the more addition of  $\text{TiO}_2$  NPs in water induced the more adsorption of mucin on the surface of  $\text{TiO}_2$  NPs via various types of interaction; small aggregates were developed unless the mass ratio ( $\text{TiO}_2$  NPs: mucin) reach 0.25. Since this point, the massive hetero aggregation was developed. Eventually, the aggregation was followed by separation phase and resulted irreversible and insoluble aggregates.



23 (A)



23 (B)

Figure 23: Illustrations of  $\text{TiO}_2$  NPs-mucin interaction at pH= 4 (A) & pH= 7 (B).

## **Conclusions**

In conclusion, the physicochemical alteration of TiO<sub>2</sub> NPs is affected significantly by pH. In acidic condition, TiO<sub>2</sub> NPs are conferred by positive charge. They become opaquer and have negative charge when the pH condition is neutral to basic. While the physicochemical properties of mucin are not pH-dependence. At acidic condition (pH=4), the interaction of TiO<sub>2</sub> NPs with mucin forms BMC. The amount of TiO<sub>2</sub> NPs exposed to mucin may play important role in the BMC formation. As the mass ratio of TiO<sub>2</sub> NPs to mucin has reached 0.25, it leads to the irreversible hetero aggregation, which are possibly caused by the polyvalent binding. This is evidenced by the augmentation of the size, the increase of turbidity, the hypochromic effect and red shift of the UV-Vis spectra, and the exothermic energy during TiO<sub>2</sub> NPs-mucin interactions. The interaction between TiO<sub>2</sub> NPs and mucin has been identified using SERS. The distinctive cluster of SERS spectra between TiO<sub>2</sub> NPs, mucin, and BMC, reveals that the exposure of TiO<sub>2</sub> NPs to mucin may transform TiO<sub>2</sub> NPs become new species with different bio-fate. It is interesting to note that although both TiO<sub>2</sub> NPs and mucin have negatively charge at a neutral condition (pH=7), SERS has exhibited that the interaction between TiO<sub>2</sub> NPs and mucin still occurs. Hence, it can be noticed that in the aquatic environment, the hydrophobic interaction may play more significant role in the formation of BMC. Therefore, further studies to observe about the interaction types, reaction products, and the possible consequences of TiO<sub>2</sub> NPs-mucin interaction on aquatic environment are required.



## CHAPTER IV

### 4. INTERACTION BETWEEN TiO<sub>2</sub> NPs AND MUCIN: EFFECTS OF PHOSPHATE-BUFFERED SALINE (PBS) AND pH ON THE ALTERATION OF SURFACE CHANGE, THE FORMATION OF BIOMOLECULAR CORONA (BMC) AND HETERO AGGREGATION

#### Introduction

Since 1966, the US FDA has approved the use of TiO<sub>2</sub> NPs in food up to 1% [286]. TiO<sub>2</sub> NPs have various forms such as rutile, brookite, and anatase [287-288]. Although anatase is more toxic [4] due to the higher photocatalytic activity than the others [287, 289], it has a higher industrial application. The massive application of TiO<sub>2</sub> NPs in food industries rises disputes regarding their safety. The International Agency for Research on Cancer (IARC) has categorized TiO<sub>2</sub> NPs as a pigment are probably carcinogenic to human (group 2B) according to mechanisms and animal experiments (particularly exposure by inhalation) [60, 290-291]. Based on the animal experiment data, the safety margin of TiO<sub>2</sub> NPs was 2.25 mg TiO<sub>2</sub> NPs/kg bw/day [292]. In 2016, the European Food Safety Authority (EFSA) reported that exposures of TiO<sub>2</sub> NPs (E171) to humans did not emerge concerns [292]; however, the admissible daily intake could not be determined owing to the insufficient research data.

Studies have examined oral intake of TiO<sub>2</sub> NPs and indicate tissue accumulations and toxic effects [156, 293-294]. The accumulation and toxic effect of TiO<sub>2</sub> NPs in tissue are still debatable [295]. Exposures of TiO<sub>2</sub> NPs at higher age resulted in a significant accumulation [296]; but, studies reported that a child is potentially to have higher accumulation in tissue. The estimation of TiO<sub>2</sub> consumption in a child is 2-4 times more than an adult person (TiO<sub>2</sub> per kg of body weight (bw) per day) because of

the higher child prefers to sweet foods (e.g. chocolate, chewing gum, and candy) containing approximately 2.5 mg TiO<sub>2</sub>/g of food [4, 297-298]. Besides, disputes associated with TiO<sub>2</sub> NPs accumulation in tissue are caused by the fact that their increase in organ levels was not always detected [44] due to very low absorptions and dose-dependent. Although there were low absorptions, they could be approved by visual detection in organs [49].

The absorption of TiO<sub>2</sub> NPs is considered at low levels, but it may be toxic enough in cells. Experiments conducted by Degabriel (2015) exhibited that bare TiO<sub>2</sub> NPs generated more free radicals (hydrogen peroxide (H<sub>2</sub>O<sub>2</sub>)) in phosphate-buffered saline (PBS) than in pure water [45]. As PBS is well known as a cell buffer, it indicates that bare TiO<sub>2</sub> NPs in the cell is toxic. However, studies have shown that TiO<sub>2</sub> NPs contained in digested food models (1.5 % w/w) did not perturb cellular proteome significantly because of their interaction with food matrix [46] and mucus [47]. Mucus is a hydrogel, consisting of water, mucin (glycoproteins), DNA, proteins, lipids, and cell debris [243]. While mucin is densely O-glycosylated proteins which coat a wide variety of wet epithelial cell [38], exclude foreign or hazardous molecules, and permit the useful ones such as nutrients [36, 48]. It can be noticed that the interaction between TiO<sub>2</sub> NPs and mucin may play a significant role in cell toxicity.

The interaction between TiO<sub>2</sub> NPs and mucin occurs in the human gastrointestinal (GI) tract at various pH; hence, the physicochemical property and toxicity of TiO<sub>2</sub> NPs may be altered due to pH. Recent studies have shown that TiO<sub>2</sub> NPs were poorly dissolved in the model gastric and intestinal environment [44]. TiO<sub>2</sub> NPs exposed with juices mimicking the gastric and intestinal compartment (pH=2 & 7) tend to alter their

size and surface charge due to agglomeration and protein adsorptions on their surface [50-51]. But, limited studies have examined the basic interaction between TiO<sub>2</sub> NPs and mucin at different pH and its effect on physicochemical alterations. Therefore, this study observed the basic interaction between TiO<sub>2</sub> NPs and mucin using deionized water (DIW) as a buffer. It can be hypothesized that pH may alter TiO<sub>2</sub> NPs-mucin interactions, and may change the physicochemical properties of TiO<sub>2</sub> NPs (particularly the surface charge), and thereby alter their bio-fate.

Subsequently, BMC which was formed by TiO<sub>2</sub> NPs-mucin interaction in the GI tract may penetrate the cell membrane [52], then can be exposed by a cell buffer. Studies associated with the penetration of cell barrier by TiO<sub>2</sub> NPs have been conducted; but, the condition was not very representative of the real gut environment, particularly they did not consider the presence of mucin [47]. *In vitro* studies on Caco-2 intestinal cells [53-56] and TR146 buccal cells [57] have shown that the translocation of NPs through the epithelial barrier occurs and they stay in the cell. However, no studies have reported the interaction between TiO<sub>2</sub> NPs and mucin in the cell. Therefore, the main purpose of this study is to know the impact of cell buffer (PBS) on TiO<sub>2</sub> NPs-mucin interaction, focusing on the change of surface charge at different pH. It was hypothesized that PBS and pH may affect the alteration of surface charge, the formation of BMC and aggregates.

## **Materials and methods**

### **Materials**

TiO<sub>2</sub> NPs (anatase; 5-15 nm) were purchased from US Research Nanomaterials (TX, US) and mucin was purchased from Sigma-Aldrich Chemicals (St. Louis, MO, USA). The following products were purchased from Life Technologies (Thermo Fisher Scientific, Agawam, MA, USA): hydrochloric acid, sodium hydroxide, sodium chloride, calcium chloride, dimethyl sulfoxide. Silver (Ag) dendrites were prepared based on (He, Lin, Li, & Kim, 2010). This study used deionized water (DIW) and PBS as a buffer.

### **Methods**

Initially, physicochemical properties of TiO<sub>2</sub> NPs and mucus in the function of pH were monitored—particularly by observing the alteration of appearance and turbidity (aggregation/size) and by investigating the change of zeta potential (surface charge). These alterations may lead to a specific interaction (e.g. electrostatic interaction), cause the formation of biomolecular corona (BMC) and aggregation, and thereby induce bio-fate alterations of TiO<sub>2</sub> NPs. These phenomena were measured through: *first*, observing the change of BMC appearance due to the interaction between TiO<sub>2</sub> NPs and mucin at different pH, and examining the surface charge of BMC in DIW a dynamic light scattering (DLS); *second*, comparing the appearance of TiO<sub>2</sub> NPs, mucin, and BMC in different buffers (DIW and PBS) and pH, then examining their zeta potential ( $\zeta$ ) using DLS; *third*, measuring the change of turbidity (O.D. 600); *fourth*, characterizing TiO<sub>2</sub> NPs-mucin interactions to examine the development of hetero aggregation by employing

surface-enhanced Raman spectroscopy (SERS). To know the effect of pH to the basic interaction between TiO<sub>2</sub> NPs and mucin, experiments used DIW as a buffer. Whereas, PBS was used when experiments observed TiO<sub>2</sub> NPs-mucin interactions in the cell.

### **Alterations of BMC Appearance & the Basic Experiment (BE)**

0.15 % (v/v) of TiO<sub>2</sub> NPs were dispersed in DIW & PBS. 5 mg/ml mucin was stirred into DIW & PBS overnight and then centrifuged. Before experiments were conducted, the pH of TiO<sub>2</sub> NPs and mucin should have been adjusted. The change of BMC appearance was observed every time a certain amount of TiO<sub>2</sub> NPs was added. The experiment was started by observing the appearance of control (mucin) by adding 0.5 mL mucin (5mg/mL) into a glass reaction tube and add 4.5 mL DIW. While the treated groups, the amount of TiO<sub>2</sub> NPs (0.1%) was increased 0.25 mL and the amount of DIW was reduced simultaneously; therefore, the total volume of the sample was 5 mL. Vortex was applied for each tube (10 s), and eventually, the alteration of appearance was observed, particularly to identify the possibility of aggregation during TiO<sub>2</sub> NPs-mucin reaction. This experiment was the primary experiment (BE) and should be understood before doing other experiments.

### **Surface charge Characterization**

The surface charge was measured by Zetasizer nano ZS series, Malvern Instruments Ltd, Worcesterhire, UK.

### **UV-Vis Spectrophotometry**

TiO<sub>2</sub> NPs and mucin were mixed and prepared as the same to BE. The absorption spectrum was determined using a SpectraMax M2 microplate reader (Molecular Devices, Sunnyvale, CA, U.S.A.).

### **Surfaced-Enhanced Raman Spectroscopy (SERS)**

Mucin and TiO<sub>2</sub> NPs were dispersed and stirred in buffer overnight. Three different categories of solution (mucin, TiO<sub>2</sub> NPs, and the mixture of TiO<sub>2</sub> NPs and mucin with the mass ratio= 0.25) were prepared. These solutions were homogenized for 10 seconds and added Ag dendrites. These mixtures were homogenized for 10 seconds and sedimented. These sediments were centrifuged and rinsed with buffer three times. Each sediment (2  $\mu$ L) was deposited on a glass slide covered with aluminum foil as a sample. The sample was air-dried before observing it under the Raman laser. A DXR Raman Microscope (Thermo Scientific, Madison, WI) equipped with a 785 nm-excitation laser and a 50x objective was employed. Spectra were collected with a 5.0 mW laser power and a 50  $\mu$ m slit aperture for 2 seconds scanning time. All SERS experiments were done minimally twice. There were 7-9 spots per sample were characterized by SERS. The elicited spectra were analyzed using TQ analyst software, version 8.0 (Thermo Fisher Scientific).

## Statistical Analysis

All experiments were conducted at least twice and reported as the results of means and deviations.

## Results and Discussions

### **Basic Interaction between TiO<sub>2</sub> NPs & Mucin in Various pH**

Studies to examine the adsorption of protein on the surface of NPs in the (GI) tract have been conducted [44, 51] in the liquid mimicking the gastric and intestinal juice (pepsin in HCl, pH=1.2 & 1.5 for gastric environment; and trypsin, KH<sub>2</sub>PO<sub>4</sub> in NaOH, pH= 6.8 & 7.4 for the intestinal environment). However, the basic interaction between TiO<sub>2</sub> NPs and mucin which forms BMC is not well known. Therefore, the aim of this experiment was to know the basic interaction between TiO<sub>2</sub> NPs and mucin using DIW as a buffer, particularly the alteration of the surface charge in various pH.

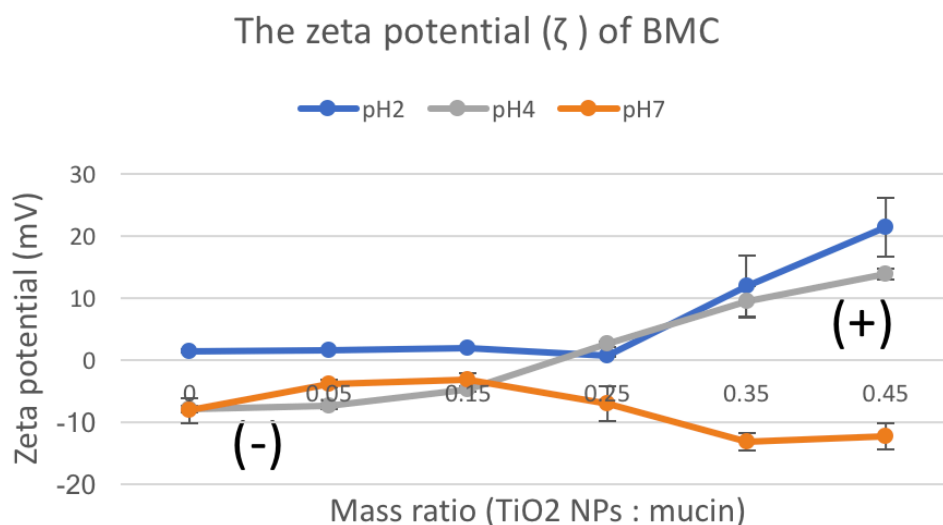


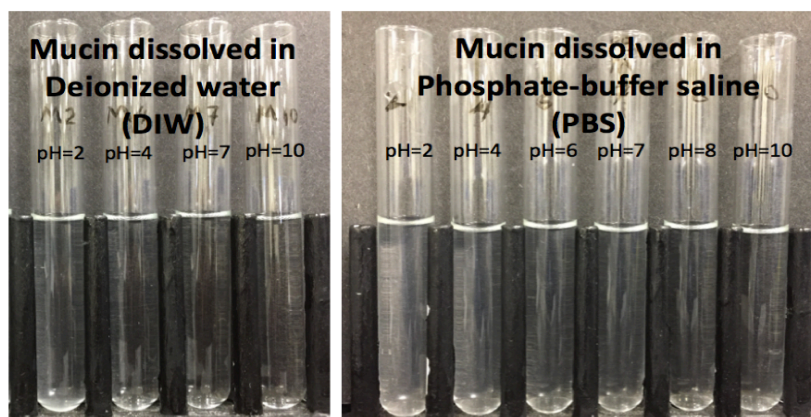
Figure 24: The surface charge of BMC in various pH.

Results showed that the surface charge of BMC was various depending on pH. In acidic conditions (pH=2 & 4) the charge was positive when the mass ratio (TiO<sub>2</sub> NPs: mucin) reached 0.25. Meanwhile, in the neutral and alkaline conditions, the surface charge of BMC was always negative (**Fig. 24**). In theory, cationic NPs features rapid uptake by cells due to their capability to have spontaneous translocation [52]. So, it can be noticed that cationic BMC may have the same capability—the direct translocation crossing the membrane cell. This study is very basic. Therefore, studies to examine TiO<sub>2</sub> NPs-mucin interactions to measure the direct translocation of cationic BMC on membrane cells, using liquids mimicking the GI tract environment, are required.

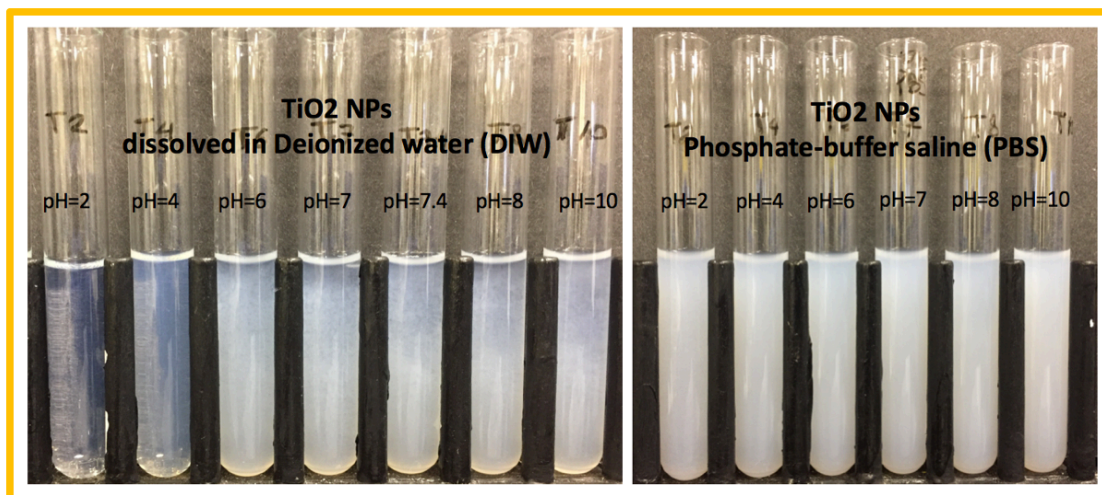
#### **TiO<sub>2</sub> NPs & mucin in different buffer and pH**

Buffers have a significant effect on the appearance of both TiO<sub>2</sub> NPs and mucin. Mucin seemed very stable, the appearance was always clear, in different pH and buffers; while TiO<sub>2</sub> NPs exhibited, various appearances depending on pH and buffers. In DIW, the appearance of TiO<sub>2</sub> NPs had a gradation, the increase of pH resulted in the opaquer the appearance. Meanwhile, the appearance of TiO<sub>2</sub> NPs was almost the same in PBS—they were all opaque.





25 (A)



25 (B)

Figure 25: Appearance of mucin (A) and TiO<sub>2</sub> NPs (B) in deionized water and PBS at different pH. Mucin is stable at different buffers and pH. There is a gradient of appearance when TiO<sub>2</sub> NPs are dispersed in DIW. They tend to be opaquer with the increase of pH; while the appearance of TiO<sub>2</sub> NPs in PBS is all the same (opaque).

Based on **Fig. 25**, it was interesting to characterizing the surface charge of TiO<sub>2</sub> NPs and mucin in DIW and PBS in various pH; therefore, the electrostatic interaction could be determined. If TiO<sub>2</sub> NPs and mucin have a different surface charge, the electrostatic attraction may occur; however, if both of them have the same charge, the electrostatic repulsion may happen.

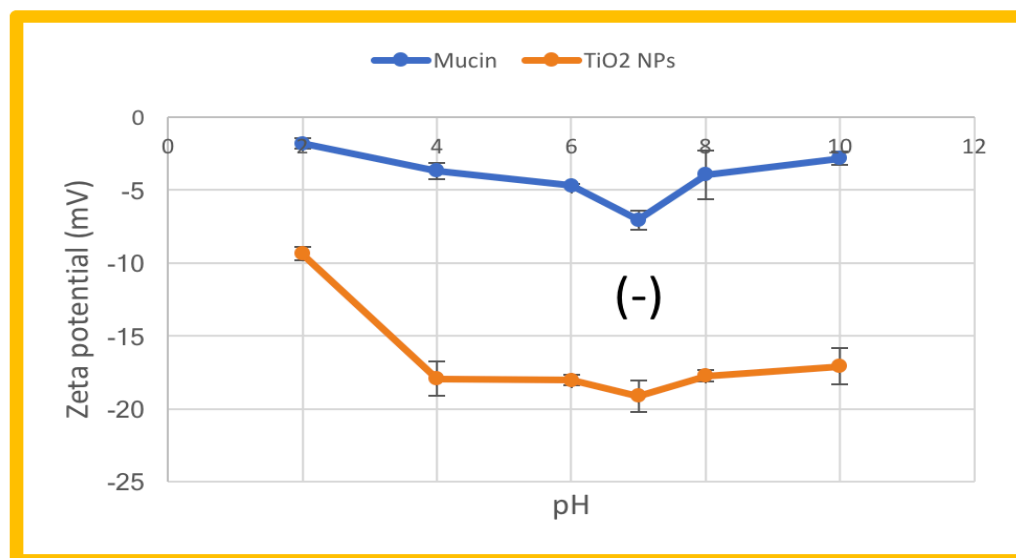
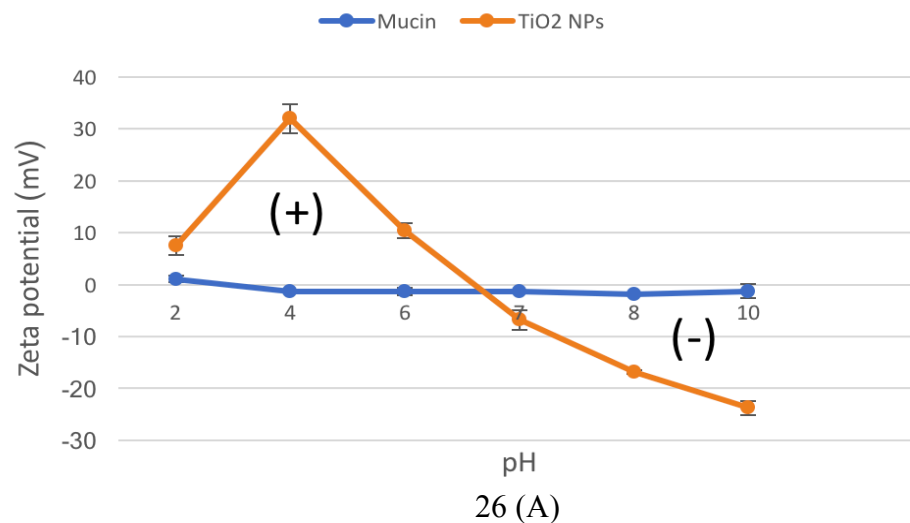


Figure 26: Surface charge of mucin and TiO<sub>2</sub> NPs in DIW (A) and PBS (B).

The surface charge of mucin both in DIW and PBS were stable and slightly negative; while the surface charge of TiO<sub>2</sub> NPs was, variable influenced by buffer and pH. In DIW and the acidic condition, TiO<sub>2</sub> NPs were positive charge; while in the neutral to basic condition, TiO<sub>2</sub> NPs were negative charge. In PBS, albeit the surface charge of

TiO<sub>2</sub> NPs was always negative (**Fig. 26 (B)**), the precipitation of TiO<sub>2</sub> NPs occurred (**Fig. 25 (B)**). It might be due to hydrophobic interaction or other interactions; thereby, studies to know the types of TiO<sub>2</sub> NPs-TiO<sub>2</sub> NPs interactions are needed.

### The formation of Biomolecular Corona BMC

In PBS, all BMC exhibited negative charge. At pH= 7, the surface charge was higher than at pH= 2 and 4 (**Fig. 27**). Whereas, the surface charge of BMC in DIW buffer can be positive and negative depending on the mass ratio (TiO<sub>2</sub> NPs: mucin) and pH (**Fig. 24**). The higher amount of TiO<sub>2</sub> NPs led to becoming more positive charge of BMC. There was a notable fact, that the charge of all BMC was relatively neutral at mass ratio (TiO<sub>2</sub> NPs: mucin) = 0.25, then BMC was becoming either more positive or negative along with the increase of mass ratio (TiO<sub>2</sub> NPs: mucin), depending on the buffer. Thereby, it was concluded that the mass ratio 0.25 is the critical point in terms of defining the charge of BMC.

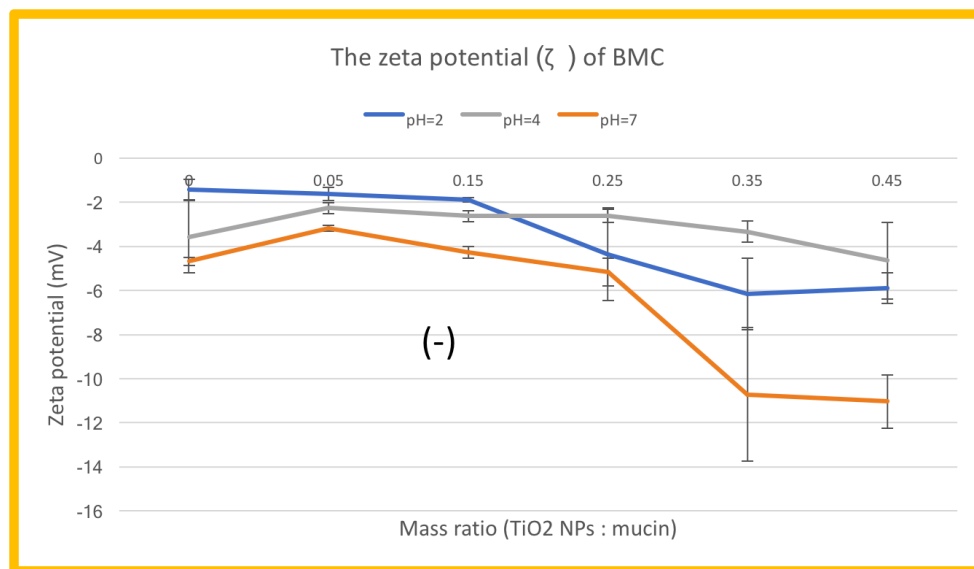
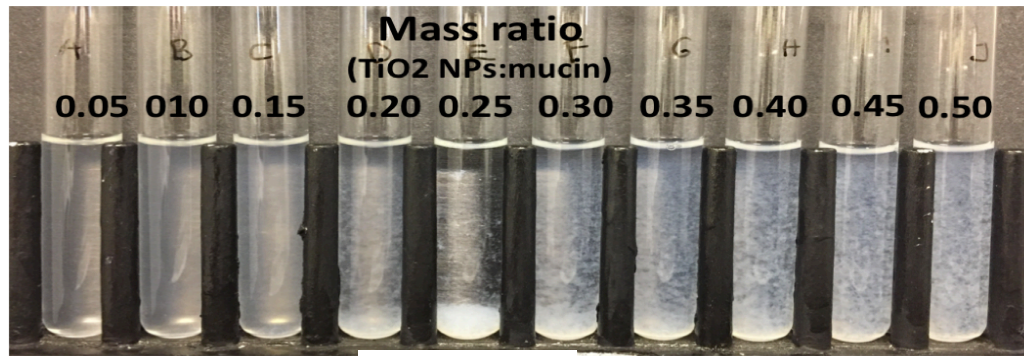
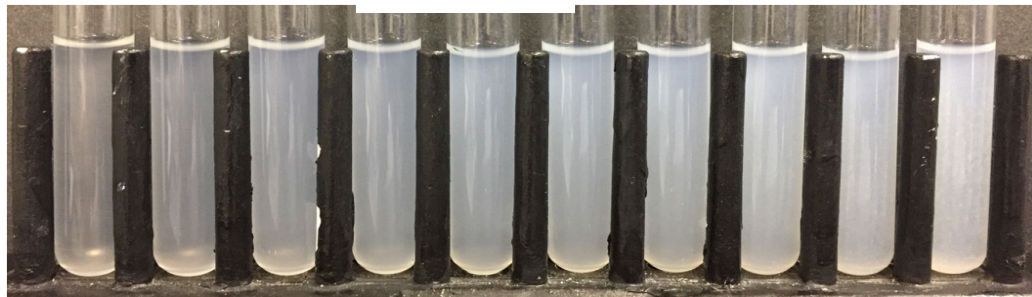


Figure 27: Zeta potentials of BMC in PBS at pH= 2, 4, and 7.

There was another interesting finding correlated with the appearance of BMC in DIW and PBS. Using the same concentration of  $\text{TiO}_2$  NPs (low concentration), **Fig. 28** showed that there was no precipitation appeared in PBS, while the massive precipitation occurred in DIW. No precipitation in PBS was likely caused by either the stable homo aggregation or the stable hetero aggregation [34]. However, once the concentration of  $\text{TiO}_2$  NPs was increased significantly (approximately three times than the initial dose), the aggregation in PBS was depicted (**Fig. 28**). That was also confirmed by the turbidity data. The more amount of  $\text{TiO}_2$  NPs added in the experiment generated the higher turbidity (**Fig. 29 B**).

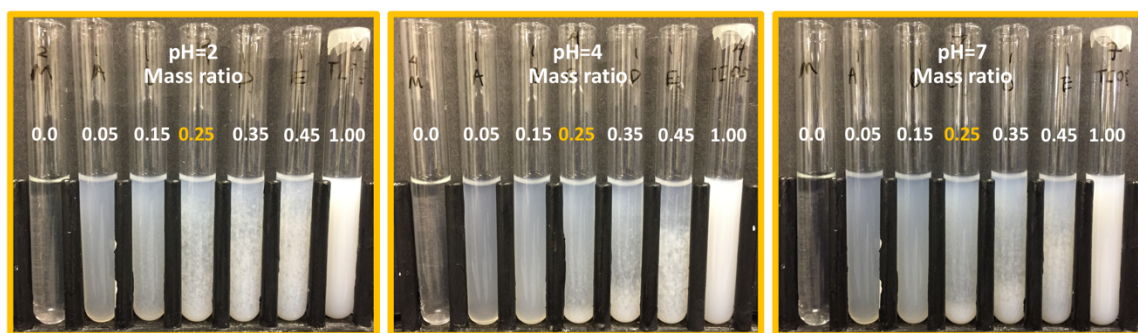


28 (A)

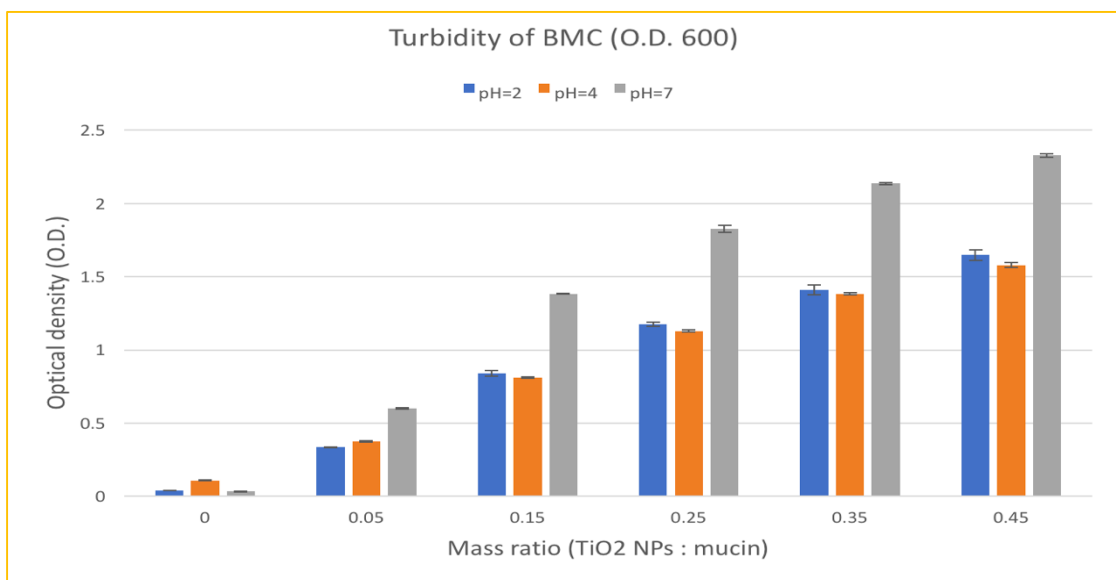


28 (B)

Figure 28: Interaction between  $\text{TiO}_2$  NPs and mucin in DIW (A) and PBS (B).



29 (A)



29 (B)

Figure 29: (A) The formation of BMC at different mass ratio (TiO<sub>2</sub> NPs: mucin) and at pH=2, 4, and 7. (B) Turbidity (O.D. 600). Turbidity at pH=7 is significantly different from turbidity at pH=2 & 4.

Albeit the accumulation of TiO<sub>2</sub> NPs is not always detected in organs due to low absorptions and concentration-dependent [44], it can be visualized [49]. It could be caused by the insufficient concentration of TiO<sub>2</sub> NPs; therefore, unstable hetero aggregation could not be developed. It can be extrapolated that PBS buffer may require a higher concentration of TiO<sub>2</sub> NPs to develop unstable aggregate compared than DIW.



This experiment was still preliminary; further studies, particularly to know the interaction between PBS, TiO<sub>2</sub> NPs and mucin, are required.

### **SERS and TiO<sub>2</sub> NPs-Mucin Interactions: Homo/Hetero Aggregation**

The interaction between TiO<sub>2</sub> NPs and mucin led to developing the unstable hetero aggregation (**Fig. 28 (A) & 29 (A)**). SERS was employed to determine the hetero aggregation, by characterizing the TiO<sub>2</sub> NPs-mucin interaction. This experiment had two steps: first, to know the impact of Ag dendrites on the spectral peaks of TiO<sub>2</sub> NPs and mucin; second, to characterize the interaction between TiO<sub>2</sub> NPs and mucin.

### **Silver (Ag) Dendrites and the Spectrum Enhancement of TiO<sub>2</sub> NPs & Mucin**

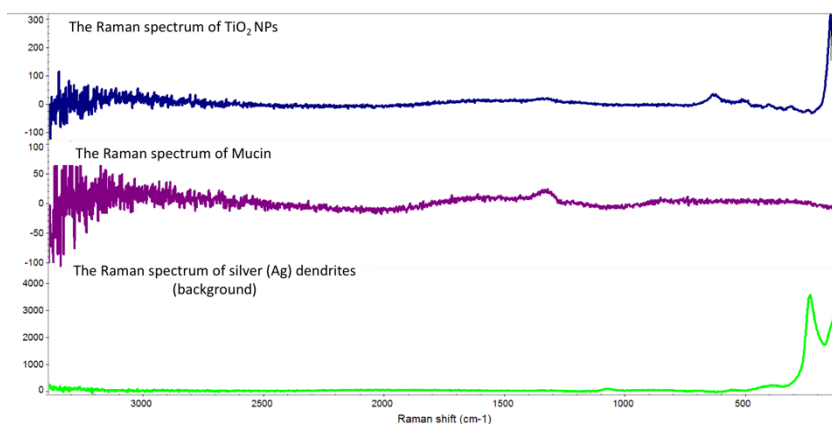
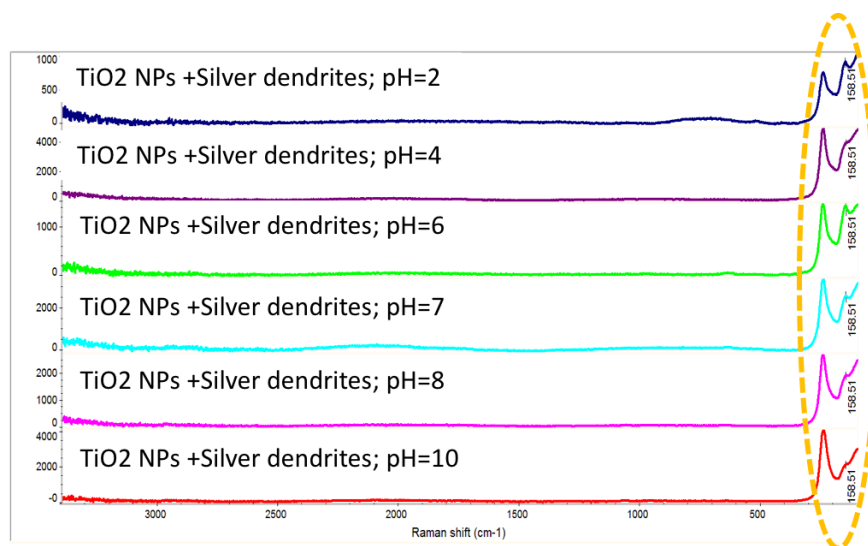


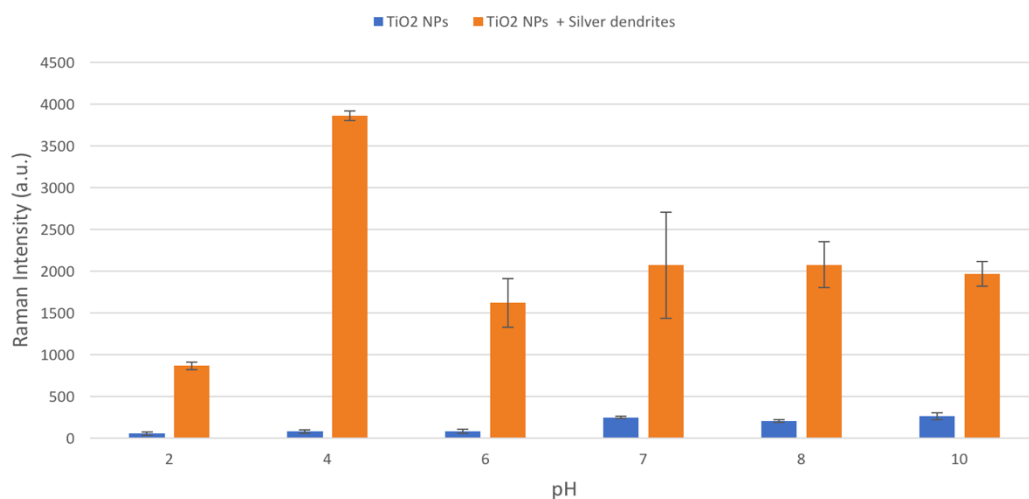
Figure 30: Spectral peaks of TiO<sub>2</sub> NPs, mucin and Ag dendrites (background).

Knowing the original spectral peak of chemicals is pivotal. TiO<sub>2</sub> NPs exhibited a specific spectral peak at wave number 158.51 cm<sup>-1</sup>; while mucin did not exhibit a significant spectral peak. Ag dendrites, as background, represented a spectral peak at wave number 243.63 cm<sup>-1</sup> (**Fig. 30**). Ag dendrites are a good substrate [257] to enhance the spectral peak of TiO<sub>2</sub> NPs and mucin. There was an enormous spectral enhancement

when TiO<sub>2</sub> NPs interacted with Ag dendrites. TiO<sub>2</sub> NPs at pH=4 generated the highest spectral peak (>3700 a.u.), while at pH= 6 to 10 showed spectral peaks from >1300 to 1700 a.u., and at pH= 2 resulted in ±800 - 900 a.u. (**Fig. 31 (A) & (B)**).



31 (A)



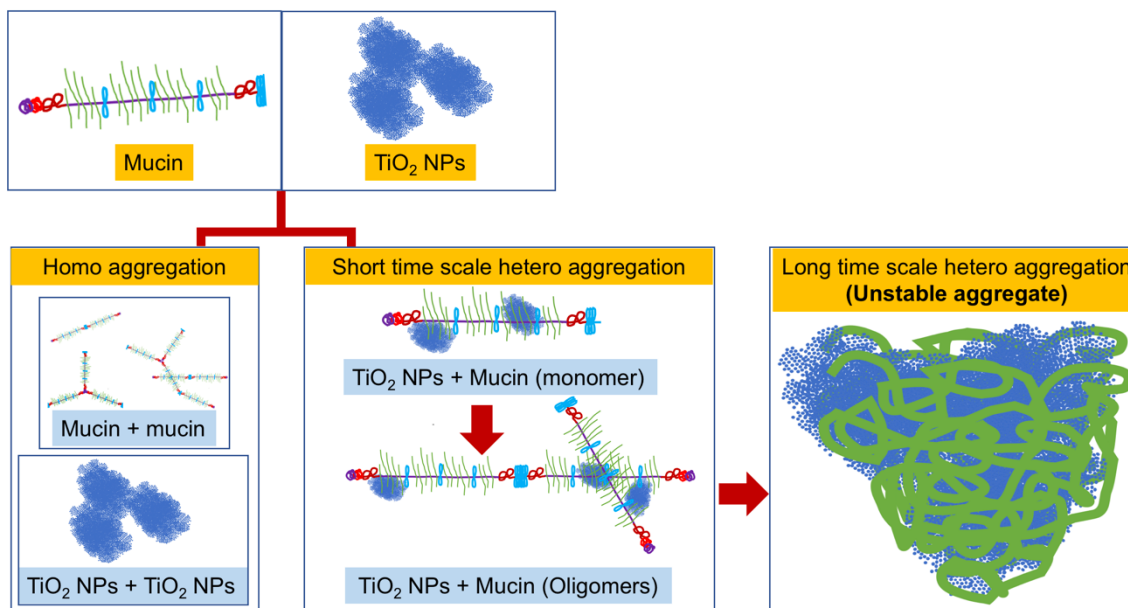
31 (B)

Figure 31: Spectral peaks of TiO<sub>2</sub> NPs and the use of Ag dendrites as a substrate (A). The addition of Ag dendrites tremendously alleviates the spectral peak of TiO<sub>2</sub> NPs (B).

## Characterizations of TiO<sub>2</sub> NPs-Mucin Interactions

Characterizing the interaction between TiO<sub>2</sub> NPs and mucin was important for determining the type of aggregations. **Fig. 29 (A)** showed that aggregates were shuttled down, but the type of aggregation was not unclear either unstable homo aggregation or hetero aggregations. Clavier, Praetorius, & Stoll, (2019) described that mono aggregation is interactions of some similar molecules/particles such as dimer and trimer. Whereas, hetero aggregations are divided into two different categories: stable hetero aggregation (which is formed by dimer/trimer aggregate) and unstable hetero aggregation (which is developed by large and complex aggregate) [34]. Based on the theory and **Fig. 29**, it can be concluded that TiO<sub>2</sub> NPs-mucin interactions are potentially to generate unstable hetero aggregations in PBS (**Fig. 32**).

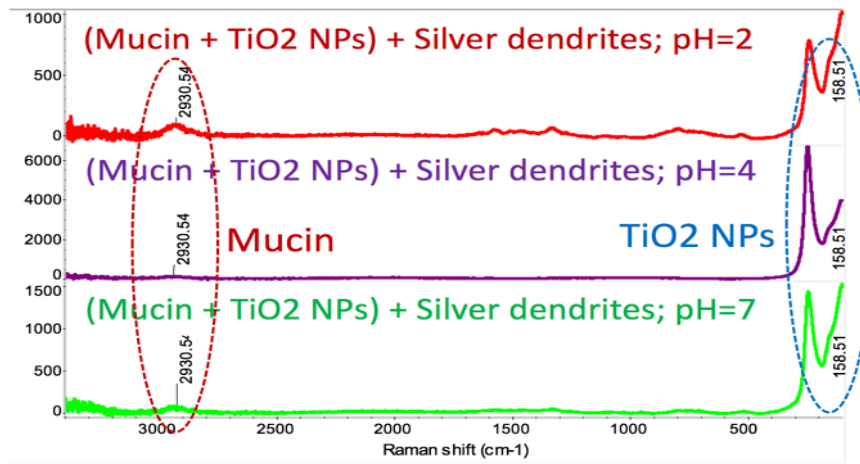
Figure 32: Illustrations of hetero aggregation between TiO<sub>2</sub> NPs and mucin in PBS,



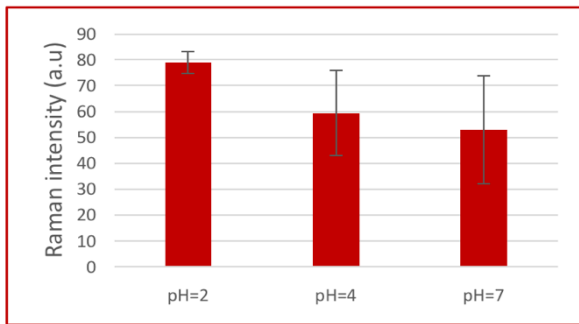
which is inspired by Clavier, Praetorius, & Stoll, (2019).



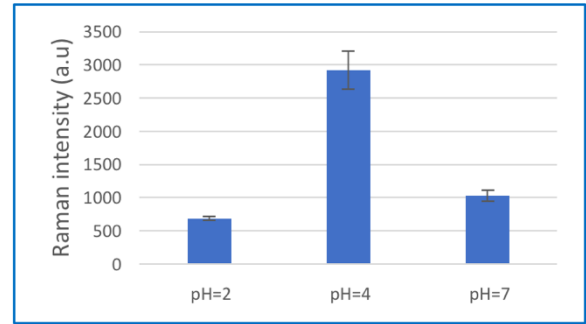
An interaction between TiO<sub>2</sub> NPs and mucin was found in various pH. The spectral peaks of TiO<sub>2</sub> NPs and mucin were appeared (**Fig. 33 (A)**). The spectral peak of TiO<sub>2</sub> NPs was 158.51 cm<sup>-1</sup> and the spectral mucin was at wavenumber at ±2930.54 cm<sup>-1</sup>. The use of mucin spectral peaks at ±2930.54 cm<sup>-1</sup> have the same reason to previous experiment in DIW. The spectral peaks of mucin at various pH were relatively low, stable, and no significant difference (**Fig. 33 (B)**). However, pH influenced the spectral peak of TiO<sub>2</sub> NPs significantly. TiO<sub>2</sub> NPs had the highest spectral peak at pH=4 (**Fig. 31**). For the same reason, the principal component (PC) score of BMC at pH= 4 was the highest than pH=2 and 7 (**Fig. 33 (C)**).



33 (A)



33 (B)



33 (C)

Figure 33: Spectral peaks of BMC (TiO<sub>2</sub> NPs-mucin interactions) at pH=2, 4 and 7 (A); Spectral peaks of mucin (B) and TiO<sub>2</sub> NPs (C).

The principal component analysis (PCA) showed that clusters of BMC at pH=2, 4, and 7 were separated obviously (**Fig. 36**). These data may indicate that the change of pH generates a different species of BMC with new bio-fate; however, more studies are needed to prove that.

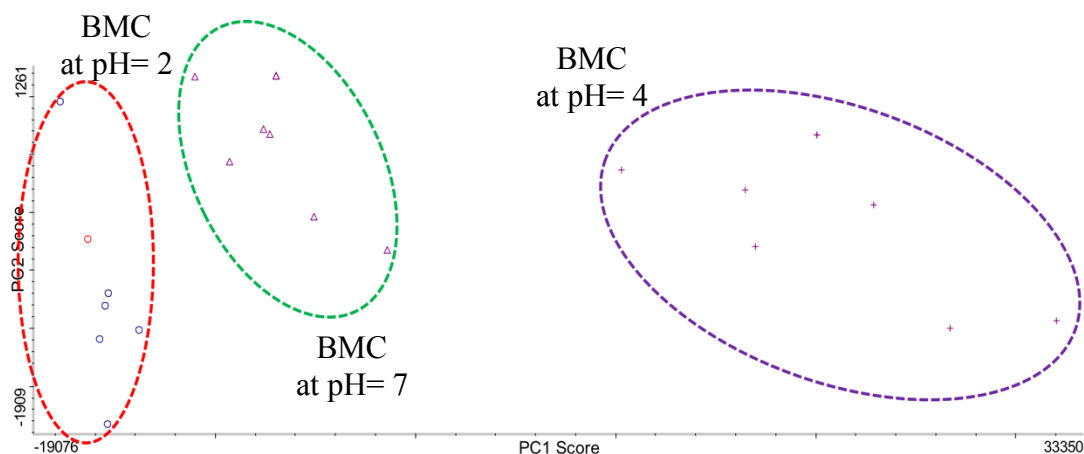


Figure 34: The principal component analysis (PCA) shows that clusters of BMC at pH= 2, 4, and 7 are separated.

### Conclusions

In summary,  $\text{TiO}_2$  NPs-mucin interactions can induce the developmet of BMC and hetero aggregation; however, the accumulation of  $\text{TiO}_2$  NPs in cells is not always detected. PBS and pH play a pivotal role in the surface charge alteration, the formation of BMC, and hetero aggregation. Using a low concentration of  $\text{TiO}_2$  NPs provides a different result in the formation of BMC in DIW and PBS. BMC formations in DIW tend to generate unstable aggregate due to large hetero aggregation, especially when the mass ratio ( $\text{TiO}_2$  NPs: mucin) has reached 0.25; while in PBS, unstable aggregates are not formed. In other words, unstable aggregates are not developed in PBS when the

concentration of  $\text{TiO}_2$  NPs is insufficient. The PCA of SERS spectra show that a different pH is possibly to develop a different species of BMC. It may be concluded that the formation of hetero aggregation in PBS is dose-dependent. The difficulty to detect the  $\text{TiO}_2$  NPs accumulation in the cell may be caused by an insufficient concentration of  $\text{TiO}_2$  NP. Studies to know the type of interactions, and the effect of different species BMC in cells are required.

## CHAPTER V

### 5. CONCLUDING REMARKS

TiO<sub>2</sub> NPs are one of the most inorganic nanoparticles used worldwide. Estimation of the global production of TiO<sub>2</sub> NPs increases continuously until 2020. The potential adverse effect of TiO<sub>2</sub> NPs has raised ecological and health concerns. However, the fate, transformation, transport, and toxicity of TiO<sub>2</sub> NPs in a complex environment are not well known. Therefore, understanding the mechanistic interaction of TiO<sub>2</sub> NPs with mucin—well known as a host barrier from biological, physical, and chemical perils [36]—is necessary. The physicochemical properties (such as surface charge and size) may play a significant role in the TiO<sub>2</sub> NPs-mucin interaction.

The interaction between mucin and TiO<sub>2</sub> NPs occurred in both DIW and PBS. In DIW, the results exhibited that surface charges of TiO<sub>2</sub> NPs were changed from positive to negative with the increase of pH, while surface charges of mucin were not pH-dependent. The development of the strongest cationic TiO<sub>2</sub> NPs was at pH= 4. Meanwhile, pH values near the zero point of charge (pH<sub>pzc</sub>) of TiO<sub>2</sub> NPs were between 6.0 to 7.4. Hence, TiO<sub>2</sub> NPs at pH= 4 was stable due to electrostatic repulsion, while TiO<sub>2</sub> NPs at pH<sub>pzc</sub> between 6.0 to 7.4 was supposed to aggregate. At pH=4, surface charges of BMC were changed from negative to positive along with the increase of TiO<sub>2</sub> NPs concentration. The interaction between TiO<sub>2</sub> NPs and mucin induced the formation of BMC, massive aggregations at mass ratio (TiO<sub>2</sub> NPs: mucin) = 0.25, and an increase in BMC size.

The interaction between TiO<sub>2</sub> NPs and mucin might develop polyvalent binding and result in exothermic energy at pH=4. Despite no electrostatic interaction, hydrophobic interaction, and other interactions may play a crucial role in TiO<sub>2</sub> NPs-mucin interactions at pH= 7.

Cationic BMC was possibly formed at acidic condition (pH= 2 & 4) at mass ratio (TiO<sub>2</sub> NPs: mucin) = 0.25. Based on the result and literature [52], this study proposed that cationic BMC might be able to cross the cell membrane.

In PBS, although surface charges of TiO<sub>2</sub> NPs were all negative, the appearance of TiO<sub>2</sub> NPs in PBS was opaque due to unstable aggregation that may be induced by the hydrophobic interactions or other interactions.

The development of hetero aggregation of BMC in PBS required a higher concentration of TiO<sub>2</sub> NPs than in DIW. Different pH might generate different species and bio-fate of BMC in PBS.

Knowing the interaction between TiO<sub>2</sub> NPs and mucin is required to understand the fate, transformation, transport, and toxicological effect of TiO<sub>2</sub> NPs in a complex environment. Interactions between TiO<sub>2</sub> NPs and mucin may induce polyvalent binding, formation of BMC, and hetero aggregation. In PBS, the development of unstable hetero aggregate of BMC requires a higher concentration of TiO<sub>2</sub> NPs than in DIW. The cationic BMC may be developed in DIW with acidic conditions; while in PBS, there is no cationic BMC formed. This study presented the primary interaction between TiO<sub>2</sub> NPs and mucin in DIW and PBS. Further studies to determine of the types, products, and effects of TiO<sub>2</sub> NPs-mucin interaction are required, particularly in the complex environment mimicking a real ecological and physiological condition.

## BIBLIOGRAPHY

- [1] U.S. EPA. Nanomaterial Case Studies: Nanoscale Titanium Dioxide In Water Treatment And In Topical Sunscreen (Final). U.S. Environmental Protection Agency, Washington, DC, EPA/600/R-09/057F, 2010.
- [2] Landsiedel, R., Ma-Hock, L., Kroll, A., Hahn, D., Schnekenburger, J., Wiench, K., & Wohlleben, W. (2010). Testing metal-oxide nanomaterials for human safety. *Advanced Materials* (Deerfield Beach, Fla.), 22(24), 2601-2627. doi:10.1002/adma.200902658 [doi]
- [3] Gea, M., Bonetta, S., Iannarelli, L., Giovannozzi, A. M., Maurino, V., Bonetta, S., . . . Schilirò, T. (2019). Shape-engineered titanium dioxide nanoparticles (TiO<sub>2</sub>-NPs): Cytotoxicity and genotoxicity in bronchial epithelial cells. *Food and Chemical Toxicology : An International Journal Published for the British Industrial Biological Research Association*, 127, 89-100. doi:S0278-6915(19)30103-6 [pii]
- [4] Weir, A., Westerhoff, P., Fabricius, L., Hristovski, K., & von Goetz, N. (2012). Titanium dioxide nanoparticles in food and personal care products. *Environmental Science & Technology*, 46(4), 2242-2250. doi:10.1021/es204168d
- [5] Chen, H., Zhao, R., Wang, B., Cai, C., Zheng, L., Wang, H., . . . Feng, W. (2017). The effects of orally administered ag, TiO<sub>2</sub> and SiO<sub>2</sub> nanoparticles on gut microbiota composition and colitis induction in mice. *Nanoimpact*, 8, 80-88. doi:https://doi.org/10.1016/j.impact.2017.07.005
- [6] Khan, M. R., Adam, V., Rizvi, T. F., Zhang, B., Ahamad, F., Joško, I., . . . Mao, C. (2019). Nanoparticle–Plant interactions: Two-way traffic. *Small*, 15(37), 1901794. doi:https://doi.org/10.1002/sml.201901794
- [7] Zhu, Y., Xu, F., Liu, Q., Chen, M., Liu, X., Wang, Y., . . . Zhang, L. (2019). Nanomaterials and plants: Positive effects, toxicity and the remediation of metal and metalloid pollution in soil. *Science of the Total Environment*, 662, 414-421. doi:https://doi.org/10.1016/j.scitotenv.2019.01.234
- [8] Tan, W., Peralta-Videa, J., & Gardea-Torresdey, J. (2018). Interaction of titanium dioxide nanoparticles with soil components and plants: Current knowledge and future research needs – a critical review. *Environmental Science: Nano*, 5(2), 257-278. doi:10.1039/C7EN00985B
- [9] Abdel-Latif, H. M. R., Dawood, M. A. O., Menanteau-Ledouble, S., & El-Matbouli, M. (2020). Environmental transformation of n-TiO<sub>2</sub> in the aquatic systems and their ecotoxicity in bivalve mollusks: A systematic review. *Ecotoxicology and Environmental Safety*, 200, 110776. doi:https://doi.org/10.1016/j.ecoenv.2020.110776
- [10] Zhang, D., Qiu, J., Shi, L., Liu, Y., Pan, B., & Xing, B. (2020). The mechanisms and environmental implications of engineered nanoparticles dispersion. *Science of the Total Environment*, 722, 137781. doi:https://doi.org/10.1016/j.scitotenv.2020.137781

- [11] Sengul, A. B., & Asmatulu, E. (2020). Toxicity of metal and metal oxide nanoparticles: A review. *Environmental Chemistry Letters*, 18(5), 1659-1683. doi:10.1007/s10311-020-01033-6
- [12] Lehutso, R. F., Tancu, Y., Maity, A., & Thwala, M. (2020). Aquatic toxicity of transformed and product-released engineered nanomaterials: An overview of the current state of knowledge. *Process Safety and Environmental Protection*, 138, 39-56. doi:https://doi.org/10.1016/j.psep.2020.03.002
- [13] Judy, J. D., & Bertsch, P. M. (2014). Chapter one - bioavailability, toxicity, and fate of manufactured nanomaterials in terrestrial ecosystems. In D. L. Sparks (Ed.), *Advances in agronomy* (pp. 1-64) Academic Press. doi:https://doi.org/10.1016/B978-0-12-420225-2.00001-7 Retrieved from <http://www.sciencedirect.com/science/article/pii/B9780124202252000017>
- [14] Hasanzadeh Kafshgari, M., & Goldmann, W. H. (2020). Insights into theranostic properties of titanium dioxide for nanomedicine. *Nano-Micro Letters*, 12(1), 22. doi:10.1007/s40820-019-0362-1
- [15] Musial, J., Krakowiak, R., Mlynarczyk, D. T., Goslinski, T., & Stanisz, B. J. (2020). Titanium dioxide nanoparticles in food and personal care products-what do we know about their safety? *Nanomaterials (Basel, Switzerland)*, 10(6), 1110. doi:10.3390/nano10061110
- [16] Danielsson, K., Gallego-Urrea, J., Hasselov, M., Gustafsson, S., & Jonsson, C. M. (2017). Influence of organic molecules on the aggregation of TiO<sub>2</sub> nanoparticles in acidic conditions. *Journal of Nanoparticle Research*, 19(4), 133. doi:10.1007/s11051-017-3807-9
- [17] Ellis, L. A., Valsami-Jones, E., Lead, J. R., & Baalousha, M. (2016). Impact of surface coating and environmental conditions on the fate and transport of silver nanoparticles in the aquatic environment. *The Science of the Total Environment*, 568, 95-106. doi:S0048-9697(16)31136-6 [pii]
- [18] Koelmans, A. A., Diepens, N. J., Velzeboer, I., Besseling, E., Quik, J. T., & van de Meent, D. (2015). Guidance for the prognostic risk assessment of nanomaterials in aquatic ecosystems. *The Science of the Total Environment*, 535, 141-149. doi:S0048-9697(15)00168-0 [pii]
- [19] Nowack, B., & Bucheli, T. D. (2007). Occurrence, behavior and effects of nanoparticles in the environment. *Environmental Pollution (Barking, Essex : 1987)*, 150(1), 5-22. doi:S0269-7491(07)00273-4 [pii]
- [20] Nowack, B., Ranville, J. F., Diamond, S., Gallego-Urrea, J. A., Metcalfe, C., Rose, J., . . . Klaine, S. J. (2012). Potential scenarios for nanomaterial release and subsequent alteration in the environment. *Environmental Toxicology and Chemistry*, 31(1), 50-59. doi:10.1002/etc.726 [doi]
- [21] Keller, A. A., Wang, H., Zhou, D., Lenihan, H. S., Cherr, G., Cardinale, B. J., . . . Ji, Z. (2010). Stability and aggregation of metal oxide nanoparticles in natural aqueous matrices. *Environmental Science & Technology*, 44(6), 1962-1967. doi:10.1021/es902987d

- [22] Davis, J. A., James, R. O., & Leckie, J. O. (1978). Surface ionization and complexation at the oxide/water interface: I. computation of electrical double layer properties in simple electrolytes. *Journal of Colloid and Interface Science*, 63(3), 480-499. doi:[https://doi.org/10.1016/S0021-9797\(78\)80009-5](https://doi.org/10.1016/S0021-9797(78)80009-5)
- [23] Garcia, J., Markovski, J., McKay Gifford, J., Apul, O., & Hristovski, K. D. (2017). The effect of metal (hydr)oxide nano-enabling on intraparticle mass transport of organic contaminants in hybrid granular activated carbon. *The Science of the Total Environment*, 586, 1219-1227. doi:S0048-9697(17)30363-7 [pii]
- [24] Li, Z., Sahle-Demessie, E., Aly Hassan, A., Pressman, J. G., Sorial, G. A., & Han, C. (2017). Effects of source and seasonal variations of natural organic matters on the fate and transport of CeO(2) nanoparticles in the environment. *The Science of the Total Environment*, 609, 1616-1626. doi:S0048-9697(17)31857-0 [pii]
- [25] Praetorius, A., Scheringer, M., & Hungerbühler, K. (2012). Development of environmental fate models for engineered Nanoparticles—A case study of TiO<sub>2</sub> nanoparticles in the rhine river. *Environmental Science & Technology*, 46(12), 6705-6713. doi:10.1021/es204530n
- [26] Cornelis, G., Pang, L., Doolette, C., Kirby, J. K., & McLaughlin, M. J. (2013). Transport of silver nanoparticles in saturated columns of natural soils. *The Science of the Total Environment*, 463-464, 120-130. doi:S0048-9697(13)00642-6 [pii]
- [27] Thio, B. J. R., Zhou, D., & Keller, A. A. (2011). Influence of natural organic matter on the aggregation and deposition of titanium dioxide nanoparticles. *Journal of Hazardous Materials*, 189(1-2), 556-563. doi:10.1016/j.jhazmat.2011.02.072
- [28] Baalousha, M. (2009). Aggregation and disaggregation of iron oxide nanoparticles: Influence of particle concentration, pH and natural organic matter. *The Science of the Total Environment*, 407(6), 2093-2101. doi:10.1016/j.scitotenv.2008.11.022 [doi]
- [29] Carnal, F., Clavier, A., & Stoll, S. (2015). Modelling the interaction processes between nanoparticles and biomacromolecules of variable hydrophobicity: Monte carlo simulations. *Environmental Science: Nano*, 2(4), 327-339. doi:10.1039/C5EN00054H
- [30] Braun, A., Klumpp, E., Azzam, R., & Neukum, C. (2015). Transport and deposition of stabilized engineered silver nanoparticles in water saturated loamy sand and silty loam. *The Science of the Total Environment*, 535, 102-112. doi:S0048-9697(14)01725-2 [pii]
- [31] Auffan, M., Rose, J., Bottero, J. Y., Lowry, G. V., Jolivet, J. P., & Wiesner, M. R. (2009). Towards a definition of inorganic nanoparticles from an environmental, health and safety perspective. *Nature Nanotechnology*, 4(10), 634-641. doi:10.1038/nnano.2009.242 [doi]



- [32] Labib, M. E. (1988). The origin of the surface charge on particles suspended in organic liquids. *Colloids and Surfaces*, 29(3), 293-304. doi:[https://doi.org/10.1016/0166-6622\(88\)80124-0](https://doi.org/10.1016/0166-6622(88)80124-0)
- [33] USGS. (2019). Water, the Universal Solvent. November 21, 2020. [https://www.usgs.gov/special-topic/water-science-school/science/water-universal-solvent?qt-science\\_center\\_objects=0#qt-science\\_center\\_objects](https://www.usgs.gov/special-topic/water-science-school/science/water-universal-solvent?qt-science_center_objects=0#qt-science_center_objects)
- [34] Clavier, A., Praetorius, A., & Stoll, S. (2019). Determination of nanoparticle heteroaggregation attachment efficiencies and rates in presence of natural organic matter monomers. monte carlo modelling. *Science of the Total Environment*, 650, 530-540. doi:<https://doi.org/10.1016/j.scitotenv.2018.09.017>
- [35] Seijo, M., Ulrich, S., Filella, M., Buffle, J., & Stoll, S. (2009). Modeling the adsorption and coagulation of fulvic acids on colloids by brownian dynamics simulations. *Environmental Science & Technology*, 43(19), 7265-7269. doi:10.1021/es9002394
- [36] Bansil, R., & Turner, B. S. (2006). Mucin structure, aggregation, physiological functions and biomedical applications. *Current Opinion in Colloid & Interface Science*, 11(2), 164-170. doi:<https://doi.org/10.1016/j.cocis.2005.11.001>
- [37] Witten, J., Samad, T., & Ribbeck, K. (2018). Selective permeability of mucus barriers. *Current Opinion in Biotechnology*, 52, 124-133. doi:S0958-1669(17)30232-X [pii]
- [38] Wagner, C. E., Wheeler, K. M., & Ribbeck, K. (2018). Mucins and their role in shaping the functions of mucus barriers. *Annual Review of Cell and Developmental Biology*, 34(1), 189-215. doi:10.1146/annurev-cellbio-100617-062818
- [39] Reverter, M., Tapissier-Bontemps, N., Lecchini, D., Banaigs, B., & Sasal, P. (2018). Biological and ecological roles of external fish mucus: A review. *Fishes*, 3(4) doi:<http://dx.doi.org/10.3390/fishes3040041>
- [40] Hajirezaee, S., Mohammadi, G., & Naserabad, S. S. (2020). The protective effects of vitamin C on common carp (*cyprinus carpio*) exposed to titanium oxide nanoparticles (TiO<sub>2</sub>-NPs). *Aquaculture*, 518, 734734. doi:<https://doi.org/10.1016/j.aquaculture.2019.734734>
- [41] Davies, M. S., & Hawkins, S. J. (1998). Mucus from marine molluscs. In J. H. S. Blaxter, A. J. Southward & P. A. Tyler (Eds.), *Advances in marine biology* (pp. 1-71) Academic Press. doi:[https://doi.org/10.1016/S0065-2881\(08\)60210-2](https://doi.org/10.1016/S0065-2881(08)60210-2)  
Retrieved from <http://www.sciencedirect.com/science/article/pii/S0065288108602102>
- [42] Yin, C., Zhao, W., Liu, R., Liu, R., Wang, Z., Zhu, L., . . . Liu, S. (2017). TiO<sub>2</sub> particles in seafood and surimi products: Attention should be paid to their exposure and uptake through foods. *Chemosphere*, 188, 541-547. doi:S0045-6535(17)31400-5 [pii]
- [43] McClements, D. J., Xiao, H., & Demokritou, P. (2017). Physicochemical and colloidal aspects of food matrix effects on gastrointestinal fate of ingested

- inorganic nanoparticles. *Advances in Colloid and Interface Science*, 246, 165-180. doi:<https://doi.org/10.1016/j.cis.2017.05.010>
- [44] Cho, W. S., Kang, B. C., Lee, J. K., Jeong, J., Che, J. H., & Seok, S. H. (2013). Comparative absorption, distribution, and excretion of titanium dioxide and zinc oxide nanoparticles after repeated oral administration. *Particle and Fibre Toxicology*, 10, 9-8977-10-9. doi:10.1186/1743-8977-10-9 [doi]
  - [45] Degabriel, T. (2015). Study of the interaction between proteins and TiO<sub>2</sub> NPs : Nature of the interfacial processes
  - [46] Cao, X., Zhang, T., DeLoid, G. M., Gaffrey, M. J., Weitz, K. K., Thrall, B. D., . . . Demokritou, P. (2020). Evaluation of the cytotoxic and cellular proteome impacts of food-grade TiO<sub>2</sub> (E171) using simulated gastrointestinal digestions and a tri-culture small intestinal epithelial model. *Nanoimpact*, 17, 10.1016/j.impact.2019.100202. doi:10.1016/j.impact.2019.100202 [doi]
  - [47] Mercier-Bonin, M., Despax, B., Raynaud, P., Houdeau, E., & Thomas, M. (2018). Mucus and microbiota as emerging players in gut nanotoxicology: The example of dietary silver and titanium dioxide nanoparticles. *Critical reviews in food science and nutrition*, 58(6), 1023–1032. <https://doi.org/10.1080/10408398.2016.1243088>
  - [48] Linden, S. K., Sutton, P., Karlsson, N. G., Korolik, V., & McGuckin, M. A. (2008). Mucins in the mucosal barrier to infection. *Mucosal Immunology*, 1(3), 183-197. doi:10.1038/mi.2008.5
  - [49] Tassinari, R., Cubadda, F., Moracci, G., Aureli, F., D'Amato, M., Valeri, M., . . . Maranghi, F. (2014). Oral, short-term exposure to titanium dioxide nanoparticles in sprague-dawley rat: Focus on reproductive and endocrine systems and spleen. *Nanotoxicology*, 8(6), 654-662. doi:10.3109/17435390.2013.822114 [doi]
  - [50] Brun, E., Barreau, F., Veronesi, G., Fayard, B., Sorieul, S., Chanéac, C., . . . Carrière, M. (2014). Titanium dioxide nanoparticle impact and translocation through ex vivo, in vivo and in vitro gut epithelia. *Particle and Fibre Toxicology*, 11, 13-8977-11-13. doi:10.1186/1743-8977-11-13 [doi]
  - [51] Wang, Y., Chen, Z., Ba, T., Pu, J., Chen, T., Song, Y., . . . Jia, G. (2013). Susceptibility of young and adult rats to the oral toxicity of titanium dioxide nanoparticles. *Small (Weinheim an Der Bergstrasse, Germany)*, 9(9-10), 1742-1752. doi:10.1002/sml.201201185 [doi]
  - [52] Lin, J., & Alexander-Katz, A. (2013). Cell membranes open "doors" for cationic nanoparticles/biomolecules: Insights into uptake kinetics. *ACS Nano*, 7(12), 10799-10808. doi:10.1021/nn4040553 [doi]
  - [53] Böhmert, L., Girod, M., Hansen, U., Maul, R., Knappe, P., Niemann, B., . . . Lampen, A. (2014). Analytically monitored digestion of silver nanoparticles and their toxicity on human intestinal cells. *Nanotoxicology*, 8(6), 631-642. doi:10.3109/17435390.2013.815284 [doi]

- [54] Böhmert, L., Niemann, B., Lichtenstein, D., Juling, S., & Lampen, A. (2015). Molecular mechanism of silver nanoparticles in human intestinal cells. *9*(7), 852-860. doi:10.3109/17435390.2014.980760
- [55] Gerloff, K., Fenoglio, I., Carella, E., Kolling, J., Albrecht, C., Boots, A. W., . . . Schins, R. P. (2012). Distinctive toxicity of TiO<sub>2</sub> rutile/anatase mixed phase nanoparticles on caco-2 cells. *Chemical Research in Toxicology*, *25*(3), 646-655. doi:10.1021/tx200334k [doi]
- [56] Gerloff, K., Pereira, D. I. A., Faria, N., Boots, A. W., Kolling, J., Förster, I., . . . Schins, R. P. F. (2013). Influence of simulated gastrointestinal conditions on particle-induced cytotoxicity and interleukin-8 regulation in differentiated and undifferentiated caco-2 cells. *Nanotoxicology*, *7*(4), 353-366. doi:10.3109/17435390.2012.662249
- [57] Tay, C. Y., Fang, W., Setyawati, M. I., Chia, S. L., Tan, K. S., Hong, C. H., & Leong, D. T. (2014). Nano-hydroxyapatite and nano-titanium dioxide exhibit different subcellular distribution and apoptotic profile in human oral epithelium. *ACS Applied Materials & Interfaces*, *6*(9), 6248-6256. doi:10.1021/am501266a [doi]
- [58] ISO. (2015). Vocabulary: Part 1—core terms, International standardisation organisation (ISO), technical specification ISO/TSU.S.C.
- [59] Jomini, S., Clivot, H., Bauda, P., & Pagnout, C. (2015). Impact of manufactured TiO<sub>2</sub> nanoparticles on planktonic and sessile bacterial communities. *Environmental Pollution (Barking, Essex : 1987)*, *202*, 196-204. doi:S0269-7491(15)00144-X [pii]
- [60] IARC Working Group on the Evaluation of Carcinogenic Risks to Humans. (2010). Carbon black, titanium dioxide, and talc. Lyon (FR): International Agency for Research on Cancer (IARC Monographs on the Evaluation of Carcinogenic Risks to Humans, No. 93.). Retrieved from <https://www.ncbi.nlm.nih.gov/books/NBK326521/>
- [61] Smijs, T. G., & Pavel, S. (2011). Titanium dioxide and zinc oxide nanoparticles in sunscreens: Focus on their safety and effectiveness. *Nanotechnology, Science and Applications*, *4*, 95-112. doi:10.2147/NSA.S19419
- [62] Zhang, J., Thurber, A., Hanna, C., & Punnoose, A. (2010). Highly shape-selective synthesis, silica coating, self-assembly, and magnetic hydrogen sensing of hematite nanoparticles. *Langmuir*, *26*(7), 5273-5278. doi:10.1021/la903544a
- [63] Lundqvist, M., Stigler, J., Elia, G., Lynch, I., Cedervall, T., & Dawson, K. A. (2008). Nanoparticle size and surface properties determine the protein corona with possible implications for biological impacts. *Proc Natl Acad Sci USA*, *105*(38), 14265. doi:10.1073/pnas.0805135105
- [64] Mano, S. S., Kanehira, K., Sonezaki, S., & Taniguchi, A. (2012). Effect of polyethylene glycol modification of TiO<sub>2</sub> nanoparticles on cytotoxicity and gene expressions in human cell lines. *International Journal of Molecular Sciences*, *13*(3), 3703-3717. doi:10.3390/ijms13033703

- [65] Albanese, A., Walkey, C. D., Olsen, J. B., Guo, H., Emili, A., & Chan, W. C. (2014). Secreted biomolecules alter the biological identity and cellular interactions of nanoparticles. *ACS Nano*, 8(6), 5515-5526. doi:10.1021/nm4061012 [doi]
- [66] Jain, P., Pawar, R. S., Pandey, R. S., Madan, J., Pawar, S., Lakshmi, P. K., & Sudheesh, M. S. (2017). In-vitro in-vivo correlation (IVIVC) in nanomedicine: Is protein corona the missing link? *Biotechnology Advances*, 35(7), 889-904. doi:https://doi.org/10.1016/j.biotechadv.2017.08.003
- [67] Domingos, R. F., Tufenkji, N., & Wilkinson, K. J. (2009). Aggregation of titanium dioxide nanoparticles: Role of a fulvic acid. *Environmental Science & Technology*, 43(5), 1282-1286. doi:10.1021/es8023594
- [68] Abdal Dayem, A., Hossain, M. K., Lee, S. B., Kim, K., Saha, S. K., Yang, G. M., . . . Cho, S. G. (2017). The role of reactive oxygen species (ROS) in the biological activities of metallic nanoparticles. *International Journal of Molecular Sciences*, 18(1), 120. doi: 10.3390/ijms18010120. doi:10.3390/ijms18010120 [doi]
- [69] Bourgeault, A., Veronique Legros, Gonnet, F., Daniel, R., Aurelie Paquirissamy, Clemence Benatar, . . . Pin, S. (2017). Interaction of TiO<sub>2</sub> nanoparticles with proteins from aquatic organisms: The case of gill mucus from blue mussel. *Environmental Science and Pollution Research International*, 24(15), 13474-13483. doi:10.1007/s11356-017-8801-3
- [70] Tsai, S., Duran-Robles, E., Goshia, T., Mesina, M., Garcia, C., Young, J., . . . Chin, W. (2018). CeO<sub>2</sub> nanoparticles attenuate airway mucus secretion induced by TiO<sub>2</sub> nanoparticles. *Science of the Total Environment*, 631, 262-269.
- [71] Fadeel, B., Feliu, N., Vogt, C., Abdelmonem, A. M., & Parak, W. J. (2013). Bridge over troubled waters: Understanding the synthetic and biological identities of engineered nanomaterials. *Wiley Interdisciplinary Reviews.Nanomedicine and Nanobiotechnology*, 5(2), 111-129. doi:10.1002/wnan.1206 [doi]
- [72] Vilanova, O., Mittag, J. J., Kelly, P. M., Milani, S., Dawson, K. A., Rädler, J., O., & Franzese, G. (2016). Understanding the kinetics of protein-nanoparticle corona formation. *ACS Nano*, 10(12), 10842-10850. doi:10.1021/acsnano.6b04858
- [73] Cedervall, T., Lynch, I., Foy, M., Berggård, T., Donnelly, S. C., Cagney, G., . . . Dawson, K. A. (2007). Detailed identification of plasma proteins adsorbed on copolymer nanoparticles. *Angewandte Chemie (International Ed.in English)*, 46(30), 5754-5756. doi:10.1002/anie.200700465 [doi]
- [74] Cedervall, T., Lynch, I., Lindman, S., Berggård, T., Thulin, E., Nilsson, H., . . . Linse, S. (2007). Understanding the nanoparticle-protein corona using methods to quantify exchange rates and affinities of proteins for nanoparticles. *Proceedings of the National Academy of Sciences of the United States of America*, 104(7), 2050-2055. doi:0608582104 [pii]
- [75] Park, Y., Bae, H. C., Jang, Y., Jeong, S. H., Lee, H. N., Ryu, W., . . . Son, S. W. (2013). Effect of the size and surface charge of silica nanoparticles on cutaneous

- toxicity. *Molecular & Cellular Toxicology*, 9(1), 67-74. doi:10.1007/s13273-013-0010-7
- [76] Xiong, D., Fang, T., Yu, L., Sima, X., & Zhu, W. (2011). Effects of nano-scale TiO<sub>2</sub>, ZnO and their bulk counterparts on zebrafish: Acute toxicity, oxidative stress and oxidative damage. *The Science of the Total Environment*, 409(8), 1444-1452. doi:10.1016/j.scitotenv.2011.01.015 [doi]
  - [77] Hu, Z., Zhang, H., Zhang, Y., Wu, R., & Zou, H. (2014). Nanoparticle size matters in the formation of plasma protein coronas on Fe<sub>3</sub>O<sub>4</sub> nanoparticles. *Colloids and Surfaces B: Biointerfaces*, 121, 354-361. doi:https://doi.org/10.1016/j.colsurfb.2014.06.016
  - [78] Schäffler, M., Semmler-Behnke, M., Sarioglu, H., Takenaka, S., Wenk, A., Schleh, C., . . . Kreyling, W. G. (2013). Serum protein identification and quantification of the corona of 5, 15 and 80 nm gold nanoparticles. *Nanotechnology*, 24(26), 265103-4484/24/26/265103. doi:10.1088/0957-4484/24/26/265103 [doi]
  - [79] Klein, J. (2007). Probing the interactions of proteins and nanoparticles. *Proceedings of the National Academy of Sciences of the United States of America*, 104(7), 2029-2030. doi:0611610104 [pii]
  - [80] Kurylowicz, M., Paulin, H., Mogyoros, J., Giuliani, M., & Dutcher, J. R. (2014). The effect of nanoscale surface curvature on the oligomerization of surface-bound proteins. *Journal of the Royal Society, Interface*, 11(94), 20130818. doi:10.1098/rsif.2013.0818 [doi]
  - [81] Piella, J., Bastús, N. G., & Puntès, V. (2017). Size-dependent protein-nanoparticle interactions in citrate-stabilized gold nanoparticles: The emergence of the protein corona. *Bioconjugate Chemistry*, 28(1), 88-97. doi:10.1021/acs.bioconjchem.6b00575 [doi]
  - [82] Gessner, A., Lieske, A., Paulke, B., & Müller, R. (2002). Influence of surface charge density on protein adsorption on polymeric nanoparticles: Analysis by two-dimensional electrophoresis. *European Journal of Pharmaceutics and Biopharmaceutics : Official Journal of Arbeitsgemeinschaft Fur Pharmazeutische Verfahrenstechnik e.V.*, 54(2), 165-170. doi:S0939641102000814 [pii]
  - [83] Hühn, J., Fedeli, C., Zhang, Q., Masood, A., del Pino, P., Khashab, N. M., . . . Parak, W. J. (2016). Dissociation coefficients of protein adsorption to nanoparticles as quantitative metrics for description of the protein corona: A comparison of experimental techniques and methodological relevance. *The International Journal of Biochemistry & Cell Biology*, 75, 148-161. doi:https://doi.org/10.1016/j.biocel.2015.12.015
  - [84] Calatayud, M. P., Sanz, B., Raffa, V., Riggio, C., Ibarra, M. R., & Goya, G. F. (2014). The effect of surface charge of functionalized Fe<sub>3</sub>O<sub>4</sub> nanoparticles on protein adsorption and cell uptake. *Biomaterials*, 35(24), 6389-6399. doi:S0142-9612(14)00383-4 [pii]

- [85] Stoehr, L. C., Gonzalez, E., Stampfl, A., Casals, E., Duschl, A., Puentes, V., & Oostingh, G. J. (2011). Shape matters: Effects of silver nanospheres and wires on human alveolar epithelial cells. *Particle and Fibre Toxicology*, 8, 36-36. doi:10.1186/1743-8977-8-36
- [86] Jansch, M., Jindal, A. B., Sharmila, B. M., Samad, A., Devarajan, P. V., & Müller, R. H. (2013). Influence of particle shape on plasma protein adsorption and macrophage uptake. *Die Pharmazie*, 68(1), 27-33.
- [87] Deng, Z. J., Mortimer, G., Schiller, T., Musumeci, A., Martin, D., & Minchin, R. F. (2009). Differential plasma protein binding to metal oxide nanoparticles. *Nanotechnology*, 20(45), 455101-4484/20/45/455101. Epub 2009 Oct 13. doi:10.1088/0957-4484/20/45/455101 [doi]
- [88] Meder, F., Brandes, C., Treccani, L., & Rezwani, K. (2013). Controlling protein-particle adsorption by surface tailoring colloidal alumina particles with sulfonate groups. *Acta Biomaterialia*, 9(3), 5780-5787. doi:https://doi.org/10.1016/j.actbio.2012.11.012
- [89] Mortensen, N. P., Hurst, G. B., Wang, W., Foster, C. M., Nallathamby, P. D., & Retterer, S. T. (2013). Dynamic development of the protein corona on silica nanoparticles: Composition and role in toxicity. *Nanoscale*, 5(14), 6372-6380. doi:10.1039/c3nr33280b [doi]
- [90] Pozzi, D., Caracciolo, G., Capriotti, A. L., Cavaliere, C., La Barbera, G., Anchordoqui, T. J., & Laganà, A. (2015). Surface chemistry and serum type both determine the nanoparticle-protein corona. *Journal of Proteomics*, 119, 209-217. doi:10.1016/j.jprot.2015.02.009
- [91] Moyano, D. F., Saha, K., Prakash, G., Yan, B., Kong, H., Yazdani, M., & Rotello, V. M. (2014). Fabrication of corona-free nanoparticles with tunable hydrophobicity. *ACS Nano*, 8(7), 6748-6755. doi:10.1021/nn5006478 [doi]
- [92] Lindman, S., Lynch, I., Thulin, E., Nilsson, H., Dawson, K. A., & Linse, S. (2007). Systematic investigation of the thermodynamics of HSA adsorption to N-isopropylacrylamide/N-tert-butylacrylamide copolymer nanoparticles. effects of particle size and hydrophobicity. *Nano Letters*, 7(4), 914-920. doi:10.1021/nl062743+ [doi]
- [93] Roach, P., Farrar, D., & Perry, C. C. (2006). Surface tailoring for controlled protein adsorption: effect of topography at the nanometer scale and chemistry. *Journal of the American Chemical Society*, 128(12), 3939-3945. doi:10.1021/ja056278e
- [94] Belitz, H. D., Grosch, W., & Schieberle, P. (2009). *Food chemistry* (4th ed.). Berlin, Germany: Springer.
- [95] Damodaran, S., Parkin, K. L., & Fennema, O. R. (2007). *Fennema's food chemistry* (4th ed.). Boca Raton, FL: CRC Press.
- [96] Klaine, S. J., Alvarez, P. J. J., Batley, G. E., Fernandes, T. F., Handy, R. D., Lyon, D. Y., . . . Lead, J. R. (2008). *Nanomaterials in the environment: Behavior, fate,*

- bioavailability, and effects. *Environmental Toxicology and Chemistry*, 27(9), 1825-1851. doi:10.1897/08-090.1
- [97] Buffle, J., Wilkinson, K. J., Stoll, S., Filella, M., & Zhang, J. (1998). A generalized description of aquatic colloidal interactions: the three-colloidal component approach. *Environmental Science & Technology*, 32(19), 2887-2899. doi:10.1021/es980217h
  - [98] Chetwynd, A. J., Zhang, W., Thorn, J. A., Lynch, I., & Ramautar, R. (2020). The nanomaterial metabolite corona determined using a quantitative metabolomics approach: A pilot study. *Small*, 16(21), 2000295. doi:10.1002/sml.202000295
  - [99] Peters, R., Kramer, E., Oomen, A. G., Rivera, Z. E., Oegema, G., Tromp, P. C., . . . Bouwmeester, H. (2012). Presence of nano-sized silica during in vitro digestion of foods containing silica as a food additive. *ACS Nano*, 6(3), 2441-2451. doi:10.1021/nn204728k [doi]
  - [100] Di Silvio, D., Rigby, N., Bajka, B., Mackie, A., & Baldelli Bombelli, F. (2016). Effect of protein corona magnetite nanoparticles derived from bread in vitro digestion on caco-2 cells morphology and uptake. *The International Journal of Biochemistry & Cell Biology*, 75, 212-222. doi:https://doi.org/10.1016/j.biocel.2015.10.019
  - [101] Ranjan, S., Dasgupta, N., Srivastava, P., & Ramalingam, C. (2016). A spectroscopic study on interaction between bovine serum albumin and titanium dioxide nanoparticle synthesized from microwave-assisted hybrid chemical approach. *Journal of Photochemistry and Photobiology B: Biology*, 161, 472-481. doi:https://doi.org/10.1016/j.jphotobiol.2016.06.015
  - [102] Dekkers, B. L., Kolodziejczyk, E., Acquistapace, S., Engmann, J., & Wooster, T. J. (2016). Impact of gastric pH profiles on the proteolytic digestion of mixed  $\beta$ lg-xanthan biopolymer gels. *Food & Function*, 7(1), 58-68. doi:10.1039/c5fo01085c [doi]
  - [103] Zhang, Z., Decker, E. A., & McClements, D. J. (2014). Encapsulation, protection, and release of polyunsaturated lipids using biopolymer-based hydrogel particles. *Food Research International*, 64, 520-526. doi:https://doi.org/10.1016/j.foodres.2014.07.020
  - [104] Dickinson, E. (2010). Flocculation of protein-stabilized oil-in-water emulsions. *Colloids and Surfaces.B, Biointerfaces*, 81(1), 130-140. doi:10.1016/j.colsurfb.2010.06.033 [doi]
  - [105] Cao, X., Han, Y., Li, F., Li, Z., McClements, D. J., He, L., . . . Xiao, H. (2019). Impact of protein-nanoparticle interactions on gastrointestinal fate of ingested nanoparticles: Not just simple protein corona effects. *Nanoimpact*, 13, 37-43. doi:https://doi.org/10.1016/j.impact.2018.12.002
  - [106] Sit, I., Xu, Z., & Grassian, V. H. (2019). Plasma protein adsorption on TiO<sub>2</sub> nanoparticles: Impact of surface adsorption on temperature-dependent structural changes. *Polyhedron*, 171, 147-154. doi:https://doi.org/10.1016/j.poly.2019.06.036

- [107] Nikpasand, A., & Parvizi, M. R. (2019). Evaluation of the effect of titanium dioxide Nanoparticles/Gelatin composite on infected skin wound healing; an animal model study. *Bulletin of Emergency and Trauma*, 7(4), 366-372. doi:10.29252/beat-070405
- [108] Giri, K., Shameer, K., Zimmermann, M. T., Saha, S., Chakraborty, P. K., Sharma, A., . . . Mukherjee, P. (2014). Understanding protein-nanoparticle interaction: A new gateway to disease therapeutics. *Bioconjugate Chemistry*, 25(6), 1078-1090. doi:10.1021/bc500084f [doi]
- [109] Monopoli, M. P., Aberg, C., Salvati, A., & Dawson, K. A. (2012). Biomolecular coronas provide the biological identity of nanosized materials. *Nature Nanotechnology*, 7(12), 779-786. doi:10.1038/nnano.2012.207 [doi]
- [110] McClements, D. J., & Xiao, H. (2017). Is nano safe in foods? establishing the factors impacting the gastrointestinal fate and toxicity of organic and inorganic food-grade nanoparticles. *NPJ Science of Food*, 1, 6-017-0005-1. eCollection 2017. doi:10.1038/s41538-017-0005-1 [doi]
- [111] Runa, S., Lakadamyali, M., Kemp, M. L., & Payne, C. K. (2017). TiO<sub>2</sub> nanoparticle-induced oxidation of the plasma membrane: Importance of the protein corona. *The Journal of Physical Chemistry.B*, 121(37), 8619-8625. doi:10.1021/acs.jpcc.7b04208 [doi]
- [112] Zhao, F., Holmberg, J. P., Abbas, Z., Frost, R., Sirkka, T., Kasemo, B., . . . Svedhem, S. (2016). TiO<sub>2</sub> nanoparticle interactions with supported lipid membranes – an example of removal of membrane patches. *RSC Advances*, 6(94), 91102-91110. doi:10.1039/C6RA05693H
- [113] Aranha, M. P., Mukherjee, D., Petridis, L., & Khomami, B. (2020). An atomistic molecular dynamics study of titanium dioxide adhesion to lipid bilayers. *Langmuir*, 36(4), 1043-1052. doi:10.1021/acs.langmuir.9b03075
- [114] Foliatini, F., Yulizar, Y., & Hafizah, M. (2014). Theoretical analysis of interaction energy in alginate-capped gold nanoparticles colloidal system. *Indonesian Journal of Chemistry*, 14, 239-245. doi:10.22146/ijc.21234
- [115] McClements, D. (2015). *Nanoparticle- and microparticle-based delivery systems*. Boca Raton: CRC Press. doi:https://doi.org/10.1201/b17280
- [116] Chen, X. X., Cheng, B., Yang, Y. X., Cao, A., Liu, J. H., Du, L. J., . . . Wang, H. (2013). Characterization and preliminary toxicity assay of nano-titanium dioxide additive in sugar-coated chewing gum. *Small (Weinheim an Der Bergstrasse, Germany)*, 9(9-10), 1765-1774. doi:10.1002/sml.201201506 [doi]
- [117] Borreani, J., Llorca, E., Larrea, V., & Hernando, I. (2016). Adding neutral or anionic hydrocolloids to dairy proteins under in vitro gastric digestion conditions. *Food Hydrocolloids*, 57, 169-177. doi:https://doi.org/10.1016/j.foodhyd.2016.01.030
- [118] Fabek, H., Messerschmidt, S., Brulport, V., & Goff, H. D. (2014). The effect of in vitro digestive processes on the viscosity of dietary fibres and their influence on



- glucose diffusion. *Food Hydrocolloids*, 35, 718-726.  
doi:<https://doi.org/10.1016/j.foodhyd.2013.08.007>
- [119] Hoad, C. L., Rayment, P., Spiller, R. C., Marciani, L., Alonso Bde, C., Traynor, C., . . . Gowland, P. A. (2004). In vivo imaging of intragastric gelation and its effect on satiety in humans. *The Journal of Nutrition*, 134(9), 2293-2300.  
doi:134/9/2293 [pii]
  - [120] Dickinson, E. (2003). Hydrocolloids at interfaces and the influence on the properties of dispersed systems. *Food Hydrocolloids*, 17(1), 25-39.  
doi:[https://doi.org/10.1016/S0268-005X\(01\)00120-5](https://doi.org/10.1016/S0268-005X(01)00120-5)
  - [121] McClements, D. J., & Gumus, C. E. (2016). Natural emulsifiers - biosurfactants, phospholipids, biopolymers, and colloidal particles: Molecular and physicochemical basis of functional performance. *Advances in Colloid and Interface Science*, 234, 3-26. doi:S0001-8686(16)30038-0 [pii]
  - [122] Dona, A. C., Pages, G., Gilbert, R. G., & Kuchel, P. W. (2010). Digestion of starch: In vivo and in vitro kinetic models used to characterise oligosaccharide or glucose release. *Carbohydrate Polymers*, 80(3), 599-617.  
doi:<https://doi.org/10.1016/j.carbpol.2010.01.002>
  - [123] Lovegrove, A., Edwards, C. H., De Noni, I., Patel, H., El, S. N., Grassby, T., . . . Shewry, P. R. (2017). Role of polysaccharides in food, digestion, and health. *Critical Reviews in Food Science and Nutrition*, 57(2), 237-253. doi:939263 [pii]
  - [124] Myhrstad, M. C. W., Tunsjø, H., Charnock, C., & Telle-Hansen, V. H. (2020). Dietary fiber, gut microbiota, and metabolic regulation-current status in human randomized trials. *Nutrients*, 12(3), 859. doi: 10.3390/nu12030859.  
doi:10.3390/nu12030859 [doi]
  - [125] Chen, Z., Wang, Y., Zhuo, L., Chen, S., Zhao, L., Chen, T., . . . Jia, G. (2015). Interaction of titanium dioxide nanoparticles with glucose on young rats after oral administration. *Nanomedicine: Nanotechnology, Biology and Medicine*, 11(7), 1633-1642. doi:<https://doi.org/10.1016/j.nano.2015.06.002>
  - [126] Liu, W., Zhou, X., Xu, L., Zhu, S., Yang, S., Chen, X., . . . Song, H. (2019). Graphene quantum dot-functionalized three-dimensional ordered mesoporous ZnO for acetone detection toward diagnosis of diabetes. *Nanoscale*, 11(24), 11496-11504. doi:10.1039/C9NR00942F
  - [127] Pakulski, D., Czepa, W. I., Witomska, S., Aliprandi, A., Pawluć, P., Patroniak, V., . . . Samorì, P. (2018). Graphene oxide-branched polyethylenimine foams for efficient removal of toxic cations from water. *Journal of Materials Chemistry A*, 6(20), 9384-9390. doi:10.1039/C8TA01622D
  - [128] Gartiser, S., Flach, F., Nickel, C., Stintz, M., Damme, S., Schaeffer, A., . . . Kuhlbusch, T. A. J. (2014). Behavior of nanoscale titanium dioxide in laboratory wastewater treatment plants according to OECD 303 A. *Chemosphere*, 104, 197-204. doi:<https://doi.org/10.1016/j.chemosphere.2013.11.015>
  - [129] Gómez-Rivera, F., Field, J. A., Brown, D., & Sierra-Alvarez, R. (2012). Fate of cerium dioxide (CeO<sub>2</sub>) nanoparticles in municipal wastewater during activated

sludge treatment. *Bioresource Technology*, 108, 300-304.  
doi:<https://doi.org/10.1016/j.biortech.2011.12.113>

- [130] Auger, S., Henry, C., Péchaux, C., Lejal, N., Zanet, V., Nikolic, M. V., . . . Vidic, J. (2019). Exploring the impact of mg-doped ZnO nanoparticles on a model soil microorganism *bacillus subtilis*. *Ecotoxicology and Environmental Safety*, 182, 109421. doi:S0147-6513(19)30751-1 [pii]
- [131] Li, M., Yang, Y., Xie, J., Xu, G., & Yu, Y. (2019). In-vivo and in-vitro tests to assess toxic mechanisms of nano ZnO to earthworms. *The Science of the Total Environment*, 687, 71-76. doi:S0048-9697(19)32536-7 [pii]
- [132] Lee, W. M., & An, Y. J. (2013). Effects of zinc oxide and titanium dioxide nanoparticles on green algae under visible, UVA, and UVB irradiations: No evidence of enhanced algal toxicity under UV pre-irradiation. *Chemosphere*, 91(4), 536-544. doi:S0045-6535(12)01534-2 [pii]
- [133] Shi, W., Han, Y., Guo, C., Su, W., Zhao, X., Zha, S., . . . Liu, G. (2019). Ocean acidification increases the accumulation of titanium dioxide nanoparticles (nTiO<sub>2</sub>) in edible bivalve mollusks and poses a potential threat to seafood safety. *Scientific Reports*, 9(1), 3516-019-40047-1. doi:10.1038/s41598-019-40047-1 [doi]
- [134] Canesi, L., Frenzilli, G., Balbi, T., Bernardeschi, M., Ciacci, C., Corsolini, S., . . . Corsi, I. (2014). Interactive effects of n-TiO<sub>2</sub> and 2,3,7,8-TCDD on the marine bivalve *mytilus galloprovincialis*. *Aquatic Toxicology*, 153, 53-65. doi:<https://doi.org/10.1016/j.aquatox.2013.11.002>
- [135] Fan, X., Wang, C., Wang, P., Hu, B., & Wang, X. (2018). TiO<sub>2</sub> nanoparticles in sediments: Effect on the bioavailability of heavy metals in the freshwater bivalve *corbicula fluminea*. *Journal of Hazardous Materials*, 342, 41-50. doi:<https://doi.org/10.1016/j.jhazmat.2017.07.041>
- [136] Manske Nunes, S., Josende, M. E., González-Durruthy, M., Pires Ruas, C., Gelesky, M. A., Romano, L. A., . . . Ventura-Lima, J. (2018). Different crystalline forms of titanium dioxide nanomaterial (rutile and anatase) can influence the toxicity of copper in golden mussel *limnoperna fortunei*? *Aquatic Toxicology*, 205, 182-192. doi:<https://doi.org/10.1016/j.aquatox.2018.10.009>
- [137] Nunes, S. M., Müller, L., Simioni, C., Ouriques, L. C., Gelesky, M. A., Fattorini, D., . . . Ventura-Lima, J. (2020). Impact of different crystalline forms of nTiO<sub>2</sub> on metabolism and arsenic toxicity in *limnoperna fortunei*. *The Science of the Total Environment*, 728, 138318. doi:S0048-9697(20)31831-3 [pii]
- [138] Sendra, M., Pintado-Herrera, M. G., Aguirre-Martínez, G. V., Moreno-Garrido, I., Martín-Díaz, L. M., Lara-Martín, P. A., & J, B. (2017). Are the TiO<sub>2</sub> NPs a “Trojan horse” for personal care products (PCPs) in the clam *ruditapes philippinarum*? *Chemosphere*, 185, 192-204. doi:<https://doi.org/10.1016/j.chemosphere.2017.07.009>
- [139] Shi, W., Guan, X., Han, Y., Zha, S., Fang, J., Xiao, G., . . . Liu, G. (2018). The synergic impacts of TiO<sub>2</sub> nanoparticles and 17 $\beta$ -estradiol (E2) on the immune

- responses, E2 accumulation, and expression of immune-related genes of the blood clam, *tegillarca granosa*. *Fish & Shellfish Immunology*, 81, 29-36. doi:S1050-4648(18)30407-8 [pii]
- [140] Tian, S., Zhang, Y., Song, C., Zhu, X., & Xing, B. (2014). Titanium dioxide nanoparticles as carrier facilitate bioaccumulation of phenanthrene in marine bivalve, ark shell (*scapharca subcrenata*). *Environmental Pollution* (Barking, Essex : 1987), 192, 59-64. doi:S0269-7491(14)00198-5 [pii]
- [141] Farkas, J., Bergum, S., Nilsen, E. W., Olsen, A. J., Salaberria, I., Ciesielski, T. M., . . . Jenssen, B. M. (2015). The impact of TiO<sub>2</sub> nanoparticles on uptake and toxicity of benzo(a)pyrene in the blue mussel (*mytilus edulis*). *Science of the Total Environment*, 511, 469-476.  
doi:https://doi.org/10.1016/j.scitotenv.2014.12.084
- [142] Della Torre, C., Balbi, T., Grassi, G., Frenzilli, G., Bernardeschi, M., Smerilli, A., . . . Corsi, I. (2015). Titanium dioxide nanoparticles modulate the toxicological response to cadmium in the gills of *mytilus galloprovincialis*. *Journal of Hazardous Materials*, 297, 92-100. doi:S0304-3894(15)00370-2 [pii]
- [143] Rocco, L., Santonastaso, M., Nigro, M., Mottola, F., Costagliola, D., Bernardeschi, M., . . . Frenzilli, G. (2015). Genomic and chromosomal damage in the marine mussel *mytilus galloprovincialis*: Effects of the combined exposure to titanium dioxide nanoparticles and cadmium chloride. *Marine Environmental Research*, 111, 144-148. doi:S0141-1136(15)30043-X [pii]
- [144] Balbi, T., Smerilli, A., Fabbri, R., Ciacci, C., Montagna, M., Grasselli, E., . . . Canesi, L. (2014). Co-exposure to n-TiO<sub>2</sub> and Cd<sup>2+</sup> results in interactive effects on biomarker responses but not in increased toxicity in the marine bivalve *M. galloprovincialis*. *The Science of the Total Environment*, 493, 355-364.  
doi:S0048-9697(14)00838-9 [pii]
- [145] Banni, M., Sforzini, S., Balbi, T., Corsi, I., Viarengo, A., & Canesi, L. (2016). Combined effects of n-TiO<sub>2</sub> and 2,3,7,8-TCDD in *mytilus galloprovincialis* digestive gland: A transcriptomic and immunohistochemical study. *Environmental Research*, 145, 135-144. doi:S0013-9351(15)30164-X [pii]
- [146] Vale, G., Franco, C., Diniz, M. S., dos Santos, M. M., & Domingos, R. F. (2014). Bioavailability of cadmium and biochemical responses on the freshwater bivalve *corbicula fluminea*--the role of TiO<sub>2</sub> nanoparticles. *Ecotoxicology and Environmental Safety*, 109, 161-168. doi:S0147-6513(14)00355-8 [pii]
- [147] Chen, H. (2018). Metal based nanoparticles in agricultural system: Behavior, transport, and interaction with plants.30(1), 123-134.  
doi:10.1080/09542299.2018.1520050
- [148] Servin, A. D., Morales, M. I., Castillo-Michel, H., Hernandez-Viezcas, J., Munoz, B., Zhao, L., . . . Gardea-Torresdey, J. (2013). Synchrotron verification of TiO<sub>2</sub> accumulation in cucumber fruit: A possible pathway of TiO<sub>2</sub> nanoparticle transfer from soil into the food chain. *Environmental Science & Technology*, 47(20), 11592-11598. doi:10.1021/es403368j

- [149] Pošćić, F., Mattiello, A., Fellet, G., Miceli, F., & Marchiol, L. (2016). Effects of cerium and titanium oxide nanoparticles in soil on the nutrient composition of barley (*hordeum vulgare* L.) kernels. *International Journal of Environmental Research and Public Health*, 13(6), 577. doi: 10.3390/ijerph13060577. doi:10.3390/ijerph13060577 [doi]
- [150] Zahra, Z., Arshad, M., Rafique, R., Mahmood, A., Habib, A., Qazi, I. A., & Khan, S. A. (2015). Metallic nanoparticle (TiO<sub>2</sub> and Fe<sub>3</sub>O<sub>4</sub>) application modifies rhizosphere phosphorus availability and uptake by *lactuca sativa*. *Journal of Agricultural and Food Chemistry*, 63(31), 6876-6882. doi:10.1021/acs.jafc.5b01611 [doi]
- [151] Vittori Antisari, L., Carbone, S., Gatti, A., Vianello, G., & Nannipieri, P. (2015). Uptake and translocation of metals and nutrients in tomato grown in soil polluted with metal oxide (CeO<sub>2</sub>, Fe<sub>3</sub>O<sub>4</sub>, SnO<sub>2</sub>, TiO<sub>2</sub>) or metallic (ag, co, ni) engineered nanoparticles. *Environmental Science and Pollution Research International*, 22(3), 1841-1853. doi:10.1007/s11356-014-3509-0 [doi]
- [152] Burke, D. J., Pietrasiak, N., Situ, S. F., Abenojar, E. C., Porche, M., Kraj, P., . . . Samia, A. C. S. (2015). Iron oxide and titanium dioxide nanoparticle effects on plant performance and root associated microbes. *International Journal of Molecular Sciences*, 16(10), 23630-23650. doi:10.3390/ijms161023630
- [153] Ebrahimi, A., Galavi, M., Ramroudi, M., & Moaveni, P. (2016). Effect of TiO<sub>2</sub> nanoparticles on antioxidant enzymes activity and biochemical biomarkers in pinto bean (*phaseolus vulgaris* L.). *Journal of Molecular Biology Research*, 6, 58. doi:10.5539/jmbr.v6n1p58
- [154] Liu, Y., Yue, L., Zhenyu, W., & Xing, B. (2019). Processes and mechanisms of photosynthesis augmented by engineered nanomaterials. *Environmental Chemistry*, 16 doi:10.1071/EN19046
- [155] NIOSH. (2011). Occupational exposure to titanium dioxide. in: *Current intelligence bulletin*, 63 (national institute for occupational safety and health, cincinnati. retrieved from <https://www.cdc.gov/niosh/docs/2011-160/default.html>).
- [156] Shi, H., Magaye, R., Castranova, V., & Zhao, J. (2013). Titanium dioxide nanoparticles: A review of current toxicological data. *Particle and Fibre Toxicology*, 10, 15-8977-10-15. doi:10.1186/1743-8977-10-15 [doi]
- [157] Ursini, C. L., Cavallo, D., Freseigna, A. M., Ciervo, A., Maiello, R., Tassone, P., . . . Iavicoli, S. (2014). Evaluation of cytotoxic, genotoxic and inflammatory response in human alveolar and bronchial epithelial cells exposed to titanium dioxide nanoparticles. *Journal of Applied Toxicology : JAT*, 34(11), 1209-1219. doi:10.1002/jat.3038 [doi]
- [158] Yin, J., Liu, J., Ehrenshaft, M., Roberts, J. E., Fu, P. P., Mason, R. P., & Zhao, B. (2012). Phototoxicity of nano titanium dioxides in HaCaT keratinocytes--generation of reactive oxygen species and cell damage. *Toxicology and Applied Pharmacology*, 263(1), 81-88. doi:10.1016/j.taap.2012.06.001

- [159] Ranjan, S., & Ramalingam, C. (2016). Titanium dioxide nanoparticles induce bacterial membrane rupture by reactive oxygen species generation. *Environmental Chemistry Letters*, 14(4), 487-494. doi:10.1007/s10311-016-0586-y
- [160] Proquin, H., Rodríguez-Ibarra, C., Moonen, C. G., Urrutia Ortega, I. M., Briedé, J. J., de Kok, T. M., . . . Chirino, Y. I. (2017). Titanium dioxide food additive (E171) induces ROS formation and genotoxicity: Contribution of micro and nano-sized fractions. *Mutagenesis*, 32(1), 139-149. doi:10.1093/mutage/gew051 [doi]
- [161] Rizk, M. Z., Ali, S. A., Hamed, M. A., El-Rigal, N. S., Aly, H. F., & Salah, H. H. (2017). Toxicity of titanium dioxide nanoparticles: Effect of dose and time on biochemical disturbance, oxidative stress and genotoxicity in mice. *Biomedicine & Pharmacotherapy*, 90, 466-472. doi:https://doi.org/10.1016/j.biopha.2017.03.089
- [162] Shi, Y., Wang, F., He, J., Yadav, S., & Wang, H. (2010). Titanium dioxide nanoparticles cause apoptosis in BEAS-2B cells through the caspase 8/t-bid-independent mitochondrial pathway. *Toxicology Letters*, 196(1), 21-27. doi:10.1016/j.toxlet.2010.03.014 [doi]
- [163] Ammendolia, M. G., Iosi, F., Maranghi, F., Tassinari, R., Cubadda, F., Aureli, F., . . . De Berardis, B. (2017). Short-term oral exposure to low doses of nano-sized TiO<sub>2</sub> and potential modulatory effects on intestinal cells. *Food and Chemical Toxicology : An International Journal Published for the British Industrial Biological Research Association*, 102, 63-75. doi:S0278-6915(17)30039-X [pii]
- [164] Di Bucchianico, S., Cappellini, F., Le Bihanic, F., Zhang, Y., Dreij, K., & Karlsson, H. L. (2017). Genotoxicity of TiO<sub>2</sub> nanoparticles assessed by mini-gel comet assay and micronucleus scoring with flow cytometry. *Mutagenesis*, 32(1), 127-137. doi:10.1093/mutage/gew030 [doi]
- [165] Geppert, M., Schwarz, A., Stangassinger, L. M., Wenger, S., Wienerroither, L. M., Ess, S., . . . Himly, M. (2020). Interactions of TiO<sub>2</sub> nanoparticles with ingredients from modern lifestyle products and their effects on human skin cells. *Chemical Research in Toxicology*, 33(5), 1215-1225. doi:10.1021/acs.chemrestox.9b00428
- [166] Sadrieh, N., Wokovich, A. M., Gopee, N. V., Zheng, J., Haines, D., Parmiter, D., . . . Buhse, L. F. (2010). Lack of significant dermal penetration of titanium dioxide from sunscreen formulations containing nano- and submicron-size TiO<sub>2</sub> particles. *Toxicological Sciences : An Official Journal of the Society of Toxicology*, 115(1), 156-166. doi:10.1093/toxsci/kfq041
- [167] Newman, M. D., Stotland, M., & Ellis, J. I. (2009). The safety of nanosized particles in titanium dioxide- and zinc oxide-based sunscreens. *Journal of the American Academy of Dermatology*, 61(4), 685-692. doi:10.1016/j.jaad.2009.02.051 [doi]
- [168] Shakeel, M., Jabeen, F., Shabbir, S., Asghar, M. S., Khan, M. S., & Chaudhry, A. S. (2016). Toxicity of nano-titanium dioxide (TiO<sub>2</sub>-NP) through various routes of exposure: A review. *Biological Trace Element Research*, 172(1), 1-36. doi:10.1007/s12011-015-0550-x [doi]

- [169] Crosera, M., Prodi, A., Mauro, M., Pelin, M., Florio, C., Bellomo, F., . . . Filon, F. L. (2015). Titanium dioxide nanoparticle penetration into the skin and effects on HaCaT cells. *International Journal of Environmental Research and Public Health*, 12(8), 9282-9297. doi:10.3390/ijerph120809282
- [170] Xie, G., Lu, W., & Lu, D. (2015). Penetration of titanium dioxide nanoparticles through slightly damaged skin in vitro and in vivo. *Journal of Applied Biomaterials & Functional Materials*, 13(4), e356-61. doi:BA21F11F-D909-4174-9D26-6810E9C438F0 [pii]
- [171] Wu, J., Liu, W., Xue, C., Zhou, S., Lan, F., Bi, L., . . . Zeng, F. D. (2009). Toxicity and penetration of TiO<sub>2</sub> nanoparticles in hairless mice and porcine skin after subchronic dermal exposure. *Toxicology Letters*, 191(1), 1-8. doi:10.1016/j.toxlet.2009.05.020 [doi]
- [172] Sanches, P. L., Geaquinto, L. R. O., Cruz, R., Schuck, D. C., Lorencini, M., Granjeiro, J. M., & Ribeiro, A. R. L. (2020). Toxicity evaluation of TiO<sub>2</sub> nanoparticles on the 3D skin model: A systematic review. *Frontiers in Bioengineering and Biotechnology*, 8, 575. doi:10.3389/fbioe.2020.00575 [doi]
- [173] Horie, M., Sugino, S., Kato, H., Tabei, Y., Nakamura, A., & Yoshida, Y. (2016). Does photocatalytic activity of TiO<sub>2</sub> nanoparticles correspond to photocytotoxicity? cellular uptake of TiO<sub>2</sub> nanoparticles is important in their photocytotoxicity. *Toxicology Mechanisms and Methods*, 26(4), 284-294. doi:10.1080/15376516.2016.1175530 [doi]
- [174] Pelcova, D., Navratil, T., Kacerova, T., Zamostna, B., Fenclova, Z., Vlckova, S., & Kacer, P. (2019). NanoTiO<sub>2</sub> sunscreen does not prevent systemic oxidative stress caused by UV radiation and a minor amount of NanoTiO<sub>2</sub> is absorbed in humans. *Nanomaterials (Basel, Switzerland)*, 9(6), 888. doi:10.3390/nano9060888
- [175] Wright, C., Iyer, A. K., Wang, L., Wu, N., Yakisich, J. S., Rojanasakul, Y., & Azad, N. (2017). Effects of titanium dioxide nanoparticles on human keratinocytes. *Drug and Chemical Toxicology*, 40(1), 90-100. doi:10.1080/01480545.2016.1185111 [doi]
- [176] Pan, Z., Lee, W., Slutsky, L., Clark, R. A., Pernodet, N., & Rafailovich, M. H. (2009). Adverse effects of titanium dioxide nanoparticles on human dermal fibroblasts and how to protect cells. *Small (Weinheim an Der Bergstrasse, Germany)*, 5(4), 511-520. doi:10.1002/sml.200800798 [doi]
- [177] de Souza Carvalho, C., Daum, N., & Lehr, C. M. (2014). Carrier interactions with the biological barriers of the lung: Advanced in vitro models and challenges for pulmonary drug delivery. *Advanced Drug Delivery Reviews*, 75, 129-140. doi:S0169-409X(14)00126-4 [pii]
- [178] Löndahl, J., Möller, W., Pagels, J. H., Kreyling, W. G., Swietlicki, E., & Schmid, O. (2014). Measurement techniques for respiratory tract deposition of airborne nanoparticles: A critical review. *Journal of Aerosol Medicine and Pulmonary Drug Delivery*, 27(4), 229-254. doi:10.1089/jamp.2013.1044 [doi]

- [179] Ep, H. (1881). Essays on the floating-matter of the air in relation to putrefaction and infection. *Nature*, 25(6)
- [180] Brain, J. D., & Valberg, P. A. (1979). Deposition of aerosol in the respiratory tract. *The American Review of Respiratory Disease*, 120(6), 1325-1373. doi:10.1164/arrd.1979.120.6.1325 [doi]
- [181] Heyder, J. (2004). Deposition of inhaled particles in the human respiratory tract and consequences for regional targeting in respiratory drug delivery. *Proceedings of the American Thoracic Society*, 1(4), 315-320. doi:1/4/315 [pii]
- [182] Anderson, J. O., Thundiyil, J. G., & Stolbach, A. (2012). Clearing the air: A review of the effects of particulate matter air pollution on human health. *Journal of Medical Toxicology : Official Journal of the American College of Medical Toxicology*, 8(2), 166-175. doi:10.1007/s13181-011-0203-1 [doi]
- [183] Deng, Q., Deng, L., Miao, Y., Guo, X., & Li, Y. (2019). Particle deposition in the human lung: Health implications of particulate matter from different sources. *Environmental Research*, 169, 237-245. doi:S0013-9351(18)30596-6 [pii]
- [184] Klein, S. G., Hennen, J., Serchi, T., Blömeke, B., & Gutleb, A. C. (2011). Potential of coculture in vitro models to study inflammatory and sensitizing effects of particles on the lung. *Toxicology in Vitro*, 25(8), 1516-1534. doi:https://doi.org/10.1016/j.tiv.2011.09.006
- [185] Oberdörster, G., Oberdörster, E., & Oberdörster, J. (2005). Nanotoxicology: An emerging discipline evolving from studies of ultrafine particles. *Environmental Health Perspectives*, 113(7), 823-839. doi:10.1289/ehp.7339
- [186] Chen, S., Guo, H., Cui, M., Huang, R., Su, R., Qi, W., & He, Z. (2020). Interaction of particles with mucosae and cell membranes. *Colloids and Surfaces B: Biointerfaces*, 186, 110657. doi:https://doi.org/10.1016/j.colsurfb.2019.110657
- [187] Tu, K. W., & Knutson, E. O. (1984). Total deposition of ultrafine hydrophobic and hygroscopic aerosols in the human respiratory system. 3(4), 453-465. doi:10.1080/02786828408959032
- [188] Halappanavar, S., Saber, A. T., Decan, N., Jensen, K. A., Wu, D., Jacobsen, N. R., . . . Vogel, U. (2015). Transcriptional profiling identifies physicochemical properties of nanomaterials that are determinants of the in vivo pulmonary response. *Environmental and Molecular Mutagenesis*, 56(2), 245-264. doi:10.1002/em.21936 [doi]
- [189] Le, H. Q., Tomenson, J. A., Warheit, D. B., Fryzek, J. P., Golden, A. P., & Ellis, E. D. (2018). A review and meta-analysis of occupational titanium dioxide exposure and lung cancer mortality. *Journal of Occupational and Environmental Medicine*, 60(7), e356-e367. doi:10.1097/JOM.0000000000001314 [doi]
- [190] Nikota, J., Williams, A., Yauk, C. L., Wallin, H., Vogel, U., & Halappanavar, S. (2016). Meta-analysis of transcriptomic responses as a means to identify pulmonary disease outcomes for engineered nanomaterials. *Particle and Fibre Toxicology*, 13(1), 25-016-0137-5. doi:10.1186/s12989-016-0137-5 [doi]

- [191] PAYNE, J. M., SANSOM, B. F., GARNER, R. J., THOMSON, A. R., & MILES, B. J. (1960). Uptake of small resin particles (1–5 $\mu$  diameter) by the alimentary canal of the calf. *Nature*, 188(4750), 586-567.
- [192] SANDERS, E., & ASHWORTH, C. T. (1961). A study of particulate intestinal absorption and hepatocellular uptake. use of polystyrene latex particles. *Experimental Cell Research*, 22, 137-145. doi:0014-4827(61)90092-1 [pii]
- [193] Yamago, S., Tokuyama, H., Nakamura, E., Kikuchi, K., Kananishi, S., Sueki, K., . . . Ambe, F. (1995). In vivo biological behavior of a water-miscible fullerene: <sup>14</sup>C labeling, absorption, distribution, excretion and acute toxicity. *Chemistry & Biology*, 2(6), 385-389. doi:[https://doi.org/10.1016/1074-5521\(95\)90219-8](https://doi.org/10.1016/1074-5521(95)90219-8)
- [194] ALPAR, H. O., FIELD, W. N., HYDE, R., & LEWIS, D. A. (1989). The transport of microspheres from the gastro-intestinal tract to inflammatory air pouches in the rat. *Journal of Pharmacy and Pharmacology*, 41(3), 194-196. doi:10.1111/j.2042-7158.1989.tb06429.x
- [195] LeFevre, M. E., Boccio, A. M., & Joel, D. D. (1989). Intestinal uptake of fluorescent microspheres in young and aged mice. *Proceedings of the Society for Experimental Biology and Medicine*. Society for Experimental Biology and Medicine (New York, N.Y.), 190(1), 23-27. doi:10.3181/00379727-190-42825 [doi]
- [196] LeFevre, M. E., Vanderhoff, J. W., Laissue, J. A., & Joel, D. D. (1978). Accumulation of 2-micron latex particles in mouse peyer's patches during chronic latex feeding. *Experientia*, 34(1), 120-122. doi:10.1007/BF01921939 [doi]
- [197] Aprahamian, M., Michel, C., Humbert, W., Devissaguet, J. P., & Dange, C. (1987). Transmucosal passage of polyalkylcyanoacrylate nanocapsules as a new drug carrier in the small intestine. *Biology of the Cell*, 61(1), 69-76. doi:10.1111/j.1768-322X.1987.tb00571.x
- [198] Volkheimer, G. (1975). Hematogenous dissemination of ingested polyvinyl chloride particles. *Annals of the New York Academy of Sciences*, 246, 164-171. doi:10.1111/j.1749-6632.1975.tb51092.x [doi]
- [199] Yada, R. Y., Buck, N., Canady, R., DeMerlis, C., Duncan, T., Janer, G., . . . Thurmond, S. (2014). Engineered nanoscale food ingredients: Evaluation of current knowledge on material characteristics relevant to uptake from the gastrointestinal tract. *Comprehensive Reviews in Food Science and Food Safety*, 13(4), 730-744. doi:10.1111/1541-4337.12076
- [200] De Hoog, Els H. A., Prinz, J., Huntjens, L., Dresselhuis, D., & Van Aken, G. A. (2006). Lubrication of oral surfaces by food emulsions: The importance of surface characteristics. *Journal of Food Science*, 71(7), E337-E341. doi:10.1111/j.1750-3841.2006.00140.x
- [201] Silletti, E., Vingerhoeds, M. H., Norde, W., & van Aken, G. A. (2007). The role of electrostatics in saliva-induced emulsion flocculation. *Food Hydrocolloids*, 21(4), 596-606. doi:<https://doi.org/10.1016/j.foodhyd.2006.07.004>



- [202] Vingerhoeds, M. H., Blijdenstein, T. B. J., Zoet, F. D., & van Aken, G. A. (2005). Emulsion flocculation induced by saliva and mucin. *Food Hydrocolloids*, 19(5), 915-922. doi:<https://doi.org/10.1016/j.foodhyd.2004.12.005>
- [203] Mahmoudi, M., Lynch, I., Ejtehadi, M. R., Monopoli, M. P., Bombelli, F. B., & Laurent, S. (2011). Protein-nanoparticle interactions: Opportunities and challenges. *Chemical Reviews*, 111(9), 5610-5637. doi:10.1021/cr100440g [doi]
- [204] Aggarwal, P., Hall, J. B., McLeland, C. B., Dobrovolskaia, M. A., & McNeil, S. E. (2009). Nanoparticle interaction with plasma proteins as it relates to particle biodistribution, biocompatibility and therapeutic efficacy. *Advanced Drug Delivery Reviews*, 61(6), 428-437. doi:10.1016/j.addr.2009.03.009 [doi]
- [205] Dobrovolskaia, M. A., Patri, A. K., Zheng, J., Clogston, J. D., Ayub, N., Aggarwal, P., . . . McNeil, S. E. (2009). Interaction of colloidal gold nanoparticles with human blood: Effects on particle size and analysis of plasma protein binding profiles. *Nanomedicine: Nanotechnology, Biology and Medicine*, 5(2), 106-117. doi:<https://doi.org/10.1016/j.nano.2008.08.001>
- [206] Lacerda, S. H., Park, J. J., Meuse, C., Pristinski, D., Becker, M. L., Karim, A., & Douglas, J. F. (2010). Interaction of gold nanoparticles with common human blood proteins. *ACS Nano*, 4(1), 365-379. doi:10.1021/nn9011187 [doi]
- [207] Monopoli, M. P., Walczyk, D., Campbell, A., Elia, G., Lynch, I., Bombelli, F. B., & Dawson, K. A. (2011). Physical-chemical aspects of protein corona: Relevance to in vitro and in vivo biological impacts of nanoparticles. *Journal of the American Chemical Society*, 133(8), 2525-2534. doi:10.1021/ja107583h [doi]
- [208] Ensign, L. M., Cone, R., & Hanes, J. (2012). Oral drug delivery with polymeric nanoparticles: The gastrointestinal mucus barriers. *Advanced Drug Delivery Reviews*, 64(6), 557-570. doi:10.1016/j.addr.2011.12.009 [doi]
- [209] Fröhlich, E., & Fröhlich, E. (2016). Cytotoxicity of nanoparticles contained in food on intestinal cells and the gut microbiota. *International Journal of Molecular Sciences*, 17(4), 509-509. doi:10.3390/ijms17040509
- [210] Pietroiusti, A., Magrini, A., & Campagnolo, L. (2016). New frontiers in nanotoxicology: Gut microbiota/microbiome-mediated effects of engineered nanomaterials. *Toxicology and Applied Pharmacology*, 299, 90-95. doi:S0041-008X(15)30162-9 [pii]
- [211] Allen-Blevins, C., You, X., Hinde, K., & Sela, D. A. (2017). Handling stress may confound murine gut microbiota studies. *Peerj*, 5, e2876-e2876. doi:10.7717/peerj.2876
- [212] Aronne, L. J., & Segal, K. R. (2002). Adiposity and fat distribution outcome measures: Assessment and clinical implications. *Obesity Research*, 10 Suppl 1, 14S-21S. doi:10.1038/oby.2002.184 [doi]
- [213] Bannunah, A. M., Vllasaliu, D., Lord, J., & Stolnik, S. (2014). Mechanisms of nanoparticle internalization and transport across an intestinal epithelial cell model: Effect of size and surface charge. *Molecular Pharmaceutics*, 11(12), 4363-4373. doi:10.1021/mp500439c [doi]

- [214] Bu, Q., Yan, G., Deng, P., Peng, F., Lin, H., Xu, Y., . . . Zhao, Y. L. (2010). NMR-based metabolomic study of the sub-acute toxicity of titanium dioxide nanoparticles in rats after oral administration. *Nanotechnology*, 21(12), 125105-4484/21/12/125105. Epub 2010 Mar 5. doi:10.1088/0957-4484/21/12/125105 [doi]
- [215] Cao, X., Han, Y., Gu, M., Du, H., Song, M., Zhu, X., . . . Xiao, H. (2020). Foodborne titanium dioxide nanoparticles induce stronger adverse effects in obese mice than non-obese mice: Gut microbiota dysbiosis, colonic inflammation, and proteome alterations. *Small*, 16(36), 2001858. doi:10.1002/smll.202001858
- [216] Chithrani, B. D., & Chan, W. C. (2007). Elucidating the mechanism of cellular uptake and removal of protein-coated gold nanoparticles of different sizes and shapes. *Nano Letters*, 7(6), 1542-1550. doi:10.1021/nl070363y [doi]
- [217] Chithrani, B. D., Ghazani, A. A., & Chan, W. C. (2006). Determining the size and shape dependence of gold nanoparticle uptake into mammalian cells. *Nano Letters*, 6(4), 662-668. doi:10.1021/nl052396o [doi]
- [218] Hoggatt, A. F., Hoggatt, J., Honerlaw, M., & Pelus, L. M. (2010). A spoonful of sugar helps the medicine go down: A novel technique to improve oral gavage in mice. *Journal of the American Association for Laboratory Animal Science* : JAALAS, 49(3), 329-334. doi:2010000329 [pii]
- [219] Mao, Z., Li, Y., Dong, T., Zhang, L., Zhang, Y., Li, S., . . . Xia, Y. (2019). Exposure to titanium dioxide nanoparticles during pregnancy changed maternal gut microbiota and increased blood glucose of rat. *Nanoscale Research Letters*, 14(1), 26-26. doi:10.1186/s11671-018-2834-5
- [220] McCracken, C., Dutta, P. K., & Waldman, W. J. (2016). Critical assessment of toxicological effects of ingested nanoparticles. *Environmental Science: Nano*, 3(2), 256-282. doi:10.1039/C5EN00242G
- [221] Pinget, G., Tan, J., Janac, B., Kaakoush, N. O., Angelatos, A. S., O'Sullivan, J., . . . Macia, L. (2019). Corrigendum: Impact of the food additive titanium dioxide (E171) on gut microbiota-host interaction. *Frontiers in Nutrition*, 6, 100-100. doi:10.3389/fnut.2019.00100
- [222] Wang, H., Du, L. J., Song, Z. M., & Chen, X. X. (2013). Progress in the characterization and safety evaluation of engineered inorganic nanomaterials in food. *Nanomedicine (London, England)*, 8(12), 2007-2025. doi:10.2217/nmm.13.176 [doi]
- [223] Yao, M., He, L., McClements, D. J., & Xiao, H. (2015). Uptake of gold nanoparticles by intestinal epithelial cells: Impact of particle size on their absorption, accumulation, and toxicity. *Journal of Agricultural and Food Chemistry*, 63(36), 8044-8049. doi:10.1021/acs.jafc.5b03242 [doi]
- [224] Clarke, S. F., Murphy, E. F., Nilaweera, K., Ross, P. R., Shanahan, F., O'Toole, P. W., & Cotter, P. D. (2012). The gut microbiota and its relationship to diet and obesity: New insights. *Gut Microbes*, 3(3), 186-202. doi:10.4161/gmic.20168 [doi]

- [225] Guo, X., Li, J., Tang, R., Zhang, G., Zeng, H., Wood, R. J., & Liu, Z. (2017). High fat diet alters gut microbiota and the expression of paneth cell-antimicrobial peptides preceding changes of circulating inflammatory cytokines
- [226] Robinson, C. J., Bohannon, B. J. M., & Young, V. B. (2010). From structure to function: The ecology of host-associated microbial communities. *Microbiology and Molecular Biology Reviews* : MMBR, 74(3), 453-476. doi:10.1128/MMBR.00014-10
- [227] Bhatt, I., & Tripathi, B. N. (2011). Interaction of engineered nanoparticles with various components of the environment and possible strategies for their risk assessment. *Chemosphere*, 82(3), 308-317. doi:10.1016/j.chemosphere.2010.10.011 [doi]
- [228] Gregory, J. (2005). *Particles in water: Properties and process* . London: IWA Publishing.
- [229] Ju-Nam, Y., & Lead, J. R. (2008). Manufactured nanoparticles: An overview of their chemistry, interactions and potential environmental implications. *The Science of the Total Environment*, 400(1-3), 396-414. doi:10.1016/j.scitotenv.2008.06.042 [doi]
- [230] Lowry, G. V., Gregory, K. B., Apte, S. C., & Lead, J. R. (2012). Transformations of nanomaterials in the environment. *Environmental Science & Technology*, 46(13), 6893-6899. doi:10.1021/es300839e [doi]
- [231] Navarro, E., Baun, A., Behra, R., Hartmann, N. B., Filser, J., Miao, A. J., . . . Sigg, L. (2008). Environmental behavior and ecotoxicity of engineered nanoparticles to algae, plants, and fungi. *Ecotoxicology* (London, England), 17(5), 372-386. doi:10.1007/s10646-008-0214-0 [doi]
- [232] Lammel, T., & Sturve, J. (2018). Assessment of titanium dioxide nanoparticle toxicity in the rainbow trout (*onchorynchus mykiss*) liver and gill cell lines RTL-W1 and RTgill-W1 under particular consideration of nanoparticle stability and interference with fluorometric assays. *Nanoimpact*, 11, 1-19. doi:https://doi.org/10.1016/j.impact.2018.01.001
- [233] Lammel, T., Mackevica, A., Johansson, B. R., & Sturve, J. (2019). Endocytosis, intracellular fate, accumulation, and agglomeration of titanium dioxide (TiO<sub>2</sub>) nanoparticles in the rainbow trout liver cell line RTL-W1. *Environmental Science and Pollution Research International*, 26(15), 15354-15372. doi:10.1007/s11356-019-04856-1
- [234] Doyle, J. J., Ward, J. E., & Mason, R. (2015). An examination of the ingestion, bioaccumulation, and depuration of titanium dioxide nanoparticles by the blue mussel (*mytilus edulis*) and the eastern oyster (*crassostrea virginica*). *Marine Environmental Research*, 110, 45-52. doi:S0141-1136(15)30027-1 [pii]
- [235] Doyle, J. J., Ward, J. E., & Mason, R. (2016). Exposure of bivalve shellfish to titania nanoparticles under an environmental-spill scenario: Encounter, ingestion and egestion. *Journal of the Marine Biological Association of the United Kingdom*, 96(1), 137-149. doi:DOI: 10.1017/S0025315415001174

- [236] Johnson, B. D., Gilbert, S. L., Khan, B., Carroll, D. L., & Ringwood, A. H. (2015). Cellular responses of eastern oysters, *crassostrea virginica*, to titanium dioxide nanoparticles. *Marine Environmental Research*, 111, 135-143. doi:S0141-1136(15)30007-6 [pii]
- [237] Ward, J., & Kach, D. (2009). Marine aggregates facilitate ingestion of nanoparticles by suspension-feeding bivalves. *Marine Environmental Research*, 68, 137-142. doi:10.1016/j.marenvres.2009.05.002
- [238] Kumar, V., Sharma, N., & Maitra, S. S. (2017). In vitro and in vivo toxicity assessment of nanoparticles. *International Nano Letters*, 7(4), 243-256. doi:10.1007/s40089-017-0221-3
- [239] Badawy, A. M. E., Luxton, T. P., Silva, R. G., Scheckel, K. G., Suidan, M. T., & Tolaymat, T. M. (2010). Impact of environmental conditions (pH, ionic strength, and electrolyte type) on the surface charge and aggregation of silver nanoparticles suspensions. *Environmental Science & Technology*, 44(4), 1260-1266. doi:10.1021/es902240k
- [240] Mukherjee, B., & Weaver, J. W. (2010). Aggregation and charge behavior of metallic and nonmetallic nanoparticles in the presence of competing similarly-charged inorganic ions. *Environmental Science & Technology*, 44(9), 3332-3338. doi:10.1021/es903456e
- [241] Araújo, F., Martins, C., Azevedo, C., & Sarmento, B. (2018). Chemical modification of drug molecules as strategy to reduce interactions with mucus. *Advanced Drug Delivery Reviews*, 124, 98-106. doi:S0169-409X(17)30201-6 [pii]
- [242] Leal, J., Smyth, H. D. C., & Ghosh, D. (2017). Physicochemical properties of mucus and their impact on transmucosal drug delivery. *International Journal of Pharmaceutics*, 532(1), 555-572. doi:10.1016/j.ijpharm.2017.09.018
- [243] Murgia, X., Loretz, B., Hartwig, O., Hittinger, M., & Lehr, C. M. (2018). The role of mucus on drug transport and its potential to affect therapeutic outcomes. *Advanced Drug Delivery Reviews*, 124, 82-97. doi:S0169-409X(17)30233-8 [pii]
- [244] Witten, J., & Ribbeck, K. (2017). The particle in the spider's web: Transport through biological hydrogels. *Nanoscale*, 9(24), 8080-8095. doi:10.1039/c6nr09736g [doi]
- [245] Wu, L., Shan, W., Zhang, Z., & Huang, Y. (2018). Engineering nanomaterials to overcome the mucosal barrier by modulating surface properties. *Advanced Drug Delivery Reviews*, 124, 150-163. doi:S0169-409X(17)30205-3 [pii]
- [246] Marcos-López, M., Caldach-Giner, J., Mirimin, L., MacCarthy, E., Rodger, H. D., O'Connor, I., . . . Piazzon, M. C. (2018). Gene expression analysis of atlantic salmon gills reveals mucin 5 and interleukin 4/13 as key molecules during amoebic gill disease. *Scientific Reports*, 8(1), 13689. doi:10.1038/s41598-018-32019-8
- [247] Minniti, G., Hagen, L. H., Porcellato, D., Jørgensen, S. M., Pope, P. B., & Vaaje-Kolstad, G. (2017). The skin-mucus microbial community of farmed atlantic

- salmon (*salmo salar*). *Frontiers in Microbiology*, 8, 2043. doi:10.3389/fmicb.2017.02043 [doi]
- [248] Padra, J. T., Murugan, A. V. M., Sundell, K., Sundh, H., Benktander, J., & Lindén, S. K. (2019). Fish pathogen binding to mucins from atlantic salmon and arctic char differs in avidity and specificity and is modulated by fluid velocity. *PloS One*, 14(5), e0215583. doi:10.1371/journal.pone.0215583 [doi]
- [249] Pérez-Sánchez, J., Estensoro, I., Redondo, M. J., Calduch-Giner, J., Kaushik, S., & Sitjà-Bobadilla, A. (2013). Mucins as diagnostic and prognostic biomarkers in a fish-parasite model: Transcriptional and functional analysis. *PloS One*, 8(6), e65457. doi:10.1371/journal.pone.0065457
- [250] Sveen, L. R., Grammes, F. T., Ytteborg, E., Takle, H., & Jørgensen, S. M. (2017). Genome-wide analysis of atlantic salmon (*salmo salar*) mucin genes and their role as biomarkers. *Plos One*, 12(12) doi:e0189103. <https://doi.org/10.1371/journal.pone.0189103>
- [251] Melrose, J. (2019). Mucin-like glycopolymer gels in electrosensory tissues generate cues which direct electrolocation in amphibians and neuronal activation in mammals. *Neural Regeneration Research*, 14(7), 1191-1195. doi:10.4103/1673-5374.251298 [doi]
- [252] Pales Espinosa, E., Koller, A., & Allam, B. (2016). Proteomic characterization of mucosal secretions in the eastern oyster, *crassostrea virginica*. *Journal of Proteomics*, 132, 63-76. doi:S1874-3919(15)30191-3 [pii]
- [253] Jevtov, I., Samuelsson, T., Yao, G., Amsterdam, A., & Ribbeck, K. (2014). Zebrafish as a model to study live mucus physiology. *Scientific Reports*, 4, 6653-6653. doi:10.1038/srep06653
- [254] van der Marel, M.C., (2012). *Carp Mucus and its Role in Mucosal Defense*. PhD Thesis. Wageningen University. The Netherlands.
- [255] Fudge, D. S., Levy, N., Chiu, S., & Gosline, J. M. (2005). Composition, morphology and mechanics of hagfish slime. *The Journal of Experimental Biology*, 208(Pt 24), 4613-4625. doi:208/24/4613 [pii]
- [256] Wang, D., Wang, P., Wang, C., & Ao, Y. (2019). Effects of interactions between humic acid and heavy metal ions on the aggregation of TiO<sub>2</sub> nanoparticles in water environment. *Environmental Pollution*, 248, 834-844. doi:<https://doi.org/10.1016/j.envpol.2019.02.084>
- [257] He, L., Lin, M., Li, H., & Kim, N. (2010). Surface-enhanced raman spectroscopy coupled with dendritic silver nanosubstrate for detection of restricted antibiotics. *Journal of Raman Spectroscopy*, 41(7), 739-744. doi:10.1002/jrs.2505
- [258] Colvin, V. L. (2003). The potential environmental impact of engineered nanomaterials. *Nature Biotechnology*, 21(10), 1166-1170. doi:10.1038/nbt875
- [259] Dale, A. L., Casman, E. A., Lowry, G. V., Lead, J. R., Viparelli, E., & Baalousha, M. (2015). Modeling nanomaterial environmental fate in aquatic systems. *Environmental Science & Technology*, 49(5), 2587-2593. doi:10.1021/es505076w

- [260] Fabrega, J., Luoma, S. N., Tyler, C. R., Galloway, T. S., & Lead, J. R. (2011). Silver nanoparticles: Behaviour and effects in the aquatic environment. *Environment International*, 37(2), 517-531. doi:10.1016/j.envint.2010.10.012
- [261] Moore, M. N. (2006). Do nanoparticles present ecotoxicological risks for the health of the aquatic environment? *Environment International*, 32(8), 967-976. doi:S0160-4120(06)00085-7 [pii]
- [262] Sani-Kast, N., Scheringer, M., Slomberg, D., Labille, J., Praetorius, A., Ollivier, P., & Hungerbühler, K. (2015). Addressing the complexity of water chemistry in environmental fate modeling for engineered nanoparticles. *The Science of the Total Environment*, 535, 150-159. doi:S0048-9697(14)01727-6 [pii]
- [263] Cao, X., Ma, C., Gao, Z., Zheng, J., He, L., McClements, D. J., & Xiao, H. (2016). Characterization of the interactions between titanium dioxide nanoparticles and polymethoxyflavones using surface-enhanced raman spectroscopy. *Journal of Agricultural and Food Chemistry*, 64(49), 9436-9441. doi:10.1021/acs.jafc.6b03906
- [264] Nur, Y., Lead, J. R., & Baalousha, M. (2015). Evaluation of charge and agglomeration behavior of TiO<sub>2</sub> nanoparticles in ecotoxicological media. *Science of the Total Environment; Special Issue: Engineered Nanoparticles in Soils and Waters*, 535, 45-53. doi:https://doi.org/10.1016/j.scitotenv.2014.11.057
- [265] von Moos, N., Bowen, P., & Slaveykova, V. I. (2014). Bioavailability of inorganic nanoparticles to planktonic bacteria and aquatic microalgae in freshwater. *Environmental Science: Nano*, 1(3), 214-232. doi:10.1039/C3EN00054K
- [266] Kesimer, M., & Sheehan, J. K. (2012). Mass spectrometric analysis of mucin core proteins. *Methods in Molecular Biology (Clifton, N.J.)*, 842, 67-79. doi:10.1007/978-1-61779-513-8\_4 [doi]
- [267] Ou, Y., Lin, J., Zou, H., & Liao, D. (2005). Effects of surface modification of TiO<sub>2</sub> with ascorbic acid on photocatalytic decolorization of an azo dye reactions and mechanisms. *Journal of Molecular Catalysis A: Chemical*, 241(1), 59-64. doi:https://doi.org/10.1016/j.molcata.2005.06.054
- [268] Kaniyankandy, S., Rawalekar, S., Sen, A., Ganguly, B., & Ghosh, H. N. (2012). Does bridging geometry influence interfacial electron transfer dynamics? case of the enediol-TiO<sub>2</sub> system. *The Journal of Physical Chemistry C*, 116(1), 98-103. doi:10.1021/jp207054f
- [269] Ambort, D., Johansson, M. E., Gustafsson, J. K., Nilsson, H. E., Ermund, A., Johansson, B. R., . . . Hansson, G. C. (2012). Calcium and pH-dependent packing and release of the gel-forming MUC2 mucin. *Proceedings of the National Academy of Sciences of the United States of America*, 109(15), 5645-5650. doi:10.1073/pnas.1120269109 [doi]
- [270] Yang, X., Forier, K., Steukers, L., Van Vlierberghe, S., Dubruel, P., Braeckmans, K., & Glorieux, S. (2012). Immobilization of pseudorabies virus in porcine

- tracheal respiratory mucus revealed by single particle tracking. *PloS One*, 7, e51054. doi:10.1371/journal.pone.0051054
- [271] Wagner, C. E., Turner, B. S., Rubinstein, M., McKinley, G. H., & Ribbeck, K. (2017). A rheological study of the association and dynamics of MUC5AC gels. *Biomacromolecules*, 18(11), 3654-3664. doi:10.1021/acs.biomac.7b00809 [doi]
  - [272] Cone, R. A. (2009). Barrier properties of mucus. *Advanced Drug Delivery Reviews*, 61(2), 75-85. doi:10.1016/j.addr.2008.09.008 [doi]
  - [273] Houtman, J. C., Brown, P. H., Bowden, B., Yamaguchi, H., Appella, E., Samelson, L. E., & Schuck, P. (2007). Studying multisite binary and ternary protein interactions by global analysis of isothermal titration calorimetry data in SEDPHAT: Application to adaptor protein complexes in cell signaling. *Protein Science : A Publication of the Protein Society*, 16(1), 30-42. doi:16/1/30 [pii]
  - [274] Omanovic-Miklicanin, E., Manfield, I., & Wilkins, T. (2017). Application of isothermal titration calorimetry in evaluation of protein–nanoparticle interactions. *Journal of Thermal Analysis and Calorimetry*, 127(1), 605-613. doi:10.1007/s10973-016-5764-4
  - [275] Freyer, M. W., & Lewis, E. A. (2008). Isothermal titration calorimetry: Experimental design, data analysis, and probing Macromolecule/Ligand binding and kinetic interactions. *Methods in cell biology* (pp. 79-113) Academic Press. doi:https://doi.org/10.1016/S0091-679X(07)84004-0 Retrieved from <http://www.sciencedirect.com/science/article/pii/S0091679X07840040>
  - [276] Norde, W. (2008). My voyage of discovery to proteins in flatland ...and beyond. *Colloids and Surfaces.B, Biointerfaces*, 61(1), 1-9. doi:S0927-7765(07)00390-6 [pii]
  - [277] Guo, H., He, L., & Xing, B. (2017). Applications of surface-enhanced raman spectroscopy in the analysis of nanoparticles in the environment. *Environmental Science: Nano*, 4(11), 2093-2107. doi:10.1039/C7EN00653E
  - [278] Yang, L., Jiang, X., Ruan, W., Yang, J., Zhao, B., Xu, W., & Lombardi, J. R. (2009). Charge-transfer-induced surface-enhanced raman scattering on Ag–TiO<sub>2</sub> nanocomposites. *The Journal of Physical Chemistry C*, 113(36), 16226-16231. doi:10.1021/jp903600r
  - [279] Bansil, R., Yannas, I. V., & Stanley, H. E. (1978). Raman spectroscopy: A structural probe of glycosaminoglycans. *Biochimica Et Biophysica Acta (BBA) - General Subjects*, 541(4), 535-542. doi:https://doi.org/10.1016/0304-4165(78)90163-0
  - [280] Ashton, L., Pudney, P. D., Blanch, E. W., & Yakubov, G. E. (2013). Understanding glycoprotein behaviours using raman and raman optical activity spectroscopies: Characterising the entanglement induced conformational changes in oligosaccharide chains of mucin. *Advances in Colloid and Interface Science*, 199-200, 66-77. doi:S0001-8686(13)00071-7 [pii]

- [281] De Gelder, J., De Gussem, K., Vandenabeele, P., & Moens, L. (2007). Reference database of raman spectra of biological molecules. *Journal of Raman Spectroscopy*, 38(9), 1133-1147. doi:<https://doi.org/10.1002/jrs.1734>
- [282] Steen, P., Rudd, P., Wormald, M., Dwek, R., & Opdenakker, G. (2000). O-linked glycosylation in focus. *Trends in Glycoscience and Glycotechnology*, 12, 35-49. doi:10.4052/tigg.12.35
- [283] She, C. Y., Dinh, N. D., & Tu, A. T. (1974). Laser raman scattering of glucosamine N-acetylglucosamine, and glucuronic acid. *Biochimica Et Biophysica Acta (BBA) - General Subjects*, 372(2), 345-357. doi:[https://doi.org/10.1016/0304-4165\(74\)90196-2](https://doi.org/10.1016/0304-4165(74)90196-2)
- [284] Deters, A., Petereit, F., Schmidgall, J., & Hensel, A. (2008). N-acetyl-D-glucosamine oligosaccharides induce mucin secretion from colonic tissue and induce differentiation of human keratinocytes. *The Journal of Pharmacy and Pharmacology*, 60(2), 197-204. doi:10.1211/jpp.60.2.0008 [doi]
- [285] Zhou, H., Pandya, J. K., Tan, Y., Liu, J., Peng, S., Muriel Mundo, J. L., . . . McClements, D. J. (2019). Role of mucin in behavior of food-grade TiO(2) nanoparticles under simulated oral conditions. *Journal of Agricultural and Food Chemistry*, 67(20), 5882-5890. doi:10.1021/acs.jafc.9b01732 [doi]
- [286] FDA (Food and Drug Administration). Summary of color additives for use in the united states in foods, drugs, cosmetics, and medical devices [online]. available at : . Retrieved from <http://www.fda.gov/forindustry/color-additives/coloradditiveinventories/ucm115641.htm>. (2017)
- [287] Allen, R. (2016). The cytotoxic and genotoxic potential of titanium dioxide (TiO<sub>2</sub>) nanoparticles on human SH-SY5Y neuronal cells in vitro. *The Plymouth Student Scientist*, 9(2), 5-28.
- [288] Pandey, R. K., & Prajapati, V. K. (2018). Molecular and immunological toxic effects of nanoparticles. *International Journal of Biological Macromolecules*, 107(Pt A), 1278-1293. doi:S0141-8130(17)33394-9 [pii]
- [289] Bourikas, K., Kordulis, C., & Lycourghiotis, A. (2014). Titanium dioxide (anatase and rutile): Surface chemistry, Liquid–Solid interface chemistry, and scientific synthesis of supported catalysts. *Chemical Reviews*, 114(19), 9754-9823. doi:10.1021/cr300230q
- [290] Baan, R., Straif, K., Grosse, Y., Secretan, B., El Ghissassi, F., Coglianò, V., & WHO International Agency for Research on Cancer Monograph Working Group. (2006). Carcinogenicity of carbon black, titanium dioxide, and talc. *The Lancet.Oncology*, 7(4), 295-296. doi:10.1016/s1470-2045(06)70651-9 [doi]
- [291] Wang, J., Liu, Y., Jiao, F., Lao, F., Li, W., Gu, Y., . . . Chen, C. (2008). Time-dependent translocation and potential impairment on central nervous system by intranasally instilled TiO(2) nanoparticles. *Toxicology*, 254(1-2), 82-90. doi:10.1016/j.tox.2008.09.014 [doi]



- [292] EFSA ANS Panel. (2016). Scientific opinion on the re-evaluation of titanium dioxide (E 171) as a food additive *Efsa j.*, 14(4545), 83.  
doi:<https://doi.org/10.2903/j.efsa.2016.4545>
- [293] Iavicoli, I., Leso, V., & Bergamaschi, A. (2012). Toxicological effects of titanium dioxide nanoparticles: A review of in vivo studies. *Journal of Nanomaterials*, 2012, 964381. doi:10.1155/2012/964381
- [294] Iavicoli, I., Leso, V., Fontana, L., & Bergamaschi, A. (2011). Toxicological effects of titanium dioxide nanoparticles: A review of in vitro mammalian studies. *European Review for Medical and Pharmacological Sciences*, 15(5), 481-508.
- [295] Heringa, M. B., Geraets, L., van Eijkeren, J. C., Vandebriel, R. J., de Jong, W. H., & Oomen, A. G. (2016). Risk assessment of titanium dioxide nanoparticles via oral exposure, including toxicokinetic considerations. *Nanotoxicology*, 10(10), 1515-1525. doi:10.1080/17435390.2016.1238113 [doi]
- [296] Geraets, L., Oomen, A. G., Krystek, P., Jacobsen, N. R., Wallin, H., Laurentie, M., . . . de Jong, W. H. (2014). Tissue distribution and elimination after oral and intravenous administration of different titanium dioxide nanoparticles in rats. *Particle and Fibre Toxicology*, 11, 30-8977-11-30. doi:10.1186/1743-8977-11-30 [doi]
- [297] Dufey, W., Moniz, K., Allen-Vercoe, E., Ropers, M., & Walker, V. K. (2017). Impact of food grade and nano-TiO<sub>2</sub> particles on a human intestinal community. *Food and Chemical Toxicology*, 106, 242-249.  
doi:<https://doi.org/10.1016/j.fct.2017.05.050>
- [298] Peters, R. J., van Bommel, G., Herrera-Rivera, Z., Helsper, H. P., Marvin, H. J., Weigel, S., . . . Bouwmeester, H. (2014). Characterization of titanium dioxide nanoparticles in food products: Analytical methods to define nanoparticles. *Journal of Agricultural and Food Chemistry*, 62(27), 6285-6293.  
doi:10.1021/jf5011885 [doi]